ORGANIC MATTER DISTRIBUTION AND TURNOVER ALONG A GRADIENT FROM
FOREST TO TIDAL CREEK

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Many coastal ecosystems migrate landward under the influence of rising sea level.
Typical zonation of plant communities along coastal shorelines consists of tidal marshes, a transition zone, and adjacent upland or wetland forest. This study examined organic matter distribution along this gradient. I hypothesized that total above and belowground organic carbon mass would follow the pattern: forest $>$ transition $>$ high marsh $=$ mid marsh > low marsh > tidal creek. This study was conducted at the upper Phillips Creek study area on the eastern shore of Virginia. A zonation map of the study area was created, and low marsh was divided into two zones based on two growth forms of Spartina alterniflora. Sample sites were selected using a stratified random sampling approach. A nested plot design was used to harvest vegetation, obtain soil cores, and collect quantitative data on trees, shrubs and large wood detritus. Unharvested tree and shrub masses were estimated using regression equations. Loss on ignition was determined for vegetation and soils. Organic carbon mass was estimated to be $50 \%$ of organic matter. Total above and belowground organic carbon mass (mean $\left(\mathrm{kg} / \mathrm{m}^{2}\right) \pm$ S.E.) for each zone was: forest $24.3 \pm 2.1$, high marsh $14.2 \pm 0.7$, transition $12.8 \pm 0.6$, LMSS $12.6 \pm 0.8$, LMTS $11.3 \pm 0.7$, and tidal creek $8.7 \pm 0.3$. The greatest loss of carbon occurred in the transformation of forest to high marsh. Organic carbon turnover rates for Phillips Creek were estimated for steep and gentle slopes by projecting an 80year period of sea level rise at $5 \mathrm{~mm} / \mathrm{year}$. After 80 years, marsh and transition zones experienced $100 \%$ turnover in both profiles. The forest experienced turnover rates of
$25 \%$ and $71 \%$ in steep and gentle profiles, respectively. Horizontal turnover rates of carbon associated with state change were approximately one order of magnitude lower
than those associated with net primary production. However, horizontal turnover of ecosystem states can change coastal landscapes within the time span of a century.

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## INTRODUCTION

Eustatic sea level rise has been cited as the major force behind coastal landscape changes that include losses of salt and brackish marshes in low lying coastal areas in the eastern and Gulf Coast states of the USA (Hayden et al. 1991, Reed and Cahoon 1992, Moorhead and Brinson 1995). However, steady sea level rise is not the only force behind changes in coastal landscapes. Crustal subsidence, reduced sedimentation rates, and marsh shoreline erosion may be additional driving forces transforming salt marsh and brackish marshes into unvegetated open water systems (Stevenson et al. 1985, Downs et al. 1994). In addition, storm surges that accompany hurricanes and tropical storms may act as catalysts in the transformation of forest into high marsh (Brinson et al. 1995). Changes in ecosystem state are called "state change" (entire ecosystems replace one another as they move horizontally across the landscape). The type of state change depends on the current state, upland slope, and sediment supply. For mainland marshes, four combinations of sea level induced state changes may occur: (1) where slopes are relatively flat and sediment supplies are abundant, low and high marsh zones encroach on forest by migrating landward and low marsh zones prograde toward the estuary (Redfield 1972); (2) where slopes are flat but sediment supply is low, the low and high marsh zones still migrate overland, but the edge of low marsh erodes becoming subtidal; (3) where slopes are steep and sediment supply is high, the high marsh stalls at the forest margin and the low marsh progrades toward the estuary; and (4) where slopes are steep but sediment supply is low, the high marsh stalls at the forest edge, and the low marsh erodes (Brinson et al. 1995). For many mainland marshes along the eastern shore of Virginia, sediment supply is low. Thus erosion along the
banks of tidal creeks and margins of estuaries combined with marsh stalling or marsh overland migration are the more commonly observed responses to relative sea level rise.

Several studies have focused on local and regional processes contributing to area loss of marshes. In Louisiana, relative sea level rise, due principally to deep subsidence of the Mississippi River deltaic plain, is the major cause for the rapid deterioration of salt marshes despite high sediment accretion rates. Die-off of saltmarsh vegetation due to conditions associated with long-duration flooding, rather than shoreline erosion, has transformed the marshes into unvegetated open water areas (Reed and Cahoon 1992). In the Chesapeake Bay, a Maryland brackish marsh system has experienced extensive erosion of 2300 ha in a 40 y time span. This was largely due to a net loss of mineral sediment coupled with sea level rise exceeding organic accretion rates (Stevenson et al. 1985). In southeastern North Carolina, a loss of lagoonal salt marshes has been attributed to relative rise in sea level and inadequate sediment deposition due to coastal inlet dredging (Hackney and Cleary 1987). On the other hand, in North Inlet, South Carolina, Gardener et al. (1992) reported on the landward migration of salt marsh where relative sea level rise was believed to cause low-lying forest and high marsh to move overland.

At the Long-Term Ecological Research (LTER) site on the Virginia Coast Reserve, (VCR), the effects of relative sea level rise and storm surges on the wetlands and adjacent uplands are being studied. The process is believed to involve transgression of marshes landward that caused mainland areas of low land surface elevation to undergo a series of changes in plant community composition. For instance, wrack deposition in the high marsh, which occurs as a result storm tidal surges, reduces the productivity of
most salt marsh species. Juncus roemerianus Scheele patches were found to be susceptible to replacement by other high marsh species following wrack disturbance (Tolley and Christian 1999). State change is believed to arise from vertical changes in at least one of three surfaces: the land, the sea, and the freshwater table. At the VCR, even small changes in the relationship between land surface and sea level may result in ecosystem state change (Hayden et al. 1991, Hayden et al. 1995). For example, from 1938 to 1990 Phillips Creek marsh (on the VCR mainland) had gained approximately 27.3 ha of marsh, primarily from high marsh encroachment on uplands (Kastler and Wieberg 1996).

While the slope of the land should regulate the rate of transgression under constant sea level rise, other factors may be involved. Resistance to state change may be an important property that regulates the rate of landward migration. For instance, the transition from high marsh to intertidal low marsh is impeded by the structure of clonal saltmarsh plants, and the accumulation of organic matter in the form of peat elevating the marsh surface above intertidal elevations (Bertness 1988). A six-year study by Brinson and Christian (1999) found that the clonal species J. roemerianus maintained stable patches in a high marsh, despite wrack disturbance and marsh subsidence. In addition, transition from low marsh to subtidal creek and intertidal mud flat is impeded by Spartina alterniflora Loisel, a creekside plant whose aerial stems facilitate sediment accumulation (Christiansen 1998). Possible mechanisms for change, and components involved in maintaining each ecosystem state from low marsh to the forest, are illustrated in Figure 1, a conceptual state change model created by Brinson et al. (1995). Two of the important components, total aboveground and belowground organic matter, along a continuum from tidal creek to upland forest, have not been quantitatively


Figure 1. Conceptual model of the five states and four transitions from terrestrial forest to subtidal heterotrophic sediment. Factors that facilitate change appear at the top and components excluded with state change appear at the bottom Taken from Brinson et al. (1995)
measured on a mainland coast. Although there have been a substantial number of saltmarsh plant studies conducted along the Atlantic and Gulf coast states (Kruczynski et al. 1978, Stout 1978, Smith et al. 1979, Roman and Daiber 1984, Bellis and Gaither 1985), their focus have been on saltmarsh species productivity and biomass at a particular locale without relation to sea level rise. On the other hand, studies by Stevenson et al. (1985), Hackney and Cleary (1987), and Reed and Cahoon (1992) addressed processes involved in the conversion of salt/brackish marshes to subtidal lagoons, and Gardner's study (1992) focused on landward migration of high marsh displacing forest. However, none of these studies addressed the quantity or quality of total organic matter, above and belowground, involved in transitional changes from marsh to open water, or from upland forest into high marsh.

With the change from one ecosystem state to another, there are losses of certain ecosystem components and gains in others (Brinson et al. 1995). Therefore, the purpose for this quantitative study is to estimate the magnitude of above and belowground organic matter, consisting primarily of woody and herbaceous vegetation and soil organic matter. In turn, these estimates will enable me to evaluate the magnitude of gains and losses of organic matter and organic carbon over long time scales. Although this study will not explain the mechanisms responsible for organic carbon changes, it will provide the foundation upon which those processes can be studied.

Based on literature and field observations, I reasoned that forest would exceed the high marsh in total organic matter due to the magnitude of mature tree mass that would compensate for having less belowground organic matter than the high marsh. In
addition, I expected high marsh and mid-marsh with organic rich soil to exceed the low marsh in total organic matter.

In summary, the goals of this study were to quantify and characterize above and belowground organic matter for each ecosystem state in Phillips Creek. I compare my findings with other coastal ecosystem studies of the Atlantic and Gulf Coast states in the USA. I expect the to find the following pattern for total organic matter $\left(\mathrm{kg} / \mathrm{m}^{2}\right)$ : Forest > Transition $>$ High marsh $=$ Mid-marsh $>$ Low $m a r s h>$ Tidal creek .

Phillips Creek is a saltwater tidal creek located on the eastern shore of Virginia in Northampton County within the Brownsville area of the Virginia Coast Reserve (VCR), (Figure 2). The VCR, owned by The Nature Conservancy, is one of the Long-Term Ecological Research sites (LTER) supported by the National Science Foundation.

In the earlier part of this century, the upper region of Phillips Creek (37026' $\mathrm{N}, 75^{\circ} 52^{\prime}$ W) was used as pasture. In the present day, this former use is clearly visible in weathered fence posts that dissect the high marsh, remnants of a former fence line.

Also, a piece of rusting plow equipment is partially buried in the sediment of the present day high marsh.

The northern end of Phillips Creek contains a gently sloping, broad expanse of marsh surrounded on three sides by loblolly pine (Pinus taeda) forests and upland agricultural fields (Figure 2). This configuration forms the basis for a complex of six ecosystem states beginning with a tidal creek and an intertidal or low salt marsh, followed by a mid-marsh, then grading inland to a brackish high marsh, marsh-forest transition, and adjacent forest.

The predominant soil types are Chincoteague in the low, mid, and high marshes, and Magotha in the transition zones, which are very poorly drained and poorly drained, respectively. Nimo, Munden, and Dragston are the soil types present in the forested areas (USDA Soil Conservation Service 1989). All of these soils are found on gentle slopes ranging from 0 to $2 \%$. Hmieleski (1994) previously described these soils, except for Dragston, as the principal soil types of Phillips Creek and of Northampton County.


Figure 2. Phillips Creek study area is located within the Virginia Coast Reserve Long-Term Ecological Research complex. This complex is situated on the eastern shore of Virginia on the Delmarva Peninsula. Mainland marshes ( small rectangle) are attached to the major land mass in contrast to lagoonal and back barrier marshes to the east.

The Dragston soil is found along drainages, while the Nimo is found in depressions and flats. Both soil types support wetland forest, and they are described as poorly drained to somewhat poorly drained, respectively. The Munden soil, found on upland flats, is moderately well drained, and is the principal soil type for the upland forest (USDA Soil Conservation Service 1989).

The vegetation within the zones follows a salinity gradient. The forest contains an approximately 55-year old stand of loblolly pine (Pinus taeda L.), a mixture of hardwoods, and an understory of salt intolerant vegetation. In the transition zone, Juniperus virginiana $L$. is the dominant tree species, and is considered to be an indicator of forest transition. P. taeda trees, approximately 45 years old, also inhabit the forest transition. Shrubs that commonly occur in the transition zone are Iva frutescens L., Baccharis halimifolia L., and Myrica cerifera L. Herbaceous species are comprised of a mixture of freshwater and saltmarsh graminoid species, such as Panicum virgatum L., Setaria genticulata Beauv., Spartina patens (Aiton) Muhl, Distichlis spicata (L.) Greene and Juncus roemerianus Scheele. The same species are also commonly found in transition zones along the eastern shore of Virginia (Ricker 1999). The high marsh is dominated by the saltmarsh graminoids, J. roemerianus, S. patens, and D. spicata. Also occurring in the high marsh is Spartina alterniflora Loisel, commonly found in microtopographic depressions. This distribution coincides with areas of high marsh that are losing elevation and becoming devoid of emergent vegetation. The low marsh is dominated by S. alterniflora, both tall and short forms, with the tall form occurring along creekbanks. Lastly, the creek channels are devoid of emergent vegetation.

In 1991, three water level recorders were established on Phillips Creek Marsh to capture the variations in hydroperiod along a continuum from creekbank to adjacent
forest. Stasavich (1998) described the main water input for each ecosystem along this continuum. The low marsh experiences diurnal tides. The mid-marsh, due to an increase in elevation or distance from the tidal creek, relies on spring tides as the principal water source, while precipitation and storm tides are secondary. Again, due to elevation or distance to the creek, the high marsh, transition and forest zones are dependent on rainfall as a major water source. In addition, the high marsh and transition experience tidal inundation during storm surges. The forest, on the other hand, experiences tidal inundation only during infrequent events, such as northeaster storms and hurricanes.

## METHODS

## Field Map and Sampling Design

In order to comparatively study above and belowground organic matter of the marsh ecosystem zones, I created a map that depicts the approximated zone boundaries (Figure 3a). Boundaries of zones were approximated by visually assessing vegetation composition on a 1990 USDA black and white aerial photograph enlarged to a scale of 2.54 cm:100 m, a 1990 USDA infrared photograph at an approximate scale of 2.54 cm:200 m, and ground observation. Because the tall and short forms of S. alterniflora displayed a recognizable zonation pattern on the aerial photographs, the low marsh was divided into two zones based on growth forms. For zones that could not be defined by aerial photography, field measurements of distance and direction were taken using a meter tape and a compass. These measurements were then converted onto the map as approximated zone boundaries. On the map, drawn boundary lines gave zones the appearance of polygons with variable size and shape.

This map formed the basis for the stratified random sampling design from which sampling points were selected. Because most zones (forest, high marsh and low marsh) were represented by multiple polygons that were often not contiguous, a stratified random approach was chosen to increase the possibility of an even dispersal of sampling points across each zone. The stratified zones are displayed in Figure 3a with the zone indicated by a capital letter followed by the stratum in a lower case letter. Potential sampling points were plotted on the stratified zone map, using intersection points on a 28 m grid for most stratified zones. For the large forested area (stratum Fb ), a 55 m grid provided ample sampling points (Figure 3b). Points were numbered within


Figure 3a. Map of Philips Creek shown with zone boundanes and stadicaton scheme. capial leters derote zone and smal case letters denote strata LMSS = low marsh shor-form spartina atitmifiova and LMTS = low marsh tal Larm Spartina alfernticra


Figure 3b. Map of Philips Creek displeys spatial scale of the grid used to plot potential sampling points a cross zones. Legend shows atratification of zones. Legend abbreviations. LMSS = bw marsh short-form Spartina ahterniova and LMTS = love marsh tal |form Spatharaliantiora
each stratum for random selection. I used the numbered points to separate strata that were not separated by a boundary line. An example was the transition strata Tc and Td. The sampling intensity I chose was based on variability of vegetation and soils within each zone, and the field time required to sample points. I expected the transition, mid, and high marsh zones to be the most variable zones, and the forest and low marsh tall and short Spartina zones to be the most homogenous. I selected a sampling intensity of $10 \%$ for all strata except the forest, because that intensity would have provided too many forest sample points for the limited field time. I decided to begin with five forest points, two in Fa and three in Fb , and to select more points at a later date if needed and time allowed. In contrast, the mid marsh zone contained two strata with fewer than ten potential sampling points, so I randomly selected one point for each stratum to ensure sampling within all strata.

All sampling points on the map were located in the field using map, compass, and meter tape. However, some randomly selected points could not reached. This occurred several times in the low marsh tall Spartina alterniflora zone where many small creek channels, not visible on map, barred access. Therefore, I chose a substitute that was the nearest accessible sample point to the unreachable one. A PVC pipe stake was driven into the ground, and labeled to mark each point. This enabled me to make recurring visits to points until harvesting and data gathering were complete. For the remainder of the study, I will refer to these established sampling points as sample sites.

After several months of field sampling, several changes were made in the zone assignment of sample sites. Because mid-marsh and high marsh were indistinguishable in vegetation and soil characteristics, all mid-marsh sample sites were assigned to the high marsh, except one; that mid-marsh site was assigned to the low marsh short-form

Spartina alterniflora zone. In addition, three transition zone sites were assigned to the high marsh. I decided the eastern transition "finger", which extended as a peninsula into the high marsh, had completed its transition to high marsh due to low shrub density, and its disconnection from the forest. A second transition "finger" remains, except that it was reduced in area from that previously mapped (Figure 4). This transition area is comprised of Iva frutescens shrubs and Juniperus virginiana snags, and is an extension of an inland transition comprised of scattered trees (J. virginiana and P. taeda), and mixed shrubs species (l. frutescens and Baccharis halimifolia). Thus, the final number of sample sites per zone is as follows: 5 forest, 14 transition, 25 high marsh, 9 low marsh short-form S. alterniflora, and 7 low marsh tall-form S. alterniflora.

## GPS Survey

A global positioning system was used to determine coordinates and elevation of all sample sites. The roving unit was set up next to or within 1 m of the PVC stake marking the site. Static readings of 8 to 10 min duration were taken for the marsh zones, and 20 to 30 minute static readings were taken for the transition and forest sites to ensure accurate elevation readings (< 1.0 cm error). The main purpose of the GPS survey was to display the location and the relative elevation of all sites. The GPS receiver used was a Trimble 4000 SE unit (L1 only) and the software used to process GPS data points was GPS Survey version 3.20a. The GPS data generated (position coordinates and elevation) were tied to VCR1, a permanent bench mark that was established as a cignet global network tie. Elevations are tied into mean sea level based on the 1929 National Geodetic Survey. Due to the high precision of VCR1, position coordinates and elevation above mean sea level are accurate to the nearest centimeter. Further information


Figure 4. Revised map of Phillips Creek shown with adjusted zone boundanes
concerning the VCR1 benchmark can be obtained at the internet address: www.vcrlter.virginia.edu/~crc7m/gps.html.

## Aboveground Organic Matter

## Sampling Techniques

Marsh vegetation was harvested in early October 1997 from all marsh zones. Before harvesting, percent species composition and percent groundcover (Daubenmire 1959) were visually assessed within a $1 \mathrm{~m}^{2}$ quadrat at each site. From a $0.0625 \mathrm{~m}^{2}$ quadrat $(25 \times 25 \mathrm{~cm})$ area within the $1 \mathrm{~m}^{2}$ quadrat, all graminoid vegetation was clipped and surface litter collected, and placed in labeled plastic bags.

Sampling within transition and forest zones was more complex. Because of the presence of shrubs, trees, vines, and forbs, a nested sampling design was used. Trees were measured within a fixed 10 m radius plot and shrubs within a 5 m radius plot. Small woody vegetation $<1 \mathrm{~m}$ height, forbs, fruits and cones and small wood detritus, was harvested within $1 \mathrm{~m}^{2}$ quadrats. Graminoid vegetation and herbaceous litter were harvested within $0.0625 \mathrm{~m}^{2}$ quadrats. For trees and shrubs $\geq 2.54 \mathrm{~cm}$ dbh, speciesspecific regression equations were used to determine biomass (explained below).

The following data were collected at transition and forest sample sites: (1) diameter at breast height (dbh) at 1.4 m from tree base for $J$. virginiana, $P$. taeda, and hardwood tree species, (2) basal diameter, height, and average crown diameter of shrub species, (3) species, height, dbh, and decay class of all standing snags leaning less than 45 degrees, (4) species, length, decay class, and diameter at mid length of large wood (LW), defined as dead stems greater than 10 cm diameter lying prostrate or leaning greater than 45 degrees from vertical, (5) number and species identification of woody
vines, (6) tree age estimate of $P$. taeda and $J$. virginiana by obtaining increment cores from two of the largest trees of each species per plot, and (7) percent groundcover and percent herbaceous species composition within five-1 $\mathrm{m}^{2}$ quadrats.

In the transition zone, graminoids, small wood (SW) (defined as dead stems $<10 \mathrm{~cm}$ diameter lying on the ground), vegetation less than 1 m in height and less than 2.54 cm in diameter, and fruits and cones were harvested within $1 \mathrm{~m}^{2}$ quadrats at five points, equidistant, along a north-south directed transect within a 10 m radius plot. Graminoids and surface litter were harvested within a $0.0625 \mathrm{~m}^{2}$ quadrat ( $25 \times 25 \mathrm{~cm}$ right corner of a $1 \mathrm{~m}^{2}$ quadrat). The remaining vegetation categories were harvested within the $1 \mathrm{~m}^{2}$ quadrat. Forest vegetation and detritus were harvested in the same fashion, except that herbaceous species were harvested within a $1 \mathrm{~m}^{2}$ quadrat. All collected samples were stored in labeled plastic bags and later refrigerated.

In the fruit and cones category, only pine cones were found and their mass was later included in the SW category. Also, a variation in plot size occurred at one transition site. Because this site was comprised of shrubs and lacked trees, a 5 m radius plot with three $1 \mathrm{~m}^{2}$ quadrats was used.

## Processing of Vegetation and Detritus

Marsh vegetation was sorted by species, except for litter, and live and dead plants were separated. Totally brown plants were considered dead. Plant fragments that could not be identified to species were added to the surface litter category. In contrast,
herbaceous vegetation samples from transition and forest sites were not separated by species. All vegetation and detritus samples were placed in labeled paper bags and dried at $85^{\circ} \mathrm{C}$ in a drying oven to a constant weight. After drying, bags with vegetation
were weighed to the nearest 0.01 g . Bag weight was subtracted to obtain dry weight of plant material. Selected dried samples were ground using a Wiley mill with a 40-mesh screen, and stored in labeled plastic bottles until loss on ignition (LOI) analysis

## Qualitative Vegetation Analysis

Species dominance was derived from percent herbaceous species composition within the $1 \mathrm{~m}^{2}$ quadrats. Dominance was assigned to one or two species that comprised $50 \%$ or more of the area. Because the transition and forest sites had five-1 $\mathrm{m}^{2}$ sampling points, in addition to determining species dominance within each quadrat another step was taken to ascertain the dominant herbaceous species for the site. The dominant species of the $1 \mathrm{~m}^{2}$ quadrats were tallied for frequency of dominant occurrence among the five quadrats. Thus, the species that was dominant in three or more quadrats was considered the dominant site species, or two species occurring with equal frequency shared dominance. Similarly, Daubenmire's (1959) percent groundcover values were averaged for the five replicate quadrats. Dominant species and percent groundcover were used to characterize the different zones.

## Mass Estimations

Snags and large wood. Snag and large wood (LW) detritus mass were estimated per species with the following equation: Mass $(\mathrm{kg})=$ Height $(\mathrm{m})$ or length $(\mathrm{m})$ * mean cross-section area $\left(\mathrm{m}^{2}\right)^{*}$ wood specific gravity $\left(\mathrm{kg} / \mathrm{m}^{3}\right)$. Then, snag and LW biomass were summed separately in each sample plot, and converted to $\mathrm{kg} / \mathrm{m}^{2}$. Specific gravity values $\left(\mathrm{g} / \mathrm{cm}^{3}\right)$ for snags and LW, that had no evidence of decay (class 1 ), were obtained from Haygreen and Bowyer (1989) for P. taeda , J. virginiana, Prunus virginiana, and

Nyssa biflora. Because many snags and LW were in various stages of decay, three decay classes were described: Class 1 had bark intact with no evidence of wood decay, class 2 had some missing bark and mild to moderate wood decay, and class 3 had no bark and advanced stages of wood decay.

Because P. taeda and J. virginiana comprised the bulk of snag and LW categories, specific gravity values for classes 2 and 3 of these species were determined from multiple specimens of dead wood for each decay class. Each specimen was categorized, labeled, and placed in a drying oven at $85^{\circ} \mathrm{C}$ until there was no further weight loss. Dry weights were obtained to the nearest 0.01 g . Then specimens were wrapped securely in cellophane using minimal tape to keep them dry. Each wood specimen was submersed in a 1 L graduated cylinder filled with 800 mL of water, and volume displacement was recorded (Haygreen and Bowyer 1989). Specific gravity $\left(\mathrm{g} / \mathrm{cm}^{3}\right)$ was estimated by calculating the mean measurements in each class per species. Because specific gravity values of class 2 and 3 for $P$. taeda overlapped, their values were pooled for a calculated mean. Specific gravity values $\left(\mathrm{g} / \mathrm{cm}^{3}\right)$ of four species, for which dead wood mass was calculated, are listed in Table 1.

Woody vines. I did not find regression equations for Rhus radicans (poison ivy) or Campsis radicans (trumpet vine), the two most common vines growing in the forest and transitions zones. Because of the magnitude of tree biomass in the forest and tree, shrub, and herbaceous plant contribution in the transition, I believed that vine biomass would be inconsequential to the total site biomass. Therefore, woody vine biomass was not estimated. Instead, I characterized their abundance by counting the number of vines of each species occurring on tree trunks within a 10 m radius plot. Then I expressed species abundance as vines/ha.

Table 1. Dry specific gravity ( $\mathrm{g} / \mathrm{cm}^{3}$ ) for large wood dethius and snags. Shownas mean $\pm$ standard deviationfor decay dasses 2 and 3 . Cass 1 values were obtaned fromHaygreen and Bowyer (1989). The speaific gravily of Nyssa sy/vatica was used for Myssa biflora of this study.

| Decay <br> Cass | Pinustaeda | Junipens <br> virginiana | Prunus <br> virginiana | Nyssa <br> sylvatica |
| :---: | :--- | :--- | :--- | :--- |
| 1 | 0.510 | 0.470 | 0.470 | 0.500 |
| 2 | $0.268 \pm 0.04$ | $0.439 \pm 0.02$ <br> $(n=6)$ | $(n=2)$ |  |
| 3 | $0.268 \pm 0.04$ |  |  |  |
| $(n=6)$ |  |  |  |  |$\quad$| $0.262 \pm 0.12$ |
| :--- |
| $(n=4)$ |

In retrospect, I attempted to estimate vine biomass by using a California study by Gartner (1991) on poison oak (Toxicodendron diversilobum L.). The study did not provide a regression equation, but included wood density measurements that I had been unable to find anywhere else. I estimated biomass for the forest only because it had more numerous and larger vines than the transition. I used the wood density of 0.5 $\mathrm{g} / \mathrm{cm}^{3}$ of naturally supported vines and a mean basal diameter of 2 cm from Gartner (1991). Based on my observations at Phillips Creek, I estimated the average vine height to be half of tree height. Therefore, I gave all vines on my forest sites a height of 7 m , determined by taking half of the average tree height in this study. Thus estimated vine biomass was a product of the number of vines in a 10 m radius plot and the assigned vine dimensions. The average contribution of estimated vine biomass to the total biomass of forest sites was $0.4 \%$. The amount of error associated with the vine estimates could not be ascertained. However the $0.4 \%$ average contribution was so small that even if it were $1 \%$, vine biomass was unimportant to the overall aboveground biomass of the forest.

Shrub regression equations. A range of size classes of Myrica cerifera, Baccharis halimifolia, and Iva frutescens shrubs were harvested on the eastern shore of Virginia to develop regression equations to estimate biomass from basal and crown diameters and from height. Measurements were recorded for each specimen before cutting.

Sometimes basal diameter was not attainable due to multiple basal stems, especially for Iva. Shrubs were placed in labeled plastic garbage bags.

At the lab, shrubs were cut to fit into preweighed, labeled paper bags. They were then placed in a large drying oven $85^{\circ} \mathrm{C}$ for 4 or 5 d until there was no further weight loss. Dried shrubs were weighed to the nearest 0.01 g , and heavier shrubs to the
nearest 100 g . A simple linear regression was created for each shrub species using the Microsoft Excel 97 statistical package. Biomass was the dependent variable (y) with height, crown diameter, and basal diameter as the independent variables (x). Biomass was matched with each independent variable to see which one yielded the highest $R^{2}$.

With Iva frutescens only, the regression line was forced through the origin for both independent variables (crown diameter and shrub height), so that the resulting equations would give positive values. Though the regression which used shrub height had the lower $R^{2}$ value, its equation gave positive biomass values $(\mathrm{g})$ for smaller shrubs (<75 cm height) (Table 2).

Tree biomass estimations. In my literature review of tree biomass studies, I
discovered that multiple regression equations were derived from above-stump biomass of harvested tree species, and that their methods excluded the stump height of 6 to 21 cm (Nelson and Switzer 1975, Clark et al. 1983, Van Lear et al. 1984). Therefore, my tree biomass estimations $(\mathrm{kg})$ using equations from the literature exclude stump biomass. The independent variables used to predict above-stump biomass (kg) are dbh for J. virginiana (Schnell 1976), and both dbh and height for P. taeda and hardwood species (Nelson and Switzer 1975, Clark et al. 1983)

Tree heights less than 10 m were estimated with a 1.5 m PVC pole by placing the pole against the tree for a sight estimate of the number of pole lengths to the tree top. Taller tree heights were estimated with aid of a clinometer that measured angles in percent degrees at a measured horizontal distance from the tree. Calculations were made using the following formula obtained from a forestry catalogue for level sites: tree height $(\mathrm{ft})=(\% \text { degree of tree top }+\% \text { degree of tree base })^{*}$ distance $(\mathrm{ft})$ from base of tree. For example, at a distance of 85 ft from tree base, the tree top angle is $80 \%$

Table 2. Shrub regressions developed to estimate biomass
(kg) of Myrica cerifera, Iva frutescens, and Baccharis halimifolia.

| Species | X -variable | Equation | $\mathrm{R}^{2}$ |
| :--- | :--- | :--- | :--- |
| M. cerifera | $\mathrm{BD}(\mathrm{cm})$ | $\mathrm{Y}=0.9034 \mathrm{X}-1.7367$ | 0.90 |
| I. frutescens | Height $(\mathrm{m})$ | $\mathrm{Y}=0.686 \mathrm{X}$ | 0.63 |
|  | $\mathrm{CD}(\mathrm{m})$ | $\mathrm{Y}=0.9184 \mathrm{X}$ | 0.69 |
| B. halimifolia | $\mathrm{BD}(\mathrm{cm})$ | $\mathrm{Y}=0.2806 \mathrm{X}-0.3843$ | 0.85 |
|  | $\mathrm{CD}(\mathrm{m})$ | $\mathrm{Y}=1.6608 \mathrm{X}-0.4613$ | 0.90 |
|  |  |  |  |
| Abbreviations <br> diameter. |  |  |  |

degrees and the base angle is $5 \%$ degrees. The equation would yield the following results: $(0.80+0.05) * 80 \mathrm{ft} .=68 \mathrm{ft}$.

Biomass estimates for $J$. virginiana came from a study by Schnell (1976) conducted in eastern and middle Tennessee. Since regression coefficients were not included with the regression equation in the publication, Schnell's (1976) table of computed dry weights for a dbh range of 5 to 20 inches were used (Table 3). Because most $J$. virginiana on the study area were between 2 and 7 cm dbh, I harvested three $J$. virginiana trees in that size range from the eastern shore of Virginia. A simple linear regression equation was developed using dbh and biomass data from the harvested trees, and included the biomass of a 5 inch red cedar from Schnell's study (1976). Thus, the biomass of 1 to 5 inch size class was estimated with the following equation: Y (biomass kg$)=3.0318 \mathrm{X}$ ( $R^{2}=0.77$ ), where $X=d b h(c m)$.

For $P$. taeda biomass estimations, a multiple regression equation was chosen from a study by Nelson and Switzer (1975) performed on 50-year old natural stands of loblolly pine growing on good and poor sites in Mississippi. Nelson and Switzer derived one regression formula for their sites, but the coefficients vary according to site index. Therefore, I used the coefficients for good sites to estimate tree biomass in the forest, and poor site coefficients for trees in the transition zone (Table 4).

For hardwood tree species, I used multiple regression equations from a study by Clark et al. (1983). Their study derived regression equations to predict weight and volume for major Coastal Plain hardwood species of all size classes in the southeastern USA. The two regression equations employed in this study incorporate dbh and total height as the independent variables to predict total tree dry weight (Tables 5 and 6).

Table 3. Juniperus virginiana estimated dry mass
based on dbh (From Schnell 1976).

| Dbh |  | Tree Mass |  |
| :---: | ---: | ---: | ---: |
| inches | cm | lbs | kg |
| 5 | 13 | 109 | 49.4 |
| 6 | 15 | 164 | 74.4 |
| 7 | 18 | 231 | 104.8 |
| 8 | 20 | 310 | 140.6 |
| 9 | 23 | 404 | 183.3 |
| 10 | 25 | 511 | 231.8 |
| 11 | 28 | 770 | 349.3 |
| 13 | 33 | 922 | 418.2 |
| 14 | 36 | 1090 | 494.4 |
| 15 | 38 | 1275 | 578.3 |
| 16 | 41 | 1475 | 669.1 |
| 17 | 43 | 1693 | 767.9 |
| 18 | 46 | 1929 | 875.0 |
| 13 | 48 | 2181 | 990.0 |
| 20 | 51 | 2452 | 1112.2 |

Table 4. Pinus taeda regression coefficients and equation for estimating tree biomass.
(adapted from Nelson and Switzer,

| Coefficients | a | b | c |
| :--- | :---: | :---: | :---: |
| Good sites | 0.0808 | 2.6774 | 0.7744 |
| Poor sites | 0.0158 | 2.6435 | 1.0119 |

Equation Log Weight (lbs) $=\mathrm{a}+\mathrm{b}$ Log dbh

Antilog of computed answer yields weight (lbs).

Table 5 and $6 \quad$ Table 5. Hardwood species biomass regression equation and species specific coefficients for species less than 11 inch dbh (adapted from Clark et al. 1983). Coefficients of Nyssa sy/vatica were used for Nyssa bilfora of this study.

| Species <11 in. | a | b |
| :--- | :---: | :---: |
| Liquidamber <br> styraciflua | 0.13234 | 0.94165 |
| Nyssa sylvatica | 0.16700 | 0.92799 |
| All species | 0.20334 | 1.90850 |
| Equation | $\mathrm{Y}=\mathrm{dbh}^{2}(\mathrm{in})^{*} \mathrm{Height}^{\mathrm{b}}(\mathrm{ft})^{*} \mathrm{a}$ |  |

Coefficients of determination $\mathrm{R}^{2}=0.99$

Table 6. Biomass (lbs) regression equation specific for Nyssa sylvatica greater than 11 inches dbh (adapted from Clark et al. 1983). This equation was used to estimate Nyssa bilfora biomass (kg) of this study.

| Species $>11 \mathrm{in}$. | a | b | c |
| :--- | :---: | :---: | :---: |
| Nyssa sylvatica | 0.6657 | 1.11305 | 0.92799 |
| Equation | $\mathrm{Y}=\mathrm{a}\left(\mathrm{dbh}^{2}(\mathrm{in})\right)^{\mathrm{b}} \mathrm{Height}^{\mathrm{c}}(\mathrm{ft})$ |  |  |

Coefficient of determination $R^{2}=0.99$

Accuracy of tree biomass estimates. Under and over estimates could have been made on vegetation biomass using regression equations depending on the standard error associated with the equations. Standard errors were listed in the biomass studies by Clark et al. (1983) on coastal hardwoods and Kapeluck and Van Lear (1995) on P. taeda roots. Using the standard error terms ( 0.08 and 0.11 lbs ) associated with specific hardwood tree regressions, I calculated the mean standard error (MSE) of hardwood trees at each site. Using the MSE, the upper and lower limit of tree biomass (kg) per site was determined. The MSE associated with hardwood tree biomass was so small that it did not change the tree biomass estimate expressed as $\mathrm{kg} / \mathrm{m}^{2}$ at any site. Unfortunately, regression standard errors were not listed in the biomass study of $P$. taeda (Nelson and Switzer 1975) or in the Juniperus virginiana study by Schnell (1976) who also failed to report a $R^{2}$ value for the equation. However, the $P$. taeda multiple regression for both good and poor sites has an $R^{2}=0.99$ which would indicate small regression errors. Because $P$. taeda and hardwoods are the most numerous tree species in the forest, the cumulative MSE associated with tree regressions probably remains small. On the other hand, the MSE of transition tree biomass remains unknown, because that zone is only comprised of $P$. taeda and $J$. virginiana tree species. .

Stump biomass of trees was estimated to determine if significant underestimates of tree biomass had been made using the above- stump regressions. Late in my the field study, I had measured basal diameters at 3-8 cm aboveground depending on tree size on a subsample of trees from the forest sites. I calculated differences between dbh and basal diameters of measured trees, and distinguished three categories of differences in basal diameters based on dbh. I added the diameter difference to dbh (i.e., 2 cm to a 3 cm dbh; 4 cm to $\leq 25 \mathrm{~cm}$ dbh; 12 cm to $>25 \mathrm{~cm}$ dbh) to produce a basal diameter for all
trees. Based on a study by Clark et al. (1983) that reported stump heights of harvested trees, I assigned tree stump heights ( 6 cm stump for $\mathrm{dbh} \leq 12 \mathrm{~cm}$ and 15 cm stump for $\mathrm{dbh}>12 \mathrm{~cm}$ ) to all trees in forest and transition zones. I used stump dimensions and specific gravity of tree species to estimate stump biomass for forest and transition sites. The difference between tree biomass $\left(\mathrm{kg} / \mathrm{m}^{2}\right)$ with and without stump was $2.3 \%$ and $2.6 \%$ for forest and transition zones, respectively. Because stump contributions did not change tree biomass $\left(\mathrm{kg} / \mathrm{m}^{2}\right)$, I chose not to add to the regression estimates.

## Belowground Organic Matter

Soil organic matter was sampled and processed for macroorganic matter (MOM) > 1 mm , and total soil organic matter (TOM). MOM consisted of herbaceous and woody roots and detritus, depending on zone. Also, MOM ( $\mathrm{kg} / \mathrm{m}^{2}$ ) was comprised of mixed species, except for the majority of sample sites in the low marsh zones where monospecific stands of Spartina alterniflora were growing. TOM $\left(\mathrm{kg} / \mathrm{m}^{2}\right)$ was estimated from cores that were collected at depths that encompassed the majority of soil organic matter and fine roots for each zone. For tidal creek and marsh zones, a 30 cm depth was usually sufficient. In the transition zone which had soils saturated at or near the surface, a 30 cm core depth captured the majority of fine roots, and large root biomass ( $>0.6 \mathrm{~cm}$ ) of Pinus taeda was estimated using a regression. For the forest, TOM was the sum of organic matter from 30 cm soil cores, large root biomass (estimated for $P$. taeda), and fine root biomass of 30 to 50 cm MOM cores. For the forest zone, I found that the majority of fine root biomass occurred within 40 cm of the ground surface. The techniques used for extraction and processing of MOM and TOM cores are explained below.

## Marsh Zones and Tidal Creek

Soil cores were extracted with a 7.6 cm diameter aluminum corer, similar to the one constructed by Gallagher (1974). Two 30 cm cores per sample site were taken at a depth that would encompass the majority of the rooting depth and soil organic matter. Core length and depth of soil organic matter was measured in the field. When organic rich soil exceeded the depth of 30 cm (which occurred at five sampling points), a McCauley peat sampler was used to determine to what depth it continued. A change in soil texture from the lightweight spongy organic matter to a heavier dark silty layer was considered the cut-off point for soil organic matter depth. Determining the cut-off point was sometimes difficult because a change in texture, interpreted as a silty mineral soil, in some cases proved to be highly decomposed organic rich soil (ascertained from loss on ignition). As a result, depth of soil organic matter was underestimated for approximately five high marsh sites.

Two cores were extracted per sampling point where vegetation previously had been harvested. All cores were measured upon extraction for total length, and depth of soil organic matter. My original intent was to obtain at least 30 cm depth per core. This was not always possible in the marsh zones, because some cores broke off during extraction, and reinsertion of coring device to obtain remainder of core was unsuccessful. On the other hand, six cores, each approximately 30 cm in length, were extracted from a $5 \mathrm{~m}^{2}$ area within the creek channel at low tide. All cores were wrapped in labeled aluminum foil after measurements were obtained, and subsequently stored in a lab freezer at $-4^{\circ} \mathrm{C}$ until further processing.

## Transition and Forest

Two cores were extracted from each of the five, $1 \mathrm{~m}^{2}$ quadrats where live vegetation and surface litter had previously been harvested. A 6.8 cm bucket auger was used because heavy clay soils would have made coring extremely difficult and time consuming. The auger bucket was marked at $10,20,30$, and 50 cm intervals to estimate sample depth. Hole depths also were measured for more exact estimates of core length. Core volumes, needed for bulk density calculations, could not be determined using the diameter of the bucket auger. Therefore, diameter and depth measurements were obtained from numerous holes augured to a depth of 30 cm in the transition and forest zones. The hole diameter was measured at 10 cm increments to a depth of 30 cm . The mean diameter was calculated for each depth. An ANOVA was used to calculate the difference between mean hole diameters for all depths. Since differences were insignificant ( $p>0.05$ ), the mean diameter of 8.2 cm was used to calculate the volume for every augured soil increment.

Bulk density samples were collected to a 30 cm depth at 10 cm intervals, and placed in labeled plastic bags. MOM samples in the transition also were collected to a total depth of 30 cm . This was not the case for the forest, where samples were collected to a total depth of 50 cm to capture most of the fine rooting depth. This was performed at 10 cm intervals to a depth of 30 cm , then followed by a $20 \mathrm{~cm}(30-50 \mathrm{~cm})$ interval. All soil samples were later stored at $-4^{\circ} \mathrm{C}$.

In the forest, depth (cm) of the organic horizon (litter layer had been removed) was measured from the sides of the augured holes by using visual and tactile approximation of change in soil texture and a centimeter ruler. This technique was not practical for the transition zone, because the depth of the organic horizon was too great to be assessed
in the same manner. Therefore, depth to gray mineral soil was approximated from the augered soil samples. A mean depth was determined for each site by averaging measurements obtained within the five quadrats.

## Loblolly Pine Root Regression

Kapeluck and Van Lear (1995) provided the only available regression equations for predicting total below stump dry biomass ( kg ) of roots $>0.06 \mathrm{~cm}$ diameter for $P$. taeda. The study was conducted on an eroded site in the piedmont region of South Carolina.

The regression equation, coefficients and correction factor (CF) are: dry root biomass = $a^{*} C F^{*} d b h^{b}\left(R^{2}=0.95\right)$, where $a=0.0152, b=2.5535$, and $C F=1.0139$.

The regression standard error for $P$. taeda tree roots ( $>0.6 \mathrm{~cm}$ ) was 0.07 kg . I calculated the mean standard error for each forest site, and found the error term too small to change the root biomass $\left(\mathrm{kg} / \mathrm{m}^{2}\right)$ estimate of any forest site. I did not calculate the mean standard error for transition sites, because the $P$. taeda tree root biomass $\left(\mathrm{kg} / \mathrm{m}^{2}\right)$ contributed only $2.8 \%$ to the grand total belowground mass of the forest zone. Overall, large root biomass was underestimated for both forest and transition zones for lack of regressions on hardwood species, J. virginiana, and shrubs, and to what extent is uncertain

## Laboratory Processing of Soil Cores

One core from each sample site within the low and high marsh zones was processed to determine bulk density, total organic matter, and total organic carbon. This core was divided into 10 cm segments ( $0-10,10-20$, and $20-30 \mathrm{~cm}$ ). However, some cores fell short of having the full three segments. If a segment was less than 10 cm , it was still processed, and the actual length recorded for volume calculations.

Each segment was placed in a preweighed, numbered aluminum pan, and the sample ID was recorded. Samples were placed in a drying oven at $100^{\circ} \mathrm{C}$, or $85^{\circ} \mathrm{C}$ if sharing the oven with vegetation samples (to prevent charring of vegetation), until which time there was no further weight loss, usually 4-5 d. Oven-dried cores were weighed to 0.01 g , and weights recorded. Core dry weights and volumes were used to determine bulk density ( $\mathrm{g} / \mathrm{cm}^{3}$ ).

Following weighing, the cores were pulverized into a fine powder using a mortar and pestle. Coarse roots were separated from the powder and later ground using a Wiley mill with a 40 -mesh screen. This root material was reintroduced to the soil and stirred to create a homogenous sample. This mixture was subsampled and stored in labeled plastic bottles. These samples were analyzed for loss on ignition (LOI) and total organic carbon (TOC).

Transition and forest bulk density samples were treated in the same fashion, except the soil samples had already been separated into segments in the field. Also, the transition samples, $10-20$ and $20-30 \mathrm{~cm}$, were partially pulverized and a subsample of 100 ml was taken for grinding into a fine powder. This was done to speed the processing of over 200 samples.

TOC may have been underestimated due to the processing of the soils. Loss of dissolved organic carbon (DOC) found in soil pore water may have occurred during the oven-drying of the soils cores, when it was possible for some of the DOC to volatilize as soil water evaporated. I investigated the possible amount of loss by using data from a study by Aiosa (1996) who measured DOC at Phillips Creek. I used the moisture contents of three high marsh soil samples $(0-10 \mathrm{~cm})$ from my study and the DOC value of 55 mg C/L from Aoisa's study (1996) to calculate DOC levels in the soil samples,
which is unlikely The estimated quantity of DOC in the three samples ranged from 0.004 $-0.005 \mathrm{~kg} / \mathrm{m}^{2}$. The quantity was minute compared to the TOC content of each sample. The difference in TOC mass would not have exceeded 0.09\%, if the total amount of DOC had been lost from the samples.

MOM (Gallagher 1974) was processed from the second core obtained from each sample site in the low and high marsh zones. Each core was divided into 10 cm segments in the same manner as the bulk density samples. Each segment was washed over a 18 mesh (1 mm) brass screen in a laboratory sink until free of sediment. The samples were placed into preweighed, labeled aluminum pan, and dried at $85^{\circ} \mathrm{C}$, until no further weight loss occurred, usually 2-3 d.

Forest and transition MOM samples contained pebbles and sand $>1 \mathrm{~mm}$ in their mineral soil portion. Thus, an extra washing step was required. This involved placing each washed MOM sample > 10 cm depth into 1 L of water to float the vegetation, and separate the pebbles and sand.

All oven-dried MOM samples were weighed to the nearest 0.01 g , and subsequently ground using a Wiley mill with a 40-mesh screen. Each ground sample was stored in a labeled plastic bottle for LOI analysis.

## Laboratory Analyses

Loss on ignition was determined for vegetation and litter, and TOM and MOM samples from all zones. The following procedure was standard for the types of samples mentioned above. Triplicates of ground sample material were weighed to the nearest 0.00001 g. However, some MOM samples were replicated twice or not at all when there was not enough sample material.

Chipped sample material of SW, cones, herbaceous vegetation, all shrub species, and $J$. virginiana underwent LOI. Each component was replicated three or four times, and weighed to the nearest 0.00001 g .

All samples were incinerated at $500^{\circ} \mathrm{C}$ for 180 min in a muffle furnace, and reweighed to determine ash-free dry weight (AFDW). Percent organic matter (OM) was calculated, and reported.

CHN analyses of soils was necessary to obtain an actual measure of total organic carbon, and thus determine the error in LOI that might have occurred due to loss of structural water in soil samples with high clay content. Fifty soil samples, representative of all zones except the tidal creek, were processed with a Leeman Labs CE 440 CHN analyzer. A simple linear regression was constructed using percent organic carbon (OC) and percent organic matter (OM) from sample data. The slope of the line was 2.09 and $R^{2}=0.99$ with percent OC predicting percent OM. Thus, correction of soil LOI values was not necessary (Figure 5). A curvilinear relationship existed between bulk density (ranging from $0.05-1.75 \mathrm{~g} / \mathrm{cm}^{3}$ ) and percent OC values obtained from 49 of the CHN samples. This curvilinear relationship was similar to that of Gosselink et al. (1984), except that their study estimated bulk density from OC values. Using SPSS 6.1 statistical program, a regression was formulated using a cubic line as the best fit $\left(R^{2}=\right.$ 0.91 ) (Figure 6). The following cubic regression formula was the result: $\mathrm{Y}=\mathrm{a}$ $+b_{1} X+b_{2} X^{2}+b_{3} X^{3}(S E=2.7)$, where $Y=$ percent $O C, X=$ bulk density, $a=39.74, b_{1}=-$ $98.90, b_{2}=85.12$ and $b_{3}=-23.96$. This equation appears to be useful for predicting percent OC from a wide range of bulk density values.


Figure 5. Linear regression of percent organic carbon from CHN analyses plotted against percent organic matter from loss on ignition.

Calculations for Organic and Carbon Mass
For high marsh vegetation, LOI data was similar for D. spicata, S. patens, J. roemerianus (living \& dead), and non-woody litter. Therefore, the data were pooled for a mean organic matter (OM) value of $92 \%$. For the low marsh zones, LOI data for $S$. alterniflora (living) had a mean OM value of $83 \%$, and the combination of S. alterniflora (dead) and low marsh non-woody litter had a mean OM of $68 \%$. Thus, OM mass ( $\mathrm{kg} / \mathrm{m}^{2}$ ) was the product of mean percent OM and herbaceous vegetation dry mass $\left(\mathrm{kg} / \mathrm{m}^{2}\right)$.

Because herbaceous vegetation and non-woody litter samples from the transition zone did not undergo LOI analyses, the mean of $92 \%$ OM for the high marsh was applied to transition herbaceous vegetation and litter. I considered this to be justifiable in that the two zones share many of the same species. Shrubs, forest herbaceous vegetation, and detritus had OM values of $98 \%$. The OM of $J$. virginiana was also $98 \%$, and I considered this value to be representative of all tree species.

Total organic matter (TOM) ( $\mathrm{kg} / \mathrm{m}^{2}$ ) for soil samples from all zones was the product of percent TOM, bulk density, and core length. For marsh samples missing an entire 10 cm increment (due to coring difficulties), TOM values $\left(\mathrm{kg} / \mathrm{m}^{2}\right)$ were assigned that reflected the mean value for zone and depth. Consequently, the following mean TOM values were used: (1) for the high marsh zone, $10-20 \mathrm{~cm}=10.95 \mathrm{~kg} / \mathrm{m}^{2}$ and $20-30 \mathrm{~cm}=$ $9.25 \mathrm{~kg} / \mathrm{m}^{2}$ (2) for the LMSS zone, $20-30 \mathrm{~cm}=7.11 \mathrm{~kg} / \mathrm{m}^{2}$ (3) and, for LMTS, 20 to 30 $\mathrm{cm}=6.56 \mathrm{~kg} / \mathrm{m}^{2}$.

MOM oven-dry mass $\left(\mathrm{kg} / \mathrm{m}^{2}\right)$ equals dry mass $(\mathrm{kg})$ divided by core volume $\left(\mathrm{m}^{3}\right)$ and multiplied by core length ( m ). Subsequently, MOM ash-free dry weight (AFDW) $\left(\mathrm{kg} / \mathrm{m}^{2}\right)$ was a product of MOM oven-dry mass $\left(\mathrm{kg} / \mathrm{m}^{2}\right)$ and percent OM. Missing MOM


Figure 6. Cubic regression of bulk density predicting percent organic carbon.
increments were dealt with in a similar manner by assigning a mean MOM AFDW value $\left(\mathrm{kg} / \mathrm{m}^{2}\right)$ based on zone and depth.

Organic carbon mass $\left(\mathrm{kg} / \mathrm{m}^{2}\right)$ of all vegetation, detritus, and soil MOM and TOM was considered to be $50 \%$ of their AFDW $\left(\mathrm{kg} / \mathrm{m}^{2}\right)$.

## Statistical Analyses

I used Microsoft Excel 97 software for descriptive statistics and to develop simple linear regression equations to estimate biomass of shrubs and small Juniperus virginiana (< 13 cm dbh). Later I changed to a statistical program (SPSS 6.1) to analyze the relationships of organic carbon and organic matter and bulk density and organic carbon using simple linear and cubic regression analyses (Figures 5 and 6). The SPSS
6.1 statistical program also had the capacity to run one-way ANOVA tests, Kruskall

Wallace non-parametric tests, and independent t-tests.
A one-way ANOVA and least significant difference (LSD) posthoc test were used to compare mean above and belowground biomass, OM, and OC mass $\left(\mathrm{kg} / \mathrm{m}^{2}\right)$ across zones. I chose the LSD test over Tukey or Bonferonni for detecting mean differences, because it is a more powerful analysis for small sample sizes. With every one-way ANOVA test, I ran a test for homogeneity of variance. If the variance was significant, I followed with a Kruskall Wallace nonparametric test to check the validity of the one-way ANOVA results. In all cases the Kruskall Wallace supported the test results of the oneway ANOVA, so I confidently reported the ANOVA test results.

I used independent t-tests to compare mean mass of vegetation components sampled only in the forest and transition zones, such as shrubs, large and small wood detritus, and snags.

## RESULTS

## GPS Survey

A GPS survey of all sample sites, except the tidal creek, provided coordinates and elevation that were incorporated into a digitized georeferenced aerial photo of the study area (Figure 7). Elevations relative to mean sea level for sites in the five zones were as follows: the five forest sites ranged from 1.20-2.96 m; the 14 transition sites ranged from 0.73-1.24 m ; the 25 high marsh sites ranged from $0.70-1.23 \mathrm{~m}$; the nine low marsh short-form Spartina alterniflora (LMSS) sites ranged from 0.67-0.85 m; and, the seven low marsh tall-form Spartina alterniflora (LMTS) sites ranged from 0.27-0.52 m.

## Aboveground Biomass

The organic carbon (OC) mass data of aboveground vegetation, used in one-way ANOVA and independent t-test analyses, were derived from Tables 7a and 7b

Total aboveground OC mass was highest in the forest and followed by transition, high marsh, LMTS, and LMSS. The tidal creek zone is assumed to be zero because plants were absent. Not surprisingly, the forest OC mass was significantly greater than all other zones, and likewise the transition was significantly greater than the marsh zones $\left(F_{(4,55)}, \mathrm{p}=0.000\right)$. Organic carbon mass was not significantly different among the marsh zones (Figure 8a). The mean OC mass of herbaceous vegetation exhibited a different pattern. The high marsh contained the highest mass followed by transition, LMTS LMSS, and forest. The forest was significantly less than all zones, except LMSS.


Figure 7. Georeferenced USDA (1990) aerial photograph of Phillips Creek study area displaying sample sites within all zones, except the tidal creek

Table 7a. Organic carbon mass $\left(\mathrm{kg} / \mathrm{m}^{2}\right)$ of aboveground vegetation for forest and transition sites shown as mean $\pm$ standard error. Abbreviations: LW = large wood detritus and SW = small wood detritus.

| Zone | Site | $\begin{gathered} \text { Trees } \geq \\ 2.54 \mathrm{~cm} \\ \text { dbh } \end{gathered}$ | Woody veg. <2.54 cm dbh | Shrubs | Snags | LW | SW | Cones | Herb. Veg. | Herb. Litter | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Forest | Fa18 | 9.04 | 0.00 | 0.00 | 0.00 | 0.138 | 0.22 | 0.075 | 0.046 | 1.20 | 10.73 |
| Forest | Fa30 | 15.06 | 0.00 | 0.00 | 0.00 | 0.012 | 0.06 | 0.040 | 0.003 | 0.90 | 16.07 |
| Forest | Fb29 | 9.64 | 0.00 | 0.10 | 0.05 | 0.000 | 0.21 | 0.030 | 0.015 | 0.39 | 10.44 |
| Forest | Fb47 | 11.59 | 0.00 | 0.00 | 0.65 | 0.000 | 0.36 | 0.023 | 0.004 | 0.91 | 13.54 |
| Forest | Fb65 | 9.64 | 0.00 | 0.09 | 0.17 | 0.000 | 0.19 | 0.065 | 0.023 | 0.85 | 11.03 |
| Mean |  | $11.0 \pm 1.1$ | 0.00 | $0.1 \pm 0.1$ | $0.4 \pm 0.3$ | $0.06 \pm 0.06$ | $0.42 \pm 0.09$ | $0.10 \pm 0.02$ | $0.02 \pm 0.01$ | $0.9 \pm 0.1$ | $12.4 \pm 1.1$ |
| Transition | Ta1 | 0.09 | 0.000 | 0.02 | 0.09 | 0.000 | 0.001 | 0.000 | 0.50 | 0.14 | 0.84 |
| Transition | Ta4 | 0.15 | 0.000 | 0.01 | 0.01 | 0.050 | 0.000 | 0.000 | 0.56 | 0.21 | 0.98 |
| Transition | Ta13 | 0.11 | 0.027 | 0.10 | 0.03 | 0.000 | 0.011 | 0.001 | 0.41 | 0.16 | 0.84 |
| Transition | Tb29 | 0.00 | 0.000 | 0.11 | 0.00 | 0.000 | 0.000 | 0.000 | 0.72 | 0.07 | 0.89 |
| Transition | Tc2 | 0.14 | 0.000 | 0.02 | 0.01 | 0.000 | 0.007 | 0.000 | 0.45 | 0.15 | 0.77 |
| Transition | Tc8 | 2.58 | 0.003 | 0.48 | 0.09 | 0.025 | 0.129 | 0.001 | 0.28 | 0.32 | 3.90 |
| Transition | Tc18 | 1.30 | 0.000 | 0.39 | 0.08 | 0.005 | 0.051 | 0.000 | 0.52 | 0.10 | 2.45 |
| Transition | Tc31 | 0.31 | 0.000 | 0.02 | 0.20 | 0.012 | 0.076 | 0.000 | 0.25 | 0.11 | 0.98 |
| Transition | Td2 | 1.80 | 0.000 | 0.26 | 0.15 | 0.105 | 0.158 | 0.006 | 0.37 | 0.23 | 3.09 |
| Transition | Td9 | 1.49 | 0.000 | 0.27 | 0.51 | 0.041 | 0.020 | 0.000 | 0.53 | 0.11 | 2.97 |
| Transition | Td15 | 2.44 | 0.000 | 0.03 | 0.03 | 0.093 | 0.006 | 0.000 | 0.45 | 0.07 | 3.12 |
| Transition | Td33 | 0.37 | 0.000 | 0.00 | 0.03 | 0.000 | 0.010 | 0.000 | 0.47 | 0.13 | 1.01 |
| Transition | Td39 | 2.25 | 0.000 | 0.06 | 0.20 | 0.093 | 0.104 | 0.000 | 0.37 | 0.26 | 3.34 |
| Transition | Td50 | 0.22 | 0.000 | 0.06 | 0.02 | 0.000 | 0.020 | 0.000 | 0.74 | 0.53 | 1.60 |
| Mean |  | $1.0 \pm 0.3$ | 0.00 | $0.3 \pm 0.1$ | $0.2 \pm 0.1$ | $0.06 \pm 0.02$ | $0.08 \pm 0.03$ | 0.00 | $0.47 \pm 0.04$ | $0.3 \pm 0.1$ | $1.9 \pm 0.3$ |

Table 7b. Organic carbon mass ( $\mathrm{kn} / \mathrm{m}^{2}$ ) of aboveground vegetation for sites within the marsh zones shown as mean $\pm$ standard error. Abbreviations are LMSS = low marsh shortform Spartina alterniflora and LMTS = low marsh tall-form $S$. alterniflora.

| Zone | Site | Herb. Veg. | Herb. Litter |
| :--- | :---: | :---: | :---: |
|  |  | Total OC |  |
|  |  |  |  |
| High Marsh Ha5 | 0.58 | 0.13 | 0.71 |
| High Marsh Ha14 | 0.57 | 0.14 | 0.71 |
| High Marsh Ha19 | 0.33 | 0.30 | 0.63 |
| High Marsh Ha21 | 0.22 | 0.16 | 0.38 |
| High Marsh Ha36 | 0.47 | 0.19 | 0.66 |
| High Marsh Ha41 | 0.42 | 0.10 | 0.53 |
| High Marsh Ha52 | 0.21 | 0.01 | 0.22 |
| High Marsh Hb1 | 0.39 | 0.26 | 0.65 |
| High Marsh Hb4 | 0.49 | 0.17 | 0.66 |
| High Marsh Hc1 | 0.66 | 0.11 | 0.77 |
| High Marsh Hc30 | 1.18 | 0.24 | 1.42 |
| High Marsh Hc37 | 0.02 | 0.01 | 0.03 |
| High Marsh Hc41 | 0.21 | 0.23 | 0.44 |
| High Marsh Hc42 | 1.26 | 0.23 | 1.50 |
| High Marsh Hc48 | 0.69 | 0.13 | 0.82 |
| High Marsh Hc72 | 0.48 | 0.11 | 0.59 |
| High Marsh Hc75 | 0.82 | 0.17 | 0.99 |
| High Marsh Hc84 | 0.32 | 0.21 | 0.53 |
| High Marsh Hc92 | 0.30 | 0.15 | 0.45 |
| High Marsh Ma1 | 1.25 | 0.14 | 1.38 |
| High Marsh Ma15 | 0.70 | 0.14 | 0.84 |
| High Marsh Mc2 | 0.22 | 0.08 | 0.29 |
| High Marsh Tb4 | 0.44 | 0.06 | 0.50 |
| High Marsh Tb14 | 0.37 | 0.16 | 0.53 |
| High Marsh Tb23 | 0.60 | 0.12 | 0.72 |
| Mean | $0.53+0.06$ | $0.15+0.01$ | $0.68+0.07$ |
| LMSS LSa8 | 0.15 | 0.03 | 0.18 |
| LMSS LSa12 | 0.29 | 0.04 | 0.33 |
| LMSS LSb2 | 0.22 | 0.03 | 0.25 |
| LMSS LSb16 | 0.34 | 0.01 | 0.35 |
| LMSS LSb17 | 0.34 | 0.14 | 0.48 |
| LMSS LSb46 | 0.25 | 0.14 | 0.39 |
| LMSS LSb53 | 0.23 | 0.04 | 0.27 |
| LMSS LSb55 | 0.27 | 0.06 | 0.33 |
| LMSS Mb4 | 0.17 | 0.11 | 0.28 |
| Mean | $0.25+0.02$ | $0.07+0.02$ | $0.32+0.03$ |
|  |  |  |  |

Table 7b. Completed

| Zone |  | Site | Herb. Veg. | Herb. Litter |
| :--- | :--- | :---: | :---: | :---: |
|  |  | Total OC |  |  |
|  |  |  |  |  |
| LMTS | LTS6 | 0.33 | 0.00 | 0.33 |
| LMTS | LTS9 | 0.43 | 0.14 | 0.56 |
| LMTS | LT6 | 0.67 | 0.07 | 0.74 |
| LMTS | LT12 | 0.61 | 0.00 | 0.61 |
| LMTS | LT25 | 0.46 | 0.00 | 0.46 |
| LMTS | LT29 | 0.22 | 0.03 | 0.25 |
| LMTS | LT33 | 0.48 | 0.05 | 0.53 |
| Mean |  | $0.46+0.06$ | $0.04+0.02$ | $0.50+0.06$ |



Figure 8. Organic carbon mass of (a) total aboveground vegetation and (b) herbaceous vegetation across zones. Shown with standard error bars. Different letters above the bars indicate significant differences ( $\mathrm{F}_{(4,55)}, \mathrm{P}=0.000$ ). LMSS = low marsh short-form Spartina alterniflora and LMTS = low marsh tall-form S. alterniflora.

Additionally, the LMSS zone was significantly less than the high marsh and transition $\left(F_{(4,55)}, P=0.000\right)$ (Figure 8b).

Shrub OC mass of the transition was higher than the forest, but the difference was insignificant ( $\mathrm{t}(17$ ), $\mathrm{P}=0.07$ ) (Figure 9). This result may be due to the fact that eight of the 14 transition sites had few shrubs. Small woody vegetation, comprised of small shrubs (< 2.54 cm basal diameter) harvested at only two transition sites, was not analyzed statistically; however, it was included in the calculation of total aboveground biomass (Table 7a).

Organic carbon mass of herbaceous litter was compared across zones. The forest contained the highest mass followed by transition, high marsh, LMSS, and LMTS. There were significant differences $\left(F_{(4,55)}, P<0.0000\right)$ in OC mass (Figure 10). ANOVA results of aboveground vegetation mass compared across five zones were summarized in Table 8.

Organic carbon mass of large wood (LW) and small wood (SW) detritus components was compared between forest and transition zones. The forest exceeded the transition in snag, and snag and LW combined, and the transition was slightly higher in LW.

However, the differences were not significant (t-test, 17, P>0.05), (Figures $11 \mathrm{a}, \mathrm{b}$, and c). In contrast, the forest greatly exceeded the transition in SW (pine cones included) (ttest, 17, $\mathrm{P}=0.000$ (Figure 12).

The distribution of vegetation mass $\left(\mathrm{kg} / \mathrm{m}^{2}\right)$ varied across zones (Figure 13). For the forest, the majority of mass was distributed in living trees (90\%). Herbaceous litter, SW, and snags combined comprised the remaining 10\%. The transition zone was the most evenly distributed: $48 \%$ of total biomass in living trees, $28 \%$ in herbaceous vegetation, $10 \%$ in herbaceous litter, and $12 \%$ in the sum of shrubs and snags.


Figure 9. Shrub organic carbon mass of forest and
transition zones with standard error bars. The difference is insignificant $(\mathrm{t}=-1.94,17, \mathrm{P}=0.07)$.


Figure 10. Organic carbon mass of herbaceous litter across zones. Bars with the same letter are not significantly different $\left(F_{(4,55)}, p=0.0000\right)$.

Table 8. One-way analysis of variance comparing total vegetation, herbaceous vegetation, and herbaceous litter in three weight categories across five zones.

| Factor D | Dependent variable | SS | MS | F P |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Zones 1 to $5^{\text {a }}$ | $5^{\text {a }}$ Total vegetation biomass | 4 | 55 | 2554.82 | 638.70 | 189.48 | 0.0000 |
|  | Total vegetation AFDW | 4 | 55 | 2482.44 | 620.61 | 190.54 | 0.0000 |
|  | Total vegetation OC | 4 | 55 | 620.84 | 155.21 | 194.06 | 0.0000 |
| Zones 1 to 5 | Herbaceous vegetation biomass Herbaceous vegetation AFDW Herbaceous vegetation OC | 4 | 55 | 6.46 | 1.61 | 6.27 | 0.0003 |
|  |  | 4 | 55 | 5.60 | 1.40 | 6.56 | 0.0002 |
|  |  | 4 | 55 | 1.40 | 0.35 | 6.57 | 0.0002 |
| Zones 1 to 5 | 5 Herbaceous litter | 4 | 55 | 9.97 | 2.49 | 42.34 | 0.0000 |
|  | Herbaceous litter AFDW | 4 | 55 | 10.10 | 2.53 | 49.96 | 0.0000 |
|  | Herbaceous litter OC | 4 | 55 | 2.71 | 0.68 | 18.75 | 0.0000 |

${ }^{\text {a }} 1$ to 5 refer to the following zones: 1 = forest, $2=$ transition, $3=$ high marsh, $4=$ low marsh short-
form Spartina alterniflora, and $5=$ low marsh tall-form S. alterniflora
Abbreviations: AFDW $=$ ash-free dry weight, $\mathrm{OC}=$ organic carbon, $\mathrm{DF}_{1}=$ degrees on freedom in numerator, and $\mathrm{DF}_{2}=$ degrees of freedom in denominator.


Figure 11. Average organic carbon mass of (a) snags, (b) large wood detritus, and (c) combined snags and large wood detritus of forest and transition zones. Shown with standard error bars. No significant differences with t -tests, $\mathrm{df}=17$ in all 3 categories.


Figure 12. Average organic carbon mass of small wood detritus (pine cones included) compared between two zones. Displayed with standard error bars. Differences were highly significant $(t=5.83,17, P=0.000)$.

As expected for the marsh zones, greater than $75 \%$ of total biomass was standing herbaceous vegetation and less than $25 \%$ in litter. The high marsh contained the highest mass followed litter, and $12 \%$ in the sum of shrubs and snags (Figure 13).

## Community Composition

Overall, the forest was dominated by a mature stand of Pinus taeda, and an understory of mixed hardwood species. However, two forest sites had hardwood species that shared the canopy with $P$. taeda, and one of these sites was dominated by Nyssa biflora. The subcanopy was also comprised of a few scattered Myrica cerifera shrubs on wetland sites. The sparse herbaceous layer was dominated by Smilax spp. and Rhus radicans, and the presence of graminoid species was rare (Table 9). The forest floor was covered by a thick layer of pine needles, leaves, and woody detritus. Bare soil was a rare occurrence.

In general, the transition had few trees limited to Juniperus virginiana and $P$. taeda, a frequent occurrence of shrubs relative to other zones, and an abundant mix of brackish and freshwater herbaceous vegetation (Table 9). The transition from forest to high marsh was highly variable among sites. An advanced stage of the transition appeared have high abundance of shrubs with snags but no living trees, or sites with few trees and shrubs situated on hummocks. A less advanced transition were sites dominated by mature $P$. taeda that were stunted, evidenced by a smaller tree dbh and height than $P$. taeda occurring in the forest zone.

Herbaceous species were prevalent throughout the marsh zones. Dominant high marsh species were Spartina patens, Juncus roemerianus, and Distichlis spicata.


Figure 13. Distribution of vegetation mass within each zone. Any component less than $1 \%$ was not represented. LMSS = low marsh short-form S . alterniflora, and LMTS = low marsh tall-form S. alterniflora. SW = small wood detritus, LW = large wood detritus, and Herb.Veg. = herbaceous vegetation

Table 9. Vascular plant species of Phillips Creek. Zones are designated by: $\mathrm{F}=$ forest, T=transition, HM= high marsh, LMSS=low marsh short-form Spartina, LMTS= low marsh tall-form Spartina. Nomenclature follows Gleason and Cronquist (1991).

| Family / Species | Common Name | F | Zone |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Poaceae |  |  |  |  |  |  |  |
| Panicum virgatum L. | Switch grass | X | x |  |  |  |  |
| Setaria genticulata Beauv. | Saltmarsh foxtail grass |  | X |  |  |  |  |
| Spartina patens (Aiton) Muhl | Saltmeadow hay |  | X |  | X |  |  |
| Distichlis spicata (L.) Greene | Salt grass |  | X |  | X | $x$ |  |
| Spartina alterniflora Loisel | Saltmarsh cordgrass |  |  |  | X | X | X |
| Phragmites australis (Cav.) Trin.Reed grass |  |  |  |  |  |  |  |
| Asteraceae |  |  |  |  |  |  |  |
| Aster tenuifolius L. | Saltmarsh aster |  |  |  | X |  |  |
| Solidago spp. | Goldenrod |  | X |  |  |  |  |
| Borrichia frutescens L. | Sea oxeye |  |  |  | X |  |  |
| Pluchea purpurascens (SW)DC. Saltmarsh fleabane |  |  | $x$ |  | X |  |  |
| Baccharis halimifolia L. | Groundsel tree |  | X |  |  |  |  |
| Iva frutescens L . | Marsh elder |  | X |  | X |  |  |
| Juncaceae |  |  |  |  |  |  |  |
| Juncus roemerianus Scheele | Black needlerush |  | X |  | X | X |  |
| Typhaceae |  |  |  |  |  |  |  |
| Typha augustifolia L. | Narrow-leaved cattail |  | X |  |  |  |  |
| Smilicaceae |  |  |  |  |  |  |  |
| Smilax bona-nox L. | Fringed greenbrier | X |  |  |  |  |  |
| Smilax glauca Walt | Sawbrier | x |  |  |  |  |  |
| Cyperaceae |  |  |  |  |  |  |  |
| Fimbristylis castanea (Michx.) |  |  |  |  |  |  |  |
| Vahl. | Marsh fimbristylis |  | X |  | X |  |  |
| Scirpus olneyi Gray | AmericanThree-square |  | X |  |  |  |  |
| Scirpus robustus Pursh | Saltmarsh bulrush |  | X |  |  |  |  |
| Eleocharis spp | Spike-rush |  | X |  |  |  |  |
| Gentianaceae |  |  |  |  |  |  |  |
| Limonium carolinianum (Walt.) |  |  |  |  |  |  |  |
| Britt. | Sea lavender |  |  |  | X |  |  |
| Chenopodiaceae |  |  |  |  |  |  |  |
| Scrophulariaceae |  |  |  |  |  |  |  |
| Agalinis spp. | Gerardia |  |  | x |  |  |  |
| Myricaceae |  |  |  |  |  |  |  |
| Myrica cerifera L. | Wax myrtle X | X |  |  |  |  |  |

Table 9. Completed.

| Family / Species | Common Name | Zone |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | F | T | HM | LMSS | LMTS |
| Malvaceae |  |  |  |  |  |  |
| Hibiscus moscheutos L. | Marsh hibiscus |  | X |  |  |  |
| Kosteletzkya virginica(L.) Perl. | Marsh mallow |  | X |  |  |  |
| Pinaceae |  |  |  |  |  |  |
| Pinus taedaL. | Loblolly pine | X | X |  |  |  |
| Cupressaceae |  |  |  |  |  |  |
| Juniperus virginianaL. | Red cedar | X | X |  |  |  |
| Rosaceae |  |  |  |  |  |  |
| Prunus serotinaEhrh. | Black cherry | X |  |  |  |  |
| Hamamelidaceae |  |  |  |  |  |  |
| Nyssaceae |  |  |  |  |  |  |
| Nyssa biflora(Walter) Sargent | Swamp blackgum | X |  |  |  |  |
| Aquifoliaceae |  |  |  |  |  |  |
| Fagaceae |  |  |  |  |  |  |
| Quercus phellos L. | Willow oak | X |  |  |  |  |
| Ulmaceae |  |  |  |  |  |  |
| Celtis laevigata Willd. | Hackberry | x |  |  |  |  |
| Araliaceae |  |  |  |  |  |  |
| Aralia spinosaL. | Hercules' club | X |  |  |  |  |
| Anacardiaceae |  |  |  |  |  |  |
| Rhus radicans L. | Poison ivy | X | x |  |  |  |
| Bignoniaceae |  |  |  |  |  |  |
| Campsis radicans (L.) Seem | Trumpet vine | X |  |  |  |  |
| Rubiaceae |  |  |  |  |  |  |
| Mitchella repens L. | Partridge berry | X |  |  |  |  |
| Aceraceae |  |  |  |  |  |  |
| Totals | 39 Species | 17 | 21 | 1 | 13 | 1 |

Also found in the high marsh was an occasional occurrence of live and dead shrubs, and tree snags. However, high shrub frequency (44 within a 5 m radius) occurred at one high marsh site. This anomaly may be partly explained by an abundance of fiddler crab burrows that aerate the soil and raised the soil surface elevation. The low marsh zones were dominated by $S$. alterniflora, but other herbaceous species were found there, including a patch of $J$. roemerianus (Table 9).

The transition zone with 21 species was higher in species richness than the forest with 17, but the forest had twice the number of woody species (Table 9). The loss of woody biomass in the transition also was reflected in tree basal area, and tree and woody vine densities. Tree basal area of the forest was 8 times greater, and stem density was 1.5 times greater than the transition (Tables 10 and 11). In addition, Campsis radicans, found in all forest sites, was absent in the transition, and Rhus radicans vines dropped sharply from a range of 318-2737 to 0-688 vines/ha (Table 12).

Percent cover of herbaceous species was greatest in the high marsh and transition, and lowest in the forest (Table 12). In the high marsh, S. patens and J. roemerianus were dominant species providing the highest range of cover (98-100\% and 85-100\% respectively). On the other hand, $D$. spicata was a dominant species providing the widest range of coverage in the high marsh, 38 to $100 \%$. S. alterniflora was the dominant species in the low marsh zones, but also dominated several high marsh sites. The percent cover range of $S$. alterniflora was lowest for LMTS and greatest for LMSS zone

Table 10. Forest tree stand table renresents all livind tree snecies $>2.54 \mathrm{~cm}$ dbh

| Species | Density (stems/ | Relative density | Basa rea ${ }^{2}$ / ha | Relative dominance |  | Relative | IV |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Pinus taeda | 306 | 31.37 | 32.23 | 76.81 | 1.0 | 21.74 | 43 |
| Prunus virginiana | 115 | 11.76 | 0.49 | 1.17 | 0.4 | 8.70 | 7 |
| llex opaca | 115 | 11.76 | 0.49 | 1.17 | 0.8 | 17.39 | 10 |
| Liquidambar styraciflua | 76 | 7.84 | 1.19 | 2.84 | 0.4 | 8.70 | 6 |
| Quercus phellos | 19 | 1.96 | 0.03 | 0.08 | 0.4 | 8.70 | 4 |
| Aralia spinosa | 13 | 1.31 | 0.01 | 0.02 | 0.2 | 4.35 | 2 |
| Celtis laevigata | 6 | 0.65 | 0.00 | 0.01 | 0.2 | 4.35 | 2 |
| Juniperus virginiana | 140 | 14.38 | 2.44 | 5.81 | 0.6 | 13.04 | 11 |
| Nyssa sylvatica | 178 | 18.30 | 4.83 | 11.52 | 0.4 | 8.70 | 13 |
| Acer rubrum | 6 | 0.65 | 0.24 | 0.58 | 0.2 | 4.35 | 2 |
| Total | 974 | 100.00 | 41.97 | 100.00 | 4.6 | 100.00 | 100 |

Table 11. Transition tree stand table rebresents all livina tree snecies $>2.54 \mathrm{~cm}$ dbh.

| Species | Density (stems) ha) | Relative Density | Basal area $\mathrm{m}^{2}$ / ha | Relative dominance | Frequency | Relative frequency | IV |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Juniperus virginiana | 473 | 68.44 | 2.77 | 54.03 | 1.0 | 56.52 | 60 |
| Pinus taeda | 218 | 31.56 | 2.36 | 45.97 | 0.8 | 43.48 | 40 |
| Total | 690 | 100.00 | 5.14 | 100.00 | 1.8 | 100.00 | 100 |

Table 12. Percent cover range of dominant herbaceous species from all sample sites. Abbreviations are LMTS = low marsh tall-form S. alterniflora and LMSS = low marsh short-form S. alterniflora

| Species | LMTS | LMSS | High marsh | Transition | Forest |
| :--- | :---: | :---: | :---: | :---: | :---: |
| Spartina patens | - | - | 98 to 100 | 68 to 98 | - |
| Distichlis spicata | - | - | 38 to 100 | 68 to 98 | - |
| Spartinaalterniflora | 15 to 38 | 38 to 68 | 68 to 85 | - | - |
| Juncus roemerianus | - | - | 85 to 100 | 68 to 85 | - |
| Smilax spp. | - | - | - | - | 15 to 38 |
| Rhus radicans | - | - | - | - | 3 to 15 |
| Mitchella repens | - | - | - | - | 3 to 15 |

## Belowground Organic Matter

The depth of organic rich soil (below litter layer) varied with zones. In the forest, the organic rich horizon, comprised of humus, roots and detritus, was approximately 4 to 7 cm thick and overlay a mineral horizon. In transition and high marsh zones, a spongy organic layer derived primarily from dead herbaceous root material (frequently referred to as peat) was found, and depths were highly variable. Transition varied from 5 to 20 cm , and the high marsh varied from 5 to greater than 70 cm . The low marsh zones and tidal creek were comprised of mineral sediment without an accumulated layer of organic rich soil.

## Macroorganic Matter (MOM)

MOM carbon mass was compared across five zones for each 10 cm increment and total 30 cm depth, and differences were significant for all depths (Table 13). In the top 10 cm , high marsh and LMSS had greater OC mass than forest, transition, and LMTS zones. At 10 to 20 cm , high marsh, LMSS, and LMTS zones were greater than the transition and forest. At the depth of 20 to 30 cm , MOM OC mass was greater in LMTS and high marsh than all other zones (Figure 14a). At a total depth of 30 cm , MOM OC mass had the following trend: high marsh $\left(3.88 \pm 0.29 \mathrm{~kg} / \mathrm{m}^{2}\right)>$ LMSS $(3.54 \pm 0.43$ $\left.\mathrm{kg} / \mathrm{m}^{2}\right)>$ LMTS $\left(2.46 \pm 0.33 \mathrm{~kg} / \mathrm{m}^{2}\right)>\operatorname{transition}\left(1.93 \pm 0.15 \mathrm{~kg} / \mathrm{m}^{2}\right)>$ forest $(1.47 \pm 0.22$ $\mathrm{kg} / \mathrm{m}^{2}$ ) (Figure 14b)

## Total Organic Carbon (TOC)

TOC mass was calculated as $50 \%$ of total organic matter mass (Table 14). TOC was compared across zones for each 10 cm increment of 30 cm soil cores, and

Table 13. One-way analysis of variance of macroorganic matter (MOM) organic carbon mass compared across five zones.

| Factor | Dependent variable | $\mathrm{DF}_{(1)} \mathrm{DF}_{(2)}$ | SS | MS | F | P |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Zones 1 to $5^{\mathrm{a}}$ | $0-10 \mathrm{~cm}$ depth | 4 | 52 | 11.20 | 2.80 | 9.11 | 0.0000 |
| Zones 1 to 5 | $10-20 \mathrm{~cm}$ depth | 4 | 52 | 7.48 | 1.87 | 7.18 | 0.0001 |
| Zones 1 to 5 | $20-30 \mathrm{~cm}$ depth | 4 | 52 | 1.63 | 0.41 | 3.32 | 0.0171 |
| Zones 1 to 5 | $0-30 \mathrm{~cm}$ depth | 4 | 53 | 50.22 | 12.56 | 9.97 | 0.0000 |

${ }^{\text {a }} 1$ to 5 refers to the following zones: 1 = forest, 2 = transition, $3=$ high marsh, $4=$ low marsh
short-form Spartina alterniflora and 5 = low marsh tall-form Spartinaalterniflora
(a)



Figure 14. (a) Macroorganic matter (MOM) carbon mass for three soil depths ( $F_{(4,52)}, P<0.02$ ) and (b) MOM carbon mass for total 30 cm depth $\left(\mathrm{F}_{(4,53)}, \mathrm{P}=0.000\right)$. Bar graphs are shown with standard error bars. Significant differences are indicated by different letters above the bars.
LMSS = low marsh short-form Spartina alterniflora and
LMTS = low marsh tall-form S. alterniflora.

Table 14. Belowground total organic carbon $\left(\mathrm{kg} / \mathrm{m}^{2}\right)$ in increments of 10 cm expressed as mean + standard error for all zones. The grand total includes the carbon mass of woody roots derived from root biomass regression for Pinus taeda and forest 30 to 50 cm macroorganic matter samples.

|  |  |  |  |  |  |  |  |
| :--- | :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| Zone | Site | 0 to 10 | 10 to 20 | 20 to 30 | 0 to 30 Woody Roots | Grand Total |  |
| Forest | Fa18 | 4.04 | 1.93 | 1.45 | 7.42 | 2.37 | 9.79 |
| Forest | Fa30 | 6.25 | 2.64 | 2.70 | 11.59 | 4.14 | 15.74 |
| Forest | Fb29 | 4.64 | 1.71 | 2.15 | 8.51 | 1.60 | 10.11 |
| Forest | Fb47 | 4.89 | 2.65 | 2.23 | 9.77 | 2.43 | 12.20 |
| Forest | Fb65 | 5.85 | 1.86 | 1.59 | 9.31 | 2.75 | 12.06 |
| Mean $\pm$ SE |  | $5.13 \pm 0.40$ | $2.16 \pm 0.02$ | $2.02 \pm 0.23$ | $9.32 \pm 0.69$ | $2.66 \pm 0.42$ | $11.98 \pm 1.06$ |
|  |  |  |  |  |  |  |  |
| Transition | Ta1 | 4.80 | 4.41 | 2.52 | 15.60 | 0.00 | 15.60 |
| Transition | Ta4 | 4.12 | 3.64 | 2.83 | 10.58 | 0.00 | 10.55 |
| Transition | Ta13 | 5.59 | 3.65 | 1.76 | 11.00 | 0.00 | 11.00 |
| Transition | Tb29 | 7.09 | 3.38 | 2.54 | 13.01 | 0.00 | 13.01 |
| Transition | Tc2 | 4.87 | 2.87 | 1.83 | 9.56 | 0.01 | 9.57 |
| Transition | Tc8 | 5.04 | 6.08 | 3.43 | 14.55 | 0.68 | 15.23 |
| Transition | Tc18 | 4.35 | 3.60 | 2.68 | 10.63 | 0.01 | 10.64 |
| Transition | Tc31 | 4.83 | 3.27 | 2.13 | 10.22 | 0.00 | 10.27 |
| Transition | Td2 | 5.57 | 3.28 | 2.12 | 10.97 | 0.46 | 11.43 |
| Transition | Td9 | 4.43 | 3.31 | 1.73 | 9.46 | 0.03 | 9.49 |
| Transition | Td15 | 3.73 | 3.04 | 1.77 | 8.53 | 0.11 | 8.64 |
| Transition | Td33 | 3.05 | 3.13 | 2.71 | 8.89 | 0.03 | 8.92 |
| Transition | Td39 | 4.11 | 2.93 | 2.41 | 9.44 | 0.11 | 9.55 |
| Transition | Td50 | 3.91 | 4.88 | 3.36 | 12.15 | 0.02 | 12.16 |
| Mean $\pm$ SE |  | $4.68 \pm 0.26$ | $3.68 \pm 0.24$ | $2.42 \pm 0.15$ | $10.77 \pm 0.45$ | $0.10 \pm 0.05$ | $10.87 \pm 1.06$ |

Table 14. Continued.

| Zone | Site | 0 to 10 | 10 to 20 | 20 to 30 | 0 to 30 | Grand Total |
| :--- | :--- | ---: | ---: | ---: | ---: | ---: |
| High marsh | Ha5 | 7.73 | 5.44 | 1.69 | 14.86 | 14.86 |
| High marsh | Ha14 | 6.12 | 9.29 | 4.64 | 20.03 | 20.03 |
| High marsh | Ha19 | 1.71 | 3.40 | 3.38 | 8.49 | 8.49 |
| High marsh | Ha21 | 3.55 | 5.48 | 4.63 | 13.65 | 13.65 |
| High marsh | Ha36 | 4.99 | 4.80 | 4.63 | 14.42 | 14.42 |
| High marsh | Ha41 | 5.45 | 4.45 | 3.52 | 13.42 | 13.42 |
| High marsh | Ha52 | 4.88 | 8.36 | 7.44 | 20.69 | 20.69 |
| High marsh | Hb1 | 4.36 | 9.71 | 4.63 | 18.69 | 18.69 |
| High marsh | Hb4 | 4.63 | 6.44 | 7.08 | 18.15 | 18.15 |
| High marsh | Hc1 | 5.15 | 4.82 | 4.46 | 14.43 | 14.43 |
| High marsh | Hc30 | 3.35 | 3.90 | 4.14 | 11.39 | 11.39 |
| High marsh | Hc37 | 3.35 | 4.02 | 4.77 | 12.14 | 12.14 |
| High marsh | Hc41 | 3.78 | 4.53 | 3.93 | 12.24 | 12.24 |
| High marsh | Hc42 | 4.06 | 3.80 | 5.44 | 13.31 | 13.31 |
| High marsh | Hc68 | 4.01 | 4.54 | 4.10 | 12.65 | 12.65 |
| High marsh | Hc72 | 3.59 | 3.72 | 3.96 | 11.27 | 11.27 |
| High marsh | Hc75 | 5.56 | 5.24 | 3.55 | 14.35 | 14.35 |
| High marsh | Hc84 | 4.68 | 5.84 | 2.15 | 12.67 | 12.67 |
| High marsh | Hc92 | 5.74 | 4.97 | 3.82 | 14.53 | 14.53 |
| High marsh | Ma1 | 3.33 | 5.24 | 3.93 | 13.20 | 13.20 |
| High marsh | Ma15 | 3.37 | 3.55 | 3.13 | 10.04 | 10.04 |
| High marsh | Mc2 | 5.42 | 4.03 | 3.73 | 13.19 | 13.19 |
| High marsh | Tb4 | 3.96 | 2.41 | 2.95 | 9.32 | 9.32 |
| High marsh | Tb14 | 3.91 | 1.71 | 1.94 | 7.56 | 7.56 |
| High marsh | Tb23 | 4.02 | 5.00 | 3.07 | 12.09 | 12.09 |
| Mean $\pm$ SE |  | $4.46 \pm 0.25$ | $5.04 \pm 0.39$ | $4.03 \pm 0.28$ | $13.56 \pm 0.67$ | $13.56 \pm 0.67$ |

Table 14. Completed.

| Zone | Site | 0 to 10 | 10 to 20 | 20 to 30 | 0 to 30 | Grand Total |
| :--- | :--- | ---: | ---: | ---: | ---: | ---: |
| LMSS | LSa8 | 5.46 | 4.26 | 3.56 | 13.28 | 13.28 |
| LMSS | LSa12 | 4.36 | 3.82 | 3.56 | 11.73 | 11.73 |
| LMSS | LSb2 | 4.03 | 4.22 | 3.89 | 12.13 | 12.13 |
| LMSS | LSb16 | 2.53 | 2.09 | 2.44 | 7.05 | 7.05 |
| LMSS | LSb17 | 5.89 | 3.09 | 1.96 | 10.93 | 10.93 |
| LMSS | LSb46 | 6.07 | 4.07 | 4.06 | 14.19 | 14.19 |
| LMSS | LSb53 | 4.95 | 3.94 | 2.77 | 11.67 | 11.67 |
| LMSS | LSb54 | 5.11 | 4.39 | 6.23 | 15.73 | 15.73 |
| LMSS | Mb4 | 5.55 | 4.77 | 3.56 | 13.88 | 13.88 |
| Mean $\pm$ SE |  | $4.88 \pm 0.37$ | $3.85 \pm 0.27$ | $3.56 \pm 0.41$ | $12.29 \pm 0.8212 .29 \pm 0.82$ |  |
| LMTS | LT6 | 4.20 | 4.10 | 3.98 | 12.28 | 12.28 |
| LMTS | LT12 | 3.29 | 3.71 | 3.28 | 10.27 | 10.27 |
| LMTS | LT25 | 3.69 | 2.77 | 3.46 | 9.93 | 9.93 |
| LMTS | LT29 | 4.41 | 4.39 | 3.28 | 12.08 | 12.08 |
| LMTS | LT33 | 2.32 | 2.94 | 1.69 | 6.96 | 6.96 |
| LMTS | LTS6 | 4.69 | 4.02 | 3.28 | 11.99 | 11.99 |
| LMTS | LTS9 | 4.03 | 3.99 | 3.97 | 11.99 | 11.99 |
| Mean $\pm$ SE |  | $3.80 \pm 0.30$ | $3.70 \pm 0.23$ | $3.28 \pm 0.29$ | $10.79 \pm 0.73$ | $10.79 \pm 0.73$ |
| Tidal Creek | C1 |  |  |  | 2.49 | 3.63 |
| Tidal Creek | C2 | 2.65 | 3.02 | 2.27 | 2.90 | 8.83 |

for total 30 cm depth (Figure 15a). Soil TOC to a 30 cm depth had the following trend $\left(\right.$ mean $\pm$ SE) : high marsh $\left(13.6 \pm 0.69 \mathrm{~kg} / \mathrm{m}^{2}\right)>$ LMSS $\left(12.3 \pm 0.82 \mathrm{~kg} / \mathrm{m}^{2}\right)>$ transition $\left(10.8 \pm 0.45 \mathrm{~kg} / \mathrm{m}^{2}\right)=$ LMTS $\left(10.8 \pm 0.73 \mathrm{~kg} / \mathrm{m}^{2}\right)>$ forest $\left(9.3 \pm 0.69 \mathrm{~kg} / \mathrm{m}^{2}\right)>$ tidal creek $\left(8.7 \pm 0.29 \mathrm{~kg} / \mathrm{m}^{2}\right)$.

As mentioned earlier, the grand total of soil TOC was derived from 30 cm soil cores, $30-50 \mathrm{~cm}$ forest MOM cores, and regression estimates of $P$. taeda root biomass. TOC grand total had a different zone pattern than the 30 cm total. Expressed as mean $\pm \mathrm{SE}$ : high marsh $\left(13.6 \pm 0.67 \mathrm{~kg} / \mathrm{m}^{2}\right)>\operatorname{LMSS}\left(12.3 \pm 0.82 \mathrm{~kg} / \mathrm{m}^{2}\right)>$ forest $\left(12.0 \pm 1.06 \mathrm{~kg} / \mathrm{m}^{2}\right)$
$>\operatorname{transition}\left(11.1 \pm 0.47 \mathrm{~kg} / \mathrm{m}^{2}\right)>$ LMTS $\left(10.8 \pm 0.73 \mathrm{~kg} / \mathrm{m}^{2}\right)>$ tidal creek $(8.7 \pm 0.29$
$\mathrm{kg} / \mathrm{m}^{2}$ ) (Figure 15b). Table 15 summarizes the ANOVA results of soil TOC.
Another question pursued was what percentage of TOC is MOM OC (from 30 cm soil cores). The mean percentage was calculated for five zones, which had both TOC and MOM OC estimations. High marsh and LMSS zones were highest at 30 and 29\%, respectively. LMTS and transition followed at $23 \%$ and $18 \%$, and lastly, forest at $16 \%$.

For all zones, the percentage of MOM OC decreased with depth (Figure 16).

## Combined Above and Belowground

The total aboveground OC and the belowground grand total were combined for each site. The sum of above and belowground OC mass for each zone expressed as mean $\pm$ SE was as follows: forest $\left(24.3 \pm 2.11 \mathrm{~kg} / \mathrm{m}^{2}\right)>$ high marsh $\left(14.2 \pm 0.65 \mathrm{~kg} / \mathrm{m}^{2}\right)>$ transition $\left(12.8 \pm 0.60 \mathrm{~kg} / \mathrm{m}^{2}\right)=$ LMSS $\left(12.6 \pm 0.82 \mathrm{~kg} / \mathrm{m}^{2}\right)>\operatorname{LMTS}\left(11.3 \pm 0.72 \mathrm{~kg} / \mathrm{m}^{2}\right)>$ tidal creek ( $8.7 \pm 0.29 \mathrm{~kg} / \mathrm{m}^{2}$ ). As expected, forest OC mass greatly exceeded the other zones (Figure 17). These mean values are the estimated present day OC standing stock for each zone at Phillips Creek, and provide the information needed in the state
(a)



Figure 15. Belowground organic carbon from (a) 30 cm soil depth
compared across zones ( $F_{(5,56)}, P=0.001$ ), and $(b)$ a grand total which is the sum of 30 cm depth and woody roots (derived from $P$. taeda root
regression and 30-50 cm MOM cores), ( $F_{(5,56)}, P=0.009$ ).
Significance indicated by different letters above bars.

Table 15. One-way analysis of variance of total soil organic carbon compared across six zones. All depths represented.

| Factor | Dependent |  | $\mathrm{DF}_{(1)}$ | $\mathrm{DF}_{(2)}$ | SS | MS | F |
| :--- | ---: | :--- | :--- | :---: | :---: | :---: | :---: |
| Zones 1 to 6 | $0-10 \mathrm{~cm}$ depth | 5 | 56 | 13.46 | 2.69 | 2.36 | 0.0519 |
| Zones 1 to 6 | $10-20 \mathrm{~cm}$ depth | 5 | 56 | 49.37 | 9.87 | 5.43 | 0.0004 |
| Zones 1 to 6 | $20-30 \mathrm{~cm}$ depth | 5 | 56 | 33.02 | 6.60 | 5.85 | 0.0002 |
| Zones 1 to 6 | Total 0-30 cm | 5 | 56 | 157.55 | 31.51 | 4.81 | 0.0010 |
| Zones 1 to 6 | Grand total | 5 | 56 | 118.87 | 23.77 | 3.46 | 0.0086 |

${ }^{\text {a }} 1$ to 6 refers to the following: $1=$ forest, 2 = transition, 3 = high marsh, $4=$ low marsh short-form Spartina alterniflora, 5 = low marsh tall-form S. alterniflora, and $6=$ tidal


Figure 16. Percent MOM carbon mass of total organic carbon mass from 30 cm soil cores. Shown per unit depth across zones. LMSS = low marsh short-form S. alterniflora and LMTS = low marsh tall-form S. alterniflora.


Figure 17. Sum of total above and belowground organic carbon compared across zones. Shown with standard error bars. Significance indicated by different letters above the bars $\left(F_{(5,57)}, P=0.0000\right)$. Zone abbreviations are explained in Figure 16.
change model (Brinson et al. 1995) to address changes in organic matter distribution and turnover rates that are influenced by rising sea level.

## DISCUSSION

The main objectives of this study were to quantify and characterize above and belowground organic matter within each ecosystem state from forest to tidal creek, and to incorporate the organic carbon estimates into the state change model by Brinson et al. (1995). To achieve this, I quantified aboveground vegetation and detritus using harvest methods, and indirectly using regression equations and dimensional analysis. I also quantified belowground organic matter from soil cores, and by using regression analyses for woody roots $(>0.6 \mathrm{~cm})$ of $P$. taeda

In this section, I will compare qualitative and quantitative data from this study with other studies, and revisit my original hypothesis for Phillips Creek. Using data gathered in this study, I wi ll estimate gains and losses of organic carbon stock associated with state change. Finally, I will estimate organic carbon turnover rates of ecosystem states for different slope profiles at Phillips Creek. For the remainder of the discussion I will be referring to organic carbon simply as carbon

## Comparison With Other Coastal Marshes

The aboveground characteristics of the various states (zones) in the present study are consistent with a study by Ricker (1999) of the Virginia Coast Reserve (VCR) megasite, which consists of an area on the eastern shore of Virginia extending from Cape Charles to Wallops Island. Similar characteristics include: (1) a forest zone with a dominant cover of tree species, few shrubs, and little herbaceous groundcover, (2) a transition that has fewer, smaller trees comprised of two species $P$. taeda, and $J$. virginiana, and more shrubs, snags and halophytic and glycophytic grasses, and (3) marsh zones that are dominated by salt tolerant herbaceous species. Distribution of
plant biomass reflects the community composition of each zone. In the forest zone, living trees contribute the majority of biomass (90\%). In the transition zone, the percentage of tree biomass declines (48\%) and herbaceous vegetation, shrub, and snag mass contributions rise. In the marsh zones, saltmarsh vegetation comprises 100\% of the biomass.

Within the 56 ha study area at Phillips Creek, approximately 52\% of the land area was forest, $11 \%$ transition, $24 \%$ high marsh, $7 \%$ LMSS, and 6\% LMTS and tidal creek combined. The transition area at Phillips Creek was very similar to the average transition area (13\%) in Ricker's (1999) study. On the other hand, the high marsh area on the VCR megasite was larger (32\% vs. 24\%).

Soil organic matter depth and content increased along a continuum from forest to transition and high marsh, which was consistent with findings of Hmieleski (1994) and Ricker (1999). The LMSS and LMTS zones showed a decline in soil organic matter content relative to the high marsh. Belowground macroorganic matter (MOM) mass decreased with depth for all zones. Similarly, Blum (1993) found MOM to be highest in the first 10 cm of soil for short-form Spartina alterniflora in the low marsh. In addition, de la Cruz and Hackney (1977) found that 94\% of belowground productivity in Juncus roemerianus occurred in the top 20 cm . Soil bulk density (BD) values at 0 to 10 cm and 10 to 20 cm depths of present study were similar to Hmieleski's (1994) results, except for the upland forest $10-20 \mathrm{~cm}$ (Table 16). Average BD values for a 30 cm depth in the LMSS zone were within range of Blum's (1993) results at Phillips Creek, but my average BD values for the LMTS zone were nearly half her values. Similar to these two studies, soil BD increased with depth.

Table 16. Soil bulk density comparison with former Phillips Creek study (Hmieleski 1994) for forest to high marsh zones. Values $\left(\mathrm{g} / \mathrm{cm}^{3}\right)$ are expressed as mean $\pm S E$.

| Zone | Soil depth | Present study | Hmieleski <br> study |
| :--- | :---: | :---: | :---: |
| Forest Steep | $0-10$ | $0.50 \pm 0.04$ | $0.51 \pm 0.02$ |
| Forest Steep | $10-20$ | $1.00 \pm 0.07$ | $0.52 \pm 0.05$ |
| Forest Flat | $0-10$ | $0.38 \pm 0.04$ | $0.42 \pm 0.11$ |
| Forest Flat | $10-20$ | $0.75 \pm 0.05$ | $0.60 \pm 0.02$ |
| Transition | $0-10$ | $0.22 \pm 0.02$ | $0.35 \pm 0.04$ |
| Transition | $10-20$ | $0.72 \pm 0.04$ | $0.62 \pm 0.03$ |
| High marsh | $0-10$ | $0.20 \pm 0.04$ | $0.29 \pm 0.03$ |
| High marsh | $10-20$ | $0.50 \pm 0.10$ | $0.55 \pm 0.03$ |

Mean soil bulk density of forest upland sites (present study) were compared to the forest steep (1994 study). Likewise, forest wetland sites were compared to forest flat (1994 study). High marsh and transition (present study) were compared to the flat transect (1994 study).

## Comparison with Other Regions

Aboveground biomass of most saltmarsh species reported in this study were within range of studies conducted on the Atlantic and Gulf Coast regions (Table 17). White et al. (1978) from Louisiana and Stout (1978) from Alabama reported greater biomass of $S$. alterniflora (both forms), S. patens, and D. spicata than the Atlantic Coast studies. MOM dry mass of the various saltmarsh species in this study was compared to other studies on the Atlantic and Gulf Coasts (Table 18). S. alterniflora short-form and J. roemerianus had higher MOM mass than those reported by the Alabama and Florida studies. Spartina patens of this study had the widest range compared to other studies in Table 18. Overall, I found that aerial and MOM dry mass of saltmarsh species reported in this study fall within the range of variation reported in other studies.

## Comparison with Forests and Grasslands Studies

Forest carbon mass estimates at the Phillips Creek study area greatly exceeded that of a Pinus taeda forest and a hardwood forest of similar ages (Table 19). Van Lear and Kapeluck (1995) explained that the plantation site in their study had experienced severe soil erosion prior to conversion to forest. Therefore, the lack of soil fertility of the site may explain the low tree biomass. In a hardwood forest study by Whittaker et al.( 1974), productivity of the mesophytic forest was low compared to forests of similar environments. Thus, low forest productivity rates may have resulted in low tree biomass. My tree biomass estimates are most similar to the field studies reported by Olson et al. (1983) and that of Phillips and Shure (1990) in southern Appalachia. After comparing forest biomass estimated in this study with other studies, I am confident that the forest carbon estimate falls within normal range. In addition, marsh carbon mass

Table 17. Aboveground biomass $\left(\mathrm{kg} / \mathrm{m}^{2}\right)$ of saltmarsh species of the Atlantic and Gulf Coast states in the USA.

| Location | S. alterniflora tall-form | S. alterniflora short-form | Spartina patens | Distichlis spicata | Juncus roemerianus | Comments | Reference |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Virginia | 1.14 | 0.54 | 0.77 | 0.61 | 1.90 | End of season standing crop | Present study |
| Virginia |  |  |  |  | 1.50-2.70 | Harvest within Juncus patch | Brinson and Christian (1999) |
| Virginia |  |  | 0.80 |  | 1.69 | 1994 harvest | Tolley (1996) |
| Delaware | 1.04 | $\begin{aligned} & 0.55^{a} \\ & 0.65^{b} \end{aligned}$ | $\begin{aligned} & 0.67^{\mathrm{a}} \\ & 0.73^{\mathrm{b}} \end{aligned}$ | $0.52^{\text {b }}$ |  | Peak live with annual mean | Roman \& Daiber (1984) |
| N. Carolina | $1.20{ }^{\text {c }}$ |  |  |  |  | End of season standing crop | Stroud (1976) |
| N. Carolina |  |  |  | 0.96 | 1.17 | Peak live standing crop | Bellis \& Gaither (1985) |
| Florida |  | $\begin{aligned} & 0.70^{d} \\ & 0.33^{e} \end{aligned}$ |  | 0.58 | $\begin{aligned} & 1.24^{\mathrm{d}} \\ & 1.06^{\mathrm{e}} \end{aligned}$ | Peak and end of season standing crop | Kruczynski et al. (1978) |
| Alabama |  | 1.03 |  |  | 1.45 | Peak standing crop with annual mean | Stout (1978) |
| Louisiana | 1.47 |  | 2.19 | 1.16 | 1.96 | Peak live standing crop | White et al. (1978) |
| Range | 1.04-1.47 | 0.33-1.03 | 0.67-2.19 | 0.52-1.16 | 1.06-2.70 |  |  |

[^0]Table 18. Belowground macroorganic matter (MOM) dry mass $\left(\mathrm{kg} / \mathrm{m}^{2}\right)$ from Atlantic and Gulf Coast studies.

| Location | Species | Min | $\begin{gathered} \text { MOM } \\ \text { Max } \\ \hline \end{gathered}$ | Mean | Depth (cm) | Comments | Reference |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Virginia | Spartina alterniflora(TF) |  |  | 4.1-8.1 | 30 | biomass range | Present Study |
|  | Spartina alterniflora(SF) |  |  | 4.4-13.3 | 30 |  |  |
|  | Spartina patens ${ }^{\text {a }}$ |  |  | 5.4-16.6 | 30 |  |  |
|  | Juncus roemerianus ${ }^{\text {a }}$ |  |  | 7.3-11.2 | 30 |  |  |
|  | Distichlis spicata |  |  | 6.8-9.3 | 30 |  |  |
| Delaware ${ }^{\text {c }}$ | Spartina alterniflora (SF) ${ }^{\text {d }}$ | 13.7 | 19.3 |  | 35 |  | Roman \& Daiber (1984) |
|  | S. alterniflora (SF) ${ }^{\text {e }}$ | 9.9 | 14.3 |  | 35 |  |  |
|  | S. alterniflora(TF) ${ }^{\text {d }}$ | 4.7 | 12.4 |  | 35 |  |  |
|  | S. alterniflora(TF) ${ }^{\text {e }}$ | 4.1 | 9.4 |  | 35 |  |  |
|  | Spartina patens ${ }^{\text {d }}$ | 3.5 | 6.0 |  | 35 |  |  |
|  | Spartina patens ${ }^{\text {e }}$ | 0.6 | 4.7 |  | 35 |  |  |
| New Jersey | Spartina alterniflora(SF) |  |  | 11.2 | 30 |  | Smith et al. (1979) |
| $N$. Carolina | Distichlis spicata |  |  | 12.1 | 30 |  |  <br> Gaither (1985) |
|  | Juncus roemerianus |  |  | 11.1 | 30 |  |  |
| N. Carolina | Spartina alterniflora(TF) |  |  | 1.6-6.7 | 30 | biomass range | Reader \& Craft (1999) |
|  | Spartina alterniflora (SF) |  |  | 0.9-11.2 | 30 | biomass range |  |
| Florida | Juncus roemerianus |  |  | 5.1 | 20 |  | Kruczynski et al. (1978) |
| Alabama | Juncus roemerianus |  |  | 4.6 | 20 |  | Stout (1978) |
|  | Spartina alterniflora(SF) |  |  | 3.6 | 20 |  |  |

${ }^{\text {a }}$ MOM dry mass attributed to species included small amounts of Distichlis spicata, MOM dry mass attributed to species included small amounts of Spartina patens, ${ }^{\text {c }}$ Canary Creek site, ${ }^{d} 1975,{ }^{\text {e }} 1976$

Table 19. Aboveground biomass and carbon estimations $\left(\mathrm{kg} / \mathrm{m}^{2}\right)$ of temperate forests and grasslands.

| Forest Region | Vegetation Class | Forest Type | Tree Biomass | Carbon | Data source | Comments | Reference |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Atlantic coastal plain, USA | Coniferous evergreen | Pinus taeda (successional 55y) | 22.4 | 11.0 | Field study |  | Present study |
| Southern <br> Piedmont, USA | Coniferous evergreen | Pinus taeda (plantation 48 y ) | 13.2 | $(6.6)^{\text {b }}$ | Field study |  | Van lear \& Kapeluck (1995) |
| Gulf coastal plain, USA | Coniferous evergreen | Pinus taeda (plantation 25 y ) | 14.6 | (7.3) | Field study |  | Pehl et al. (1984) |
| New England, USA | Deciduous | Mixed hardwood (natural 50 y ) | 14.7 | (7.4) | Field study |  | Whittaker et al. (1974) |
| Southern Appalachia, USA | Deciduous | Mixed hardwood (natural) | 21.4 | (10.7) | Field study |  | Phillips and Shure (1990) |
| Atlantic \& Gulf coastal plains, USA | Deciduous | Mixed hardwood on wet flats (natural 60 y ) | 27.7 | (13.9) | Field study |  | Frederick et al. (1983) |
| Canada | Coniferous evergreen \& deciduous ${ }^{\text {a }}$ | * | * | 12.5 | Canadian forest inventories | Computer model | Kurtz and Apps (1999) |
| International | Coniferous evergreen | * | * | 11.4 | Field studies |  | Olson et al. (1983) |
| International | Coniferous evergreen | * | * | 6.0 | Global satellite | Computer model | Potter (1999) |

Table 19. Completed.

| Grassland <br> Region Vegetation <br> Class Biomass Carbon  Data source | Comments Reference |  |  |  |  |  |
| :--- | :--- | :---: | :---: | :--- | :--- | :--- |
| Atlantic coastal <br> plain, USA | Herbaceous <br> marsh | 0.87 | 0.38 | Field study | Present study |  |
| International | Grassland | $*$ | 0.48 | Field studies | Olson et al. (1983) |  |
| International | Grassland | $*$ | 0.15 | Global <br> satellite | NASA- <br> CASA <br> Codel | Potter (1999) |

${ }^{\text {a }}$ refers to vegetation class which is my interpretation of the authors' classification terms of
hardwood and softwood. * information was not reported. ${ }^{\text {b }}$ carbon data in parentheses are my calculations from the formula carbon mass $=$ tree biomass $\times 0.5$.
from the present study falls within grassland estimates of Potter (1999) and Olson et al. (1983).

Lastly, I compared total belowground carbon mass from present research with estimates from global studies of grasslands and forests. Marsh belowground carbon to a 30 cm depth in this study $\left(7.0-20.7 \mathrm{~kg} / \mathrm{m}^{2}\right)$ is similar to grassland studies of 1 to 2 m depth in Japan, USA, and Russia (13.3-26.3 kg/m²) (Schlesinger 1977). In addition, forest belowground carbon to a 30 cm depth from this study $\left(7.4-11.6 \mathrm{~kg} / \mathrm{m}^{2}\right)$ falls within range (5.6-24.0 kg/m ${ }^{2}$ ) of forest studies (swamp forests not included) encompassing depths of 0.2 to 1.3 m from Europe, Russia, Asia, and the USA (Schlesinger 1977).

## Total Above and Belowground Carbon

I had hypothesized that total above and belowground organic matter distribution would have the following trend: forest $>$ transition $>$ high marsh $=$ mid-marsh $>$ low marsh short-form $S$. alterniflora $>$ low marsh tall-form $S$. alterniflora $>$ tidal creek. As mentioned previously, I had combined mid-marsh with high marsh because vegetation and soil characteristics were indistinguishable, and there were only four sample sites. The results of this study for total above and belowground organic matter, expressed as carbon $\left(\mathrm{kg} / \mathrm{m}^{2}\right)$, revealed the following trend: forest $>$ high marsh $>$ transition $>\mathrm{LMSS}>$ LMTS $>$ tidal creek. As expected, the magnitude of aboveground carbon $\left(\mathrm{kg} / \mathrm{m}^{2}\right)$ of the forest made up for the low belowground carbon compared to other zones, except the tidal creek. The high marsh exceeded that of the transition zone by $2.7 \mathrm{~kg} / \mathrm{m}^{2}$ for total carbon, because of the higher belowground values in the marsh. However, the difference between the two zones was not statistically significant for total carbon.

The pattern of carbon loss or gain in the present study was consistent with the state change model of Brinson et al. (1995) (Figure 1). The forest loses woody vegetation and
gains soil organic matter during the transition to high marsh. The high marsh loses soil organic matter during the transition to low marsh, and a total loss of emergent vegetation occurs as low marsh becomes subtidal. The differences in carbon stock of each ecosystem state at Phillips Creek are shown in Figure 18. Both net losses and gains of carbon occur with state change. The greatest aboveground loss of carbon ( $11.6 \mathrm{~kg} / \mathrm{m}^{2}$; i.e., 10.4 from forest to transition and 1.2 from transition to high marsh) is incurred when forest completes the transformation to high marsh, during which tree material completely disappears except for large roots. When high marsh converts to LMSS, there is a loss of $1.3 \mathrm{~kg} / \mathrm{m}^{2} \mathrm{C}$ in soil organic matter to a 30 cm depth. This estimate is conservative considering that the thickness of organic rich soil in some areas of the high marsh is up to 1 m . These greater depths occur west of the transition finger, and include the area of high marsh experiencing extensive breakup and ponding. Also, a creek channel in this region is advancing headward into the high marsh. This region of high marsh is roughly 7 ha, and included 8 out of the 10 sample sites with soils of high organic content (18 to $80 \%$ from LOI ) throughout the 30 cm soil depth. If a $1 \mathrm{~m}^{2}$ area of organic rich soil, 1 m deep, were to completely oxidize, the high marsh would incur a loss of approximately 60 $\mathrm{kg} / \mathrm{m}^{2}$ of carbon. This estimate is based on a soil bulk density of $0.20 \mathrm{~g} / \mathrm{cm}^{3}$ and carbon content of $30 \%$. Thus, the net loss of belowground carbon between high marsh and LMSS can be far greater than previously stated. However, personal observations and the Ricker study (1999) of other marshes in the region indicate that deep peat deposits are not typical.


Figure 18. Organic carbon ( $\left(\mathrm{kg} / \mathrm{m}^{2}\right.$ ) losses and gains with ecosystem state change along a continuum fromforest to tiddal creek. LMSS = bw marsh short-form Sportina ailervifora and LMTS = low marsh tall-form S. alfernfiora.

## Future Changes

What changes will occur in carbon mass for each zone in response to sea level rise, and at what rate? To address this question, I created two profiles displaying migration of zones at Phillips Creek for high marsh and LMSS zones; one profile depicts a steep slope and the other a very gentle slope. I estimated mean slopes of the LMSS and high marsh zones of this study using distance between sites (measured from a field map) and GPS elevations of those sites. Based on slope data from Hmieleski (1994), Ricker (1999), Blum (1993), and my estimates, the steep profile was assigned a $1.0 \%$ slope for the forest and transition, a $0.3 \%$ slope for high marsh and LMSS zones, and a 4\% slope for LMTS. In the gentle slope profile, the forest to high marsh continuum uses a 0.05\% slope throughout, and the LMSS and LMTS zones remain the same. The elevation ranges of the different zones (obtained in the GPS survey) determined the placement of zones along a 1 m wide transect (Figures 19a and 20a). Using sea level rise (SLR) as the only driving force for state change, I projected Phillips Creek 80 y into the future with a SLR rate of $5 \mathrm{~mm} / \mathrm{y}$ ( 0.4 m rise), based on the Intergovernmental Panel on Climate Change (1996) best estimate for future sea level rise. A major assumption in my models is that migration of zones inland is unimpeded by artificial barriers such as roads. In my models, I elevated the zones 0.4 m (Figures 19b and 20b), determined the distance that underwent state change, and estimated the turnover of carbon based on zone carbon stocks from Figure 18. I focused changes in carbon stock and turnover rates for the LMSS to forest zones, and allowed those zone areas to change in response to sea level rise. I held the LMTS zone to a constant area ( $6 \mathrm{~m}^{2}$ ) from the water's edge for the present and future profiles, because tall-form S. alterniflora is typically found as a narrow zone along the creekbank, and unlikely to change substantially


Figure 19a. Profile of Phillips Creek steep slope at present year 2000. LMTS = low marsh tall-form Spartina alterniflora, LMSS = low marsh short-form S. alterniflora, and T= transition


Figure 19b. Profile of Phillips Creek steep slope at year 2080 after a $5 \mathrm{~mm} / \mathrm{y}$ sea level rise. Zone abbreviations: H = high marsh, LMSS = low marsh short-form Spartina alterniflora, and LMTS = low marsh tall-form S. alterniflora.


Figure 20a. Profile of Phillip Creek gentle slope at present year 2000. Zone abbreviations: LMSS = low marsh short-form Spartina alterniflora and LMTS = low marsh tall-form S. alterniflora.


Distance (m)
Figure 20b. Profile of Phillips Creek gentle slope at year 2080 after a $5 \mathrm{~mm} / \mathrm{y}$ sea level rise. Zone abbreviations: LMSS = low marsh short-form Spartina alterniflora and LMTS = low marsh tall-form S. alterniflora.

Changes occurred in the area of the zones following an 80 y period of marsh migration, and the following profiles show the influence of the terrestrial margin slope on the future loss or gain of marsh area. In the steep profile, the marsh zones experienced a large shift where LMSS replaced the transition zone, and high marsh and transition replaced a forest edge. The LMSS and high marsh experienced a loss of zone area, because they migrated from a $0.3 \%$ to a $1 \%$ upland slope. In contrast, migration of transition and forest zones across the same slope retained the same amount of area (Figure 19b). In the gentle slope profile, the LMSS zone expanded, because the upland slope to which it migrated was less steep (from $0.3 \%$ to $0.05 \%$ ) (Figure 20b).

High marsh to forest retained the same amount of area after migration, as slope was held constant.

In both steep and gentle profiles of Phillips Creek, there was $100 \%$ turnover of carbon for all zones, except the forest by the year 2080 (Tables 20 and 21). The forest of the steep slope profile had only a $25 \%$ turnover (Table 20). In contrast the forest in the gentle slope profile experienced a $71 \%$ turnover, being replaced by high marsh and transition zones (Table 21). Future zone carbon stocks were reflected in the losses and gains of zone area. Carbon standing stock per zone was determined with the following formula: kg C/zone $=$ zone distance $(\mathrm{m})^{*} 1 \mathrm{~m}$ zone width * initial carbon standing stock $\left(\mathrm{kg} / \mathrm{m}^{2}\right)$ (Tables 20 and 21). In the steep profile, the migration of LMSS which began with 844 kg C/zone was reduced to 252 kg C/zone following an 80 y migration, and high marsh experienced a loss of 653 kg carbon standing stock (Figure 21). In contrast for the gentle slope profile, the LMSS carbon standing stock increased from 844 to 5040 $\mathrm{kg} /$ zone as it migrated from a $0.3 \%$ to a $0.05 \%$ slope (Figure 22).

Table 20. Phillips Creek steep slope profile ( 1 m wide transect) with organic carbon turnover projected to occur due to state change over an 80 y period with a $5 \mathrm{~mm} / \mathrm{y}$ sea level rise. Abbreviations: $\mathrm{C}=$ carbon, $\mathrm{LMTS}=\mathrm{low}$ marsh tall-form Spartina alterniflora and LMSS = low marsh short-form S. alterniflora.

| Zone | Initial C standing stock ( $\mathrm{kg} / \mathrm{m}^{2}$ ) | Distance <br> (m) | Original zone C stock (kg/zone) | C (kg) loss per zone in $80 y$ | \% C <br> turnover of zone/ 80y | C turnover rate \%/y | Zone replacement | Future C standing stock ( $\mathrm{kg} / \mathrm{m}^{2}$ ) | Distance (m) | $\begin{aligned} & \frac{\mathrm{C}(\mathrm{~kg} / \text { zone })}{} \\ & \hline \text { *Original/ } \\ & \text { Future } \end{aligned}$ | Future zone C stock (kg/zone) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| LMTS | 11.3 | 6 | 68 | 68 | 100 | 1.25 | Tidal creek | 8.7 | 6 | $0 \quad 52$ | 52 |
| LMSS | 12.6 | 67 | 844 | 844 | 100 | 1.25 | Tidal creek | 8.7 | 67 | 0583 | 583 |
| High marsh | 14.2 | 66 | 937 | 937 | 100 | 1.25 | Tidal creek / LMTS | $\begin{array}{r} 8.7 \\ 11.3 \end{array}$ | $\begin{gathered} 60 \\ 6 \end{gathered}$ | $\begin{array}{lr} 0 & 522 \\ 0 & 68 \end{array}$ | $\begin{array}{r} 522 \\ 68 \end{array}$ |
| Transition | 12.8 | 20 | 256 | 256 | 100 | 1.25 | LMSS | 12.6 | 20 | 0252 | 252 |
| Forest | 24.3 | 161 | 3912 | 978 | 25 | 0.31 | High marsh / Transition | $\begin{aligned} & 14.2 \\ & 12.8 \end{aligned}$ | $\begin{aligned} & 20 \\ & 20 \end{aligned}$ | $\begin{array}{ll} 0 & 284 \\ 0 & 256 \end{array}$ | $\begin{aligned} & 284 \\ & 256 \end{aligned}$ |
|  |  |  |  |  |  |  | Forest migration | 24.3 | 40 | 2934978 | 3912 |

* Original represents the residual carbon from the previous ecosystem state. Zeros for all but the forest indicate that all original carbon was removed.

Table 21. Phillips Creek gentle slope profile ( 1 m wide) with organic carbon turnover projected to occur due to state change over an 80 y period with a $5 \mathrm{~mm} / \mathrm{y}$ sea level rise. Abbreviations: $\mathrm{C}=$ carbon, LMTS $=$ low marsh tall form Spartina alterniflora and LMSS = low marsh short-form S. alterniflora.

| Zone | Initial C standing stock $\left(\mathrm{kg} / \mathrm{m}^{2}\right)$ | $\begin{gathered} \text { Distance } \\ (\mathrm{m}) \end{gathered}$ | $\begin{array}{r} \text { Original } \\ \text { zone C } \\ \text { stock } \\ (\mathrm{kg} / \text { zone }) \end{array}$ | C (kg) loss per zone in $80 y$ | $\% \text { C }$ <br> Turnover of zone/ $80 y$ | C Turnover rate \%/y | Zone replacement | Future C standing stock D $\left(\mathrm{kg} / \mathrm{m}^{2}\right)$ | Distance (m) | C (kg /zone) Original/Future |  | Future zone C stock (kg/zone) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| LMTS | 11.3 | 6 | 68 | 68 | 100 | 1.25 | Tidal Creek | 8.7 | 6 | 0 | 52 | 52 |
| LMSS | 12.6 | 67 | 844 | 844 | 100 | 1.25 | Tidal Creek | 8.7 | 67 | 0 | 583 | 583 |
| High marsh | 14.2 | 400 | 5680 | 5680 | 100 | 1.25 | Tidal Creek/ LMTS | $\begin{aligned} & 8.7 \\ & 11.3401 \end{aligned}$ |  | $\begin{aligned} & 0 \\ & 0 \end{aligned}$ | $\begin{array}{r} 3489 \\ 68 \end{array}$ | $\begin{array}{r} 3489 \\ 68 \end{array}$ |
| Transition | 12.8 | 400 | 5120 | 5120 | 100 | 1.25 | LMSS | 12.6 | 400 | 0 | 5040 | 5040 |
| Forest | 24.3 | 1120 | 27216 | 19323 | 71 | 0.89 | High marsh/ Transition | $\begin{aligned} & 14.2 \\ & 12.8 \end{aligned}$ | $\begin{aligned} & 400 \\ & 400 \end{aligned}$ | $\begin{aligned} & 0 \\ & 0 \end{aligned}$ | $\begin{aligned} & 5680 \\ & 5120 \end{aligned}$ | $\begin{aligned} & 5680 \\ & 5120 \end{aligned}$ |
|  |  |  |  |  |  |  | Forest migration | 24.3 | 1120 | 7893 | 19323 | 27216 |


| 4 | $\begin{gathered} 4 \% \\ 0.3 \text { \% } \end{gathered}$ | Year 2000 | Year 2080 |
| :---: | :---: | :---: | :---: |
|  |  | TC | TC |
|  |  | LMTS 68 |  |
| S | 0.3 \% | LMSS 844 |  |
| 0 |  | H 937 | LMTS 68 |
| p | 1\% | T 256 | LMSS 252 |
| s |  |  | H 284 |
| $\downarrow$ | 1\% |  | T 256 |
|  |  | F 3912 | F 2934 |
|  |  |  | $\begin{gathered} \text { F migration } \\ 978 \end{gathered}$ |

Figure 21. Projected state change following 80 y of $5 \mathrm{~mm} / \mathrm{y}$ sea level rise for Phillips Creek steep profile. Original and future carbon stocks (kg/zone) are shown for marsh to forest zones along a 1 m wide transect. All zones migrate but LMSS and H migrate up a steeper slope and loss of zone area occurs. Approximately $25 \%$ of the old forest stock has turned over. TC = tidal creek, LMTS = low marsh tall-form Spartina alterniflora, LMSS = low marsh shortform S . alterniflora, $\mathrm{H}=$ high marsh, $\mathrm{T}=$ transition, and $\mathrm{F}=$ forest.


Figure 22. Projected state change following 80 y of $5 \mathrm{~mm} / \mathrm{y}$ sea level rise for Phillips Creek gentle slope profile. Original and future carbon stocks (kg/zone) are shown for marsh to forest zones. All zones migrate landward, and LMSS expands in area as it migrates across a less steep slope. Approximately 71\% of the original forest stock has turned over. Abbreviations are explained in Figure 21.

Next, I compared horizontal carbon turnover rates associated with state change and vertical carbon turnover rates (net primary productivity (NPP) rate/ standing stock) of forest and marsh zones that occur in a normal growing season (Table 22). These NPP data are based on mean values of data from Tables 23a and 23b (except for the forest), and carbon standing stock values from Figure 18. To convert the oven-dry mass from these studies to carbon mass, I used the carbon mass to oven-dry mass ratios from this study (high marsh: $46 \%$ aerial biomass and $45 \%$ MOM; low marsh zones: $41 \%$ for both aerial biomass and MOM). The forest values are the sum of aboveground production of temperate evergreen forests (Barbour et al. 1980) and woody root production (Whittaker and Woodwell 1969). Horizontal carbon turnover rate (\%/y) = carbon loss (kg/zone/y) / carbon standing stock (kg/zone) * 100. In comparison, horizontal carbon turnover rates that occurred in both profiles as a result of state change are an order of magnitude lower than vertical carbon turnover rates associated plant production during the growing season (Table 22). If the estimated carbon standing stocks $\left(\mathrm{kg} / \mathrm{m}^{2}\right)$ for each ecosystem state were over or under estimated, how would horizontal carbon turnover rates be affected? Because carbon standing stock $\left(\mathrm{kg} / \mathrm{m}^{2}\right)$ are in both numerator and denominator of the horizontal turnover equation, under or over estimates have no effect. Having compared above and belowground biomass and carbon estimates from this with others, I have found my estimates to be in range with most other studies. Therefore, I am confident that the estimated carbon standing stocks $\left(\mathrm{kg} / \mathrm{m}^{2}\right)$ of ecosystem states at Phillips Creek are valid, and may be useful estimates for other coastal areas of similar geomorphology.

Table 22. Comparison of horizontal organic carbon turnover rate due to state change and vertical carbon turnover (net primary production/
standing stock) in marsh and forest zones. Steep and gentle refer to the two state change profiles for Phillips Creek. Abbreviations: LMTS = low marsh tall-form Spartina alterniflora and LMSS = short-form S. alterniflora.

|  | Horizontal Turnover <br> $(\% / \mathrm{y})$ <br> Steep <br> Gentle | Standing <br> Stock <br> $\left(\mathrm{kg} / \mathrm{m}^{2}\right)$ | Net Primary <br> Productivity <br> $\mathrm{kg} / \mathrm{m}^{2} / \mathrm{y}$ | Vertical <br> turnover $(\% / \mathrm{y})$ |  |
| :--- | :---: | :---: | :---: | :---: | :---: |
| Zone | 1.25 | 1.25 | 11.30 | $1.66^{\mathrm{a}}$ | 14.69 |
| LMTS | 1.25 | 1.25 | 12.60 | $1.55^{\mathrm{a}}$ | 12.30 |
| LMSS | 1.25 | 1.25 | 14.20 | $2.01^{\mathrm{a}}$ | 14.15 |
| High marsh |  |  |  |  | 3.21 steep <br>  <br> Forest ${ }^{\mathrm{b}}$ |
|  | 0.31 | 0.89 | 24.30 | 0.78 | 3.21 gentle |

${ }^{a}$ Net primary production is the sum of mean aerial and MOM organic carbon values from Tables 23 a and b .
${ }^{b}$ Forest net primary production - Sum of aboveground from Barbour et al. (1980) and belowground from Whittaker and Woodwell (1969).

Table 23a. Aerial net primary production $\left(\mathrm{kg} / \mathrm{m}^{2} / \mathrm{y}\right)$ of marsh plants

|  |  | Net Primary <br> Production <br> $\mathrm{kg} / \mathrm{m}^{2} / \mathrm{y}$ | NPP C mass <br> $\mathrm{kg} / \mathrm{m}^{2} / \mathrm{y}$ | Methods |
| :--- | :---: | :---: | :---: | :--- |

Table 23b. Macroorganic matter (MOM) productivity studies of the Atlantic and Gulf Coast states. Abbreviations: LMTS = low marsh tall-form Spartina alterniflora and LMSS = low marsh short-form S. alterniflora

| Zone | Species | Net Primary Productivity $\mathrm{kg} / \mathrm{m}^{2} / \mathrm{y}$ | Net Carbon Productivity $\mathrm{kg} / \mathrm{m}^{2} / \mathrm{y}$ | Reference |
| :---: | :---: | :---: | :---: | :---: |
| LMTS | Spartina alterniflora | 6.50 | $2.67^{\text {a }}$ | Roman \& Daiber (1984) |
| LMTS | Spartina alterniflora | 0.68 | 0.28 | Blum (1993) |
| LMTS | Spartina alterniflora | 2.11 | 0.87 | Gallagher et al. (1980) |
| LMTS | Spartina alterniflora | 2.46 | 1.01 | Dame and Kenny (1986) |
| Carbon mean $\pm$ SE |  | $1.20+0.51$ |  |  |
| LMSS | Spartina alterniffora | 5.45 | 2.23 | Dame and Kenny (1986) |
| LMSS | Spartina alterniflora | 2.02 | 0.83 | Gallagher and Plumley (1979) |
| LMSS | Spartina alterniflora | 5.00 | 2.05 | Roman \& Daiber (1984) |
| LMSS | Spartina alterniflora | 2.30 | 0.94 | Smith et al. (1979) |
| LMSS | Spartina alterniflora | 2.14 | 0.88 | Blum (1993) |
| Carbon mean $\pm$ SE |  | $1.31 \pm 0.27$ |  |  |
| High marsh | Spartina patens | 3.30 | 1.49 | Roman \& Daiber (1984) |
| High marsh | Juncus roemerianus | 1.36 | 0.61 | De La Cruz \& Hackney (1977) |
| High marsh | Juncus roemerianus | 7.58 | 3.41 | Stout (1978) |
| High marsh | Distichlis spicata | 0.90 | 0.41 | Bellis \& Gaither (1985) |
| High marsh | Distichlis spicata | 3.25 | 1.46 | Bellis \& Gaither (1985) |
| Carbon <br> mean $\pm$ SE |  |  |  |  |
|  |  | $1.48 \pm 0.44$ |  |  |

${ }^{\text {a }}$ Net primary production of MOM was converted to net carbon production by using the mean ratios of MOM carbon mass to MOM dry mass obtained from data in present study (Appendix F). The ratios were 0.41 for S. alterniflora and 0.45 for all other marsh species.

These carbon turnover rates presented are based the effects of sea level rise ( $5 \mathrm{~mm} / \mathrm{y}$ ) alone. Factors that can accelerate state change (storm frequency, fires, erosion and land subsidence) and factors involved in the resistance to state change (vertical accretion of sediment and organic rich soils, structures of plant community) were held constant. Based on sea level rise, this study offers insight into the temporal and spatial scales of carbon turnover for different landscape slopes, and can be finetuned with addition of factors mentioned above.

Ecosystem state change is a spatial process involving the transformation of coastal landscapes with noticeable changes within the span of a century. The major assumption presented, that no restrictions to overland migration of marsh and forest exist, is unrealistic. In many regions, human development of coastal areas with construction of structures (buildings, parking lots, and roads) can prevent the migration of marshes and forests, and contribute to their demise. Possible exceptions are areas that are protected, such as wildlife refuges and the Virginia Coast Reserve where this study occurred. In these protected areas overland migration may occur unimpeded within the confines of their land areas. High production of plant biomass in these coastal ecosystems provides food and habitat for a wide array of organisms. Unfortunately, protected coastal land areas are not abundant, and major losses of ecologically important coastal ecosystems may be inevitable.

Carbon standing stock $\left(\mathrm{kg} / \mathrm{m}^{2}\right)$ was estimated for each ecosystem state along a continuum from forest to tidal creek, and inserted into the state change model (Brinson et al. 1995). Forest was higher compared to all other zones, followed by high marsh, transition, low marsh short-form Spartina alterniflora, low marsh tall-form S. alterniflora and tidal creek. During the process of ecosystem state change, carbon losses generally exceeded gains. For both gentle and steep slope migration profiles, the marsh and transition zones underwent an estimated $100 \%$ carbon turnover (i.e., complete replacement) following 80 y of sea level rise. In the same time period, forest turnover was $25 \%$ and $71 \%$ for the steep and gentle slope profiles, respectively.

Yearly turnover rates associated with net primary production were called "vertical turnover", because the exchange of carbon is a vertical process that occurs between the atmosphere and plant community within an ecosystem. Whereas, "Horizontal turnover" is associated with state change where one ecosystem replaces another as it moves horizontally across the landscape. Vertical turnover rates of carbon range between $3.1 \% / \mathrm{y}$ for the forest and 14.7\%/y for the LMTS zone. In comparison, horizontal carbon turnover rates for all zones are slower by approximately one order of magnitude than vertical turnover rates. Though slower, horizontal turnover is a spatial process that can change the face of coastal landscapes within the span of a century.

## LITERATURE CITED

Aiosa, D. J. 1996. Microbial metabolism of DOC. Masters Thesis, University of Virginia, Charlottesville, VA. 196 pp.

Barbour, M. G., J. H. Burk and W. D. Pitts. 1980. Terrestrial Plant Ecology. Benjamin/Cummings Publishing Company, Inc., Menlo Park, CA.

Bellis, V. J. and A. C. Gaither. 1985. Seasonality of aboveground and belowground biomass for six salt marsh plant species. The Journal of the Elisha Mitchell Scientific Society 101:95-109

Bertness, M. D. 1988. Peat accumulation and the success of marsh plants. Ecology 69: 703-713.

Blum, L. K. 1993. Spartina alterniflora root dynamics in a Virginia marsh. Marine Ecological Progress Series 102:169-178.

Brinson, M. M. and R. R. Christian. 1999. Stability of Juncus roemerianus patches in a salt marsh. Wetlands 19(1):65-70.

Brinson, M. M., R. R. Christian, and L. K. Blum. 1995. Multiple states in the sea-level induced transition from terrestrial forest to estuary. Estuaries 18:648-659

Christiansen, T. 1998. Sediment deposition of a tidal salt marsh. Ph.D. Dissertation. University of Virginia, Charlottesville, VA.

Clark, A., D. R. Phillips and D. J. Frederick. 1983. Weight, volume, and physical properties of major hardwood species in the Gulf and Atlantic Coastal Plains. USDA Forest Service, Southeastern Experiment Station, Research Paper SE-250.

Dame, R. F. and P. D. Kenny. 1986. Variability of Spartina alterniflora primary production in the euhaline North Inlet estuary. Marine Ecology Progress Series 32:71-80.

Downs, L. L., R. J. Nicholls, S. P. Leatherman and J. Hautzenroder. 1994. Historic evolution of a marsh island: Bloodsworth Island, Maryland. Journal of Coastal Research 10:1031-1044.

Daubenmire, R. 1959. A canopy coverage method of vegetational analysis. Northwest Science 33:43-64.

De la Cruz and C. T. Hackney. 1977. Energy value, elemental composition, and productivity of belowground biomass of a Juncus tidal marsh. Ecology 58:11651170.

Frederick, D. J., A. Clark, and D. R. Douglas. 1983. Biomass, nutrient and energy requirements of coastal plain hardwoods. pp. 139-147 In The hardwood resource and its utilization: where is it going? $11^{\text {th }}$ Annual Hardwood Symposium of the Hardwood Research Council.

Gallagher, J. L. 1974. Sampling macro-organic matter profiles in salt marsh plant root zones. Soil Science Society of America Proceedings 38:154-155.

Gallagher, J. L. and F. G. Plumley. 1979. Underground biomass profiles and productivity in marsh sediments. Southeastern Geology 15:17-28.

Gallagher, J. L., R. Rheimold, R. A. Linthurst, and W. J. Pfeiffer. 1980. Aerial production, mortality and mineral accumulation-export dynamics in Spartina alterniflora and Juncus roemerianus plant stands in a Georgia salt marsh. Ecology 61:303-312.

Gardner, L. R., B. R. Smith and W.K. Michener 1992. Soil evolution along a forest-salt marsh transect under a regime of slowly rising sea level, southeastern United States. Geoderma 55:141-157.

Gartner, B. L. 1991. Structural stability and architectural of vines vs. shrubs of poison oak, Toxicodendron diversilobum. Ecology 72:2005-2015.

Gleason, H. A. and A. Cronquist 1991. Manual of vascular plants of the northeastern United States and adjacent Canada. $2^{\text {nd }}$ edition. The New York Botanical Garden, Bronx, NY, USA. 910 pp.

Gosselink, J. G., R. Hatton and C. S. Hopkinson. 1984. Relationship of organic carbon and mineral content to bulk density in Louisiana marsh soils. Soil Science 137:177180.

Hackney, C. T. and J. W. Cleary. 1987. Saltmarsh loss in southeastern North Carolina lagoons: importance of sea level rise and dredging. Journal of Coastal Research 3:93-97.

Hayden, B. P., R. D. Dueser, J. T. Callahan, and H. H. Shugart. 1991. Long-term research at the Virginia Coast Reserve. BioScience 41:310-318.

Hayden, B. P., M. C. Santos, G. Shao and R. C. Kochel. 1995. Geomorphological controls on coastal vegetation at the Virginia Coast Reserve. Geomorphology 13:283-300.

Haygreen, J. G. and J. L. Bowyer. 1989. Forest Products and Wood Science, an Introduction. $2^{\text {nd }}$ edition, lowa State University Press, Ames lowa.

Hmieleski, J. 1994. High marsh-forest transitions in a brackish marsh: the effects of slope. Masters Thesis, East Carolina University, Greenville, NC. 129 pp.

Intergovernmental Panel on Climate Change. 1996. Climate Change 1995 the Science of Climate Change. J. T. Houghton, L. G. Meira Filho, B. A. Callander, N. Harris, A. Kattenburg, and K. Maskell (eds.). Cambridge University Press, Great Britain. 572 pp.

Kapeluck P. R. and D. H. Van Lear. 1995. A technique for estimating below-stump biomass of mature loblolly pine plantations. Canadian Journal of Forest Research 25:355-360

Kastler, J. A. and P. L. Wieberg. 1996. Sedimentation and boundary changes of Virginia salt marshes. Estuarine, Coastal and Shelf Science 42:683-700.

Kruczynski, W. L., C. B. Subrahmanyam, and S. H. Drake. 1978. Studies of the plant community of a North Florida salt marsh. Bulletin of Marine Science 28:316-334.

Kurtz, W. A. and M. J. Apps. 1999. A 70-year retrospective analysis of carbon fluxes in the Canadian forest sector. Ecological Applications 9:526-547.

Moorhead, K. K. and M. M. Brinson. 1995. Response of wetlands to rising sea level in the lower coastal plain of North Carolina. Ecological Applications 5:261-271.

Nelson, L. E. and G. L. Switzer. 1975. Estimating dry weights of loblolly pine trees and their components in natural stands and plantations in central Mississippi. Technical Bulletin 73, Mississippi Agriculture and Forestry Exp. St., Mississippi State, MS. 14 pp.

Olson, J. S., J. A. Watts and L. J. Allison. 1983. Carbon in live vegetation of major world ecosystems. Oak Ridge (TN): Oak Ridge National Laboratory, Environmental Sciences Division Publication Number 1997. ORNL-5862.

Pehl, C. E., C. L. Tuttle, J. N. Houser and D. M. Moehring. 1984. Total biomass and nutrients of 25 -year-old Loblolly pines (Pinus taeda L.). Forest Ecology and Management 9:155-160.

Phillips, D. L. and D. L.Shure 1990. Patch-size effects on early succession in southern Appalachian forests. Ecology 71:204-212.

Potter, C. S. 1999. Terrestrial biomass and the effects of deforestation on the global carbon cycle. BioScience 49:769-778.

Reader, J. and C. Craft. 1999. Comparison of wetland structure and function on grazed and ungrazed salt marshes. Journal of Elisha Mitchell Scientific Society 115:236249.

Redfield, A. C. 1972. Development of a New England salt marsh. Ecological Monographs 42:201-237.

Reed, D. J. and D. R. Cahoon. 1992. The relationship between marsh surface topography, hydroperiod, and growth of Spartina alterniflora in a deteriorating Louisiana salt marsh. Journal of Coastal Research 8: 77-87.

Ricker, L. D. 1999. Resistance of state change by coastal ecosystems under conditions of rising sea level. Masters Thesis, East Carolina University, Greenville, NC. 209pp.

Roman, C. T. and F. C. Daiber. 1984. Aboveground and belowground primary production dynamics of two Delaware Bay Tidal marshes. Bulletin of the Torrey Botanical Club 3:34-41.

Schlesinger, W. H. 1977. Carbon balance in terrestrial detritus. Annual Review of Ecology and Systematics 8:51-81.

Schnell, R. L. 1976. Biomass estimates of eastern redcedar tree components. Technical Note, Division of Forestry, Fisheries and Wildlife Development, Tennessee Valley Authority, Norris, TN. No. B15, 15 pp.

Smith, K. K., R. E. Good, and N. F. Good. 1979. Production for above and belowground components of a New Jersey Spartina alterniflora tidal marsh. Estuarine and Coastal Marine Science 9:189-201.

Stasavich, L. E. 1998. Hydrodynamics of a coastal wetland system. Masters Thesis, East Carolina University, Greenville, NC. 68 pp.

Stevenson, J. C., M. S. Kearney and E. C. Pendleton. 1985. Sedimentation and erosion in a Chesapeake Bay brackish marsh system. Marine Geology 67:213-235.

Stevenson, J. C., L. G. Ward and M. S. Kearney. 1986. Vertical accretion in marshes with varying rates of sea level rise. pp. 241-259 In D. A. Wolfe (ed.). Estuarine Variability. Academic Press, Orlando, Florida, USA.

Stout, J. P. 1978. An analysis of annual growth and productivity of Juncus roemerianus Scheele and Spartina alterniflora Loisel in coastal Alabama. Ph. D. Thesis, University of Alabama, Tuscaloosa. 95 pp.

Stroud, L. M. 1976. Net primary production of belowground material and carbohydrate patterns of two height forms of Spartina alterniflora in two North Carolina marshes. Ph.D. North Carolina State University, Raleigh, NC.

Thomas, C. R. 1998. The use of network analysis to compare the nitrogen cycles of three salt marsh zones experiencing relative to sea-level rise. Masters Thesis, East Carolina University, Greenville, NC

Tolley, P. M. and R. R. Christian. 1999. Effects of increased inundation and wrack deposition on a high salt marsh plant community. Estuaries 22:944-954.

Tolley P. M. 1996. Effects of increase inundation and wrack deposition on a high salt marsh plant community. Masters Thesis, East Carolina University, Greenville, NC.

Van Lear, D. H. and P. R. Kapeluck. 1995. Above and below-stump biomass and nutrient content of a mature loblolly pine plantation. Canadian Journal of Forest Research 25:361-367.

Van Lear, D. H., J. B. Waide and M. J. Teuke. 1984. Biomass and nutrient content of a 41-year-old loblolly pine (Pinus taeda L.) plantation on a poor site in South Carolina. Forest Science 30:395-404.

USDA Soil Conservation Service. 1989. Soil survey of Northampton County, Virginia 94pp.

Webb, E. C., I. A. Mendelssohn, and B. J. Wilsey. 1995. Causes for vegetation dieback in a Louisiana salt marsh: a bioassay approach. Aquatic Botany 51:281-289.

White, D. A., T. E. Weiss, J. M. J. M Trapani and L. B. Thien. 1978. Productivity and decomposition of the dominant salt marsh plants in Louisiana. Ecology 59:751-759

Whittaker, R. H., F. H, Bormann, G. E. Likens, and T. G. Siccama. 1974. The Hubbard Brook ecosystem study: forest biomass and production. Ecological Monographs 44:233-252

Whittaker, R. H. and G. M. Woodwell. 1969. Structure, production and diversity of the oak-pine forest at Brookhaven, New York. Journal of Ecology 57:155-174.

APPENDIX A. OVEN-DRY MASS ( $\mathrm{KG} / \mathrm{M}^{2}$ ) OF ABOVEGROUND VEGETATION COMPONENTS SUMMED FOR ALL SITES.

| Zone | Site | Trees | Shrubs | Woody veg | Snags | Large <br> wood | Small wood | Cones | Herb.Veg. | Herb. Litter | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Forest | Fa18 | 18.45 | 0.00 |  | 0.002 | 0.282 | 0.454 | 0.153 | 0.093 | 2.46 | 21.90 |
| Forest | Fa30 | 30.73 | 0.00 |  | 0.000 | 0.025 | 0.115 | 0.081 | 0.006 | 1.83 | 32.79 |
| Forest | Fb29 | 19.68 | 0.21 |  | 0.097 | 0.000 | 0.428 | 0.062 | 0.032 | 0.80 | 21.30 |
| Forest | Fb47 | 23.66 | 0.00 |  | 1.322 | 0.000 | 0.738 | 0.047 | 0.009 | 1.86 | 27.64 |
| Forest | Fb65 | 19.67 | 0.19 |  | 0.339 | 0.000 | 0.385 | 0.132 | 0.048 | 1.74 | 22.50 |
| Transition | Ta1 | 0.19 | 0.04 |  | 0.193 | 0.000 | 0.002 | 0.000 | 1.089 | 0.29 | 1.80 |
| Transition | Ta4 | 0.31 | 0.03 |  | 0.013 | 0.102 | 0.000 | 0.000 | 1.207 | 0.45 | 2.10 |
| Transition | Ta13 | 0.22 | 0.20 | 0.06 | 0.053 | 0.000 | 0.022 | 0.002 | 0.898 | 0.35 | 1.79 |
| Transition | Tb29 | 0.00 | 0.22 |  | 0.000 | 0.000 | 0.000 | 0.000 | 1.558 | 0.15 | 1.92 |
| Transition | Tc2 | 0.28 | 0.05 |  | 0.018 | 0.000 | 0.014 | 0.000 | 0.975 | 0.33 | 1.66 |
| Transition | Tc8 | 5.26 | 0.97 | 0.01 | 0.185 | 0.050 | 0.264 | 0.002 | 0.609 | 0.69 | 8.04 |
| Transition | Tc18 | 2.65 | 0.80 |  | 0.168 | 0.010 | 0.104 | 0.000 | 1.135 | 0.22 | 5.08 |
| Transition | Tc31 | 0.63 | 0.04 |  | 0.410 | 0.024 | 0.156 | 0.000 | 0.541 | 0.25 | 2.04 |
| Transition | Td2 | 3.68 | 0.54 |  | 0.315 | 0.214 | 0.323 | 0.012 | 0.804 | 0.51 | 6.40 |
| Transition | Td9 | 3.05 | 0.54 |  | 1.048 | 0.084 | 0.041 | 0.000 | 1.155 | 0.24 | 6.15 |
| Transition | Td15 | 4.97 | 0.05 |  | 0.064 | 0.190 | 0.013 | 0.000 | 0.975 | 0.16 | 6.43 |
| Transition | Td33 | 0.76 | 0.00 |  | 0.054 | 0.000 | 0.020 | 0.000 | 1.019 | 0.29 | 2.14 |
| Transition | Td39 | 4.60 | 0.13 |  | 0.401 | 0.189 | 0.213 | 0.001 | 0.797 | 0.57 | 6.90 |
| Transition | Td50 | 0.45 | 0.13 |  | 0.042 | 0.000 | 0.041 | 0.000 | 1.611 | 1.16 | 3.43 |
| High Marsh | Ha5 |  |  |  |  |  | 0.018 |  | 1.260 | 0.27 | 1.55 |
| High Marsh | Ha14 |  |  |  |  |  |  |  | 1.235 | 0.30 | 1.53 |
| High Marsh | Ha19 |  |  |  |  |  |  |  | 0.714 | 0.65 | 1.36 |
| High Marsh | Ha21 |  |  |  |  |  |  |  | 0.486 | 0.34 | 0.83 |
| High Marsh | Ha36 |  |  |  |  |  | 0.001 |  | 1.032 | 0.41 | 1.44 |
| High Marsh | Ha41 |  |  |  |  |  |  |  | 0.923 | 0.22 | 1.14 |
| High Marsh | Ha52 |  |  |  |  |  |  |  | 0.448 | 0.03 | 0.48 |
| High Marsh | Hb 1 |  |  |  |  |  |  |  | 0.846 | 0.57 | 1.41 |
| High Marsh | Hb4 |  |  |  |  |  |  |  | 1.056 | 0.37 | 1.43 |


| High Marsh | Hc 1 |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Zone | Site | Herb.Veg. | Herb. Litter | Total |
| High Marsh | Hc30 | 2.556 | 0.53 | 3.09 |
| High Marsh | Hc37 | 0.051 | 0.03 | 0.00 |
| High Marsh | Hc 41 | 0.458 | 0.50 | 0.00 |
| High Marsh | Hc 42 | 2.743 | 0.51 | 0.00 |
| High Marsh | Hc 48 | 1.540 | 0.29 | 0.00 |
| High Marsh | Hc 72 | 1.052 | 0.24 | 0.00 |
| High Marsh | Hc 75 | 1.828 | 0.38 | 0.00 |
| High Marsh | Hc84 | 0.693 | 0.47 | 0.00 |
| High Marsh | Hc 92 | 0.656 | 0.32 | 0.00 |
| High Marsh | Ma1 | 2.712 | 0.30 | 0.00 |
| High Marsh | Ma15 | 1.530 | 0.30 | 0.00 |
| High Marsh | Mc2 | 0.468 | 0.17 | 0.00 |
| High Marsh | Tb4 | 0.953 | 0.14 | 0.00 |
| High Marsh | Tb14 | 0.814 | 0.34 | 0.00 |
| High Marsh | Tb23 | 1.300 | 0.27 | 0.00 |
| LMSS | LSa8 | 0.365 | 0.10 | 0.00 |
| LMSS | LSa12 | 0.730 | 0.12 | 0.00 |
| LMSS | LSb2 | 0.567 | 0.08 | 0.00 |
| LMSS | LSb16 | 0.778 | 0.02 | 0.00 |
| LMSS | LSb17 | 0.873 | 0.41 | 0.00 |
| LMSS | LSb46 | 0.604 | 0.40 | 0.00 |
| LMSS | LSb53 | 0.585 | 0.12 | 0.00 |
| LMSS | LSb55 | 0.702 | 0.19 | 0.00 |
| LMSS | Mb4 | 0.420 | 0.32 | 0.00 |
| LMTS | LTS6 | 0.821 | 0.00 | 0.00 |
| LMTS | LTS9 | 1.057 | 0.41 | 0.00 |
| LMTS | LT6 | 1.664 | 0.21 | 0.00 |
| LMTS | LT12 | 1.484 | 0.00 | 0.00 |
| LMTS | LT25 | 1.119 | 0.00 | 0.00 |
| LMTS | LT29 | 0.583 | 0.09 | 0.00 |
| LMTS | LT33 | 1.231 | 0.15 | 0.00 |

APPENDIX B. SUMMATION OF BIOMASS, ASH-FREE DRY WEIGHT, AND CARBON MASS OF TREES PER FOREST AND TRANSITION SITE.

| Zone | Site | Species | Dry wt. kg | Dry wt. $\mathrm{kg} / \mathrm{m}^{2}$ | \% <br> Organic Matter | AFDW $\mathrm{kg} / \mathrm{m}^{2}$ | Organic carbon $\mathrm{kg} / \mathrm{m}^{2}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Forest | Fa18 | Pinus taeda | 509.13 | 1.621 | 98 | 1.589 | 0.795 |
| Forest | Fa18 | P. taeda | 843.76 | 2.687 | 98 | 2.633 | 1.317 |
| Forest | Fa18 | P. taeda | 1293.87 | 4.121 | 98 | 4.038 | 2.019 |
| Forest | Fa18 | P. taeda | 940.11 | 2.994 | 98 | 2.934 | 1.467 |
| Forest | Fa18 | P. taeda | 1141.59 | 3.636 | 98 | 3.563 | 1.781 |
| Forest | Fa18 | P. taeda | 858.57 | 2.734 | 98 | 2.680 | 1.340 |
| Forest | Fa18 | Prunus virginiana | 2.59 | 0.008 | 98 | 0.008 | 0.004 |
| Forest | Fa18 | P. virginiana | 9.01 | 0.029 | 98 | 0.028 | 0.014 |
| Forest | Fa18 | P. virginiana | 3.56 | 0.011 | 98 | 0.011 | 0.006 |
| Forest | Fa18 | P. virginiana | 5.98 | 0.019 | 98 | 0.019 | 0.009 |
| Forest | Fa18 | P. virginiana | 2.56 | 0.008 | 98 | 0.008 | 0.004 |
| Forest | Fa18 | P. virginiana | 1.85 | 0.006 | 98 | 0.006 | 0.003 |
| Forest | Fa18 | P. virginiana | 1.71 | 0.005 | 98 | 0.005 | 0.003 |
| Forest | Fa18 | $P$. virginiana | 4.71 | 0.015 | 98 | 0.015 | 0.007 |
| Forest | Fa18 | P. virginiana | 1.28 | 0.004 | 98 | 0.004 | 0.002 |
| Forest | Fa18 | P. virginiana | 10.76 | 0.034 | 98 | 0.034 | 0.017 |
| Forest | Fa18 | P. virginiana | 20.72 | 0.066 | 98 | 0.065 | 0.032 |
| Forest | Fa18 | P. virginiana | 1.28 | 0.004 | 98 | 0.004 | 0.002 |
| Forest | Fa18 | P. virginiana | 2.16 | 0.007 | 98 | 0.007 | 0.003 |
| Forest | Fa18 | P. virginiana | 9.21 | 0.029 | 98 | 0.029 | 0.014 |
| Forest | Fa18 | $P$. virginiana | 3.56 | 0.011 | 98 | 0.011 | 0.006 |
| Forest | Fa18 | llex opaca | 11.28 | 0.036 | 98 | 0.035 | 0.018 |
| Forest | Fa18 | llex opaca | 10.25 | 0.033 | 98 | 0.032 | 0.016 |
| Forest | Fa18 | Liriodendron styraciflua | 58.76 | 0.187 | 98 | 0.183 | 0.092 |
| Forest | Fa18 | L. styraciflua | 39.38 | 0.125 | 98 | 0.123 | 0.061 |
| Forest | Fa18 | Quercus phellos | 3.25 | 0.010 | 98 | 0.010 | 0.005 |
| Forest | Fa18 | Aralia spinosa | 1.01 | 0.003 | 98 | 0.003 | 0.002 |
| Forest | Fa18 | Aralia spinosa | 1.46 | 0.005 | 98 | 0.005 | 0.002 |
| Forest | Fa18 | Celtis spp. | 1.01 | 0.003 | 98 | 0.003 | 0.002 |
| Total |  |  | 5794.38 | 18.45 |  | 18.08 | 9.04 |
| Forest | Fa30 | P. taeda | 896.30 | 2.854 | 98 | 2.797 | 1.399 |
| Forest | Fa30 | P. taeda | 1604.41 | 5.110 | 98 | 5.007 | 2.504 |
| Forest | Fa30 | P. taeda | 991.28 | 3.157 | 98 | 3.094 | 1.547 |
| Forest | Fa30 | P. taeda | 1020.31 | 3.249 | 98 | 3.184 | 1.592 |
| Forest | Fa30 | P. taeda | 289.47 | 0.922 | 98 | 0.903 | 0.452 |
| Forest | Fa30 | P. taeda | 1065.37 | 3.393 | 98 | 3.325 | 1.663 |
| Forest | Fa30 | P. taeda | 685.39 | 2.183 | 98 | 2.139 | 1.070 |
| Forest | Fa30 | P. taeda | 1021.10 | 3.252 | 98 | 3.187 | 1.593 |
| Forest | Fa30 | P. taeda | 1043.30 | 3.323 | 98 | 3.256 | 1.628 |
| Forest | Fa30 | $Q$. phellos | 1.01 | 0.003 | 98 | 0.003 | 0.002 |


| Zone | Site | Species | Dry wt. kg | Dry wt. $\mathrm{kg} / \mathrm{m}^{2}$ |  | AFDW $\mathrm{kg} / \mathrm{m}^{2}$ | Organic carbon $\mathrm{kg} / \mathrm{m}^{2}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Forest | Fa30 | Q. phellos | 4.03 | 0.013 | 98 | 0.013 | 0.006 |
|  |  | Juniperus |  |  |  |  |  |
| Forest | Fa30 | virginiana | 140.62 | 0.448 | 98 | 0.439 | 0.219 |
| Forest | Fa30 | J. virginiana | 104.78 | 0.334 | 98 | 0.327 | 0.164 |
| Forest | Fa30 | J. virginiana | 287.13 | 0.914 | 98 | 0.896 | 0.448 |
| Forest | Fa30 | J. virginiana | 494.42 | 1.575 | 98 | 1.543 | 0.772 |
| Total |  |  | 9648.92 | 30.73 |  | 30.11 | 15.06 |
| Forest | Fb29 | P. taeda | 333.10 | 1.061 | 98 | 1.040 | 0.520 |
| Forest | Fb29 | P. taeda | 727.68 | 2.317 | 98 | 2.271 | 1.136 |
| Forest | Fb29 | P. taeda | 216.58 | 0.690 | 98 | 0.676 | 0.338 |
| Forest | Fb29 | P. taeda | 369.80 | 1.178 | 98 | 1.154 | 0.577 |
| Forest | Fb29 | P. taeda | 509.28 | 1.622 | 98 | 1.589 | 0.795 |
| Forest | Fb29 | P. taeda | 705.42 | 2.247 | 98 | 2.202 | 1.101 |
| Forest | Fb29 | Nyssa sylvatica | 773.12 | 2.462 | 98 | 2.413 | 1.206 |
| Forest | Fb29 | N. sylvatica | 420.63 | 1.340 | 98 | 1.313 | 0.656 |
| Forest | Fb29 | N. sylvatica | 1108.25 | 3.529 | 98 | 3.459 | 1.729 |
| Forest | Fb29 | N. sylvatica | 2.13 | 0.007 | 98 | 0.007 | 0.003 |
| Forest | Fb29 | N. sylvatica | 4.09 | 0.013 | 98 | 0.013 | 0.006 |
| Forest | Fb29 | N. sylvatica | 4.16 | 0.013 | 98 | 0.013 | 0.006 |
| Forest | Fb29 | N. sylvatica | 20.24 | 0.064 | 98 | 0.063 | 0.032 |
| Forest | Fb29 | N. sylvatica | 25.34 | 0.081 | 98 | 0.079 | 0.040 |
| Forest | Fb29 | N. sylvatica | 5.83 | 0.019 | 98 | 0.018 | 0.009 |
| Forest | Fb29 | N. sylvatica | 7.76 | 0.025 | 98 | 0.024 | 0.012 |
| Forest | Fb29 | N. sylvatica | 43.32 | 0.138 | 98 | 0.135 | 0.068 |
| Forest | Fb29 | N. sylvatica | 100.73 | 0.321 | 98 | 0.314 | 0.157 |
| Forest | Fb29 | N. sylvatica | 104.70 | 0.333 | 98 | 0.327 | 0.163 |
| Forest | Fb29 | N. sylvatica | 58.38 | 0.186 | 98 | 0.182 | 0.091 |
| Forest | Fb29 | N. sylvatica | 49.81 | 0.159 | 98 | 0.155 | 0.078 |
| Forest | Fb29 | N. sylvatica | 10.33 | 0.033 | 98 | 0.032 | 0.016 |
| Forest | Fb29 | N. sylvatica | 5.51 | 0.018 | 98 | 0.017 | 0.009 |
| Forest | Fb29 | N. sylvatica | 0.61 | 0.002 | 98 | 0.002 | 0.001 |
| Forest | Fb29 | N. sylvatica | 0.80 | 0.003 | 98 | 0.002 | 0.001 |
| Forest | Fb29 | N. sylvatica | 0.62 | 0.002 | 98 | 0.002 | 0.001 |
| Forest | Fb29 | N. sylvatica | 67.70 | 0.216 | 98 | 0.211 | 0.106 |
| Forest | Fb29 | l. opaca | 4.87 | 0.016 | 98 | 0.015 | 0.008 |
| Forest | Fb29 | l. opaca | 7.95 | 0.025 | 98 | 0.025 | 0.012 |
| Forest | Fb29 | l. opaca | 2.63 | 0.008 | 98 | 0.008 | 0.004 |
| Forest | Fb29 | l. opaca | 2.63 | 0.008 | 98 | 0.008 | 0.004 |
| Forest | Fb29 | l. opaca | 0.75 | 0.002 | 98 | 0.002 | 0.001 |
| Forest | Fb29 | l. opaca | 3.37 | 0.011 | 98 | 0.011 | 0.005 |
| Forest | Fb29 | l. opaca | 10.32 | 0.033 | 98 | 0.032 | 0.016 |
| Forest | Fb29 | P. virginiana | 19.77 | 0.063 | 98 | 0.062 | 0.031 |
| Forest | Fb29 | P. virginiana | 34.30 | 0.109 | 98 | 0.107 | 0.054 |


| Zone | Site | Species | Dry wt. kg | Dry wt. $\mathrm{kg} / \mathrm{m}^{2}$ |  | AFDW $\mathrm{kg} / \mathrm{m}^{2}$ | Organic carbon $\mathrm{kg} / \mathrm{m}^{2}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Forest | Fb29 | P. virginiana | 28.95 | 0.092 | 98 | 0.090 | 0.045 |
| Forest | Fb29 | Acer rubrum | 171.70 | 0.547 | 98 | 0.536 | 0.268 |
| Forest | Fb29 | J. virginiana | 9.10 | 0.029 | 98 | 0.028 | 0.014 |
| Forest | Fb29 | J. virginiana | 15.16 | 0.048 | 98 | 0.047 | 0.024 |
| Forest | Fb29 | J. virginiana | 27.29 | 0.087 | 98 | 0.085 | 0.043 |
| Forest | Fb29 | J. virginiana | 15.16 | 0.048 | 98 | 0.047 | 0.024 |
| Forest | Fb29 | J. virginiana | 74.38 | 0.237 | 98 | 0.232 | 0.116 |
| Forest | Fb29 | J. virginiana | 74.38 | 0.237 | 98 | 0.232 | 0.116 |
| Total |  |  | 6178.6 | 19.68 |  | 19.28 | 9.64 |
| Forest | Fb47 | P. taeda | 561.86 | 1.789 | 98 | 1.754 | 0.877 |
| Forest | Fb47 | P. taeda | 223.76 | 0.713 | 98 | 0.698 | 0.349 |
| Forest | Fb47 | P. taeda | 344.84 | 1.098 | 98 | 1.076 | 0.538 |
| Forest | Fb47 | P. taeda | 607.70 | 1.935 | 98 | 1.897 | 0.948 |
| Forest | Fb47 | P. taeda | 373.46 | 1.189 | 98 | 1.166 | 0.583 |
| Forest | Fb47 | P. taeda | 801.98 | 2.554 | 98 | 2.503 | 1.251 |
| Forest | Fb47 | P. taeda | 404.98 | 1.290 | 98 | 1.264 | 0.632 |
| Forest | Fb47 | P. taeda | 1223.66 | 3.897 | 98 | 3.819 | 1.910 |
| Forest | Fb47 | P. taeda | 187.19 | 0.596 | 98 | 0.584 | 0.292 |
| Forest | Fb47 | P. taeda | 396.29 | 1.262 | 98 | 1.237 | 0.618 |
| Forest | Fb47 | P. taeda | 477.00 | 1.519 | 98 | 1.489 | 0.744 |
| Forest | Fb47 | P. taeda | 641.58 | 2.043 | 98 | 2.002 | 1.001 |
| Forest | Fb47 | P. taeda | 84.39 | 0.269 | 98 | 0.263 | 0.132 |
| Forest | Fb47 | L. styraciflua | 47.06 | 0.150 | 98 | 0.147 | 0.073 |
| Forest | Fb47 | L. styraciflua | 64.89 | 0.207 | 98 | 0.203 | 0.101 |
| Forest | Fb47 | L. styraciflua | 151.52 | 0.483 | 98 | 0.473 | 0.236 |
| Forest | Fb47 | L. styraciflua | 177.01 | 0.564 | 98 | 0.552 | 0.276 |
| Forest | Fb47 | L. styraciflua | 36.90 | 0.118 | 98 | 0.115 | 0.058 |
| Forest | Fb47 | L. styraciflua | 18.49 | 0.059 | 98 | 0.058 | 0.029 |
| Forest | Fb47 | L. styraciflua | 16.00 | 0.051 | 98 | 0.050 | 0.025 |
| Forest | Fb47 | L. styraciflua | 15.81 | 0.050 | 98 | 0.049 | 0.025 |
| Forest | Fb47 | L. styraciflua | 11.94 | 0.038 | 98 | 0.037 | 0.019 |
| Forest | Fb47 | L. styraciflua | 13.47 | 0.043 | 98 | 0.042 | 0.021 |
| Forest | Fb47 | N. sylvatica | 189.43 | 0.603 | 98 | 0.591 | 0.296 |
| Forest | Fb47 | N. sylvatica | 35.08 | 0.112 | 98 | 0.109 | 0.055 |
| Forest | Fb47 | N. sylvatica | 17.28 | 0.055 | 98 | 0.054 | 0.027 |
| Forest | Fb47 | N. sylvatica | 9.25 | 0.029 | 98 | 0.029 | 0.014 |
| Forest | Fb47 | N. sylvatica | 46.42 | 0.148 | 98 | 0.145 | 0.072 |
| Forest | Fb47 | N. sylvatica | 84.19 | 0.268 | 98 | 0.263 | 0.131 |
| Forest | Fb47 | N. sylvatica | 26.66 | 0.085 | 98 | 0.083 | 0.042 |
| Forest | Fb47 | I. opaca | 1.72 | 0.005 | 98 | 0.005 | 0.003 |
| Forest | Fb47 | l. opaca | 4.36 | 0.014 | 98 | 0.014 | 0.007 |


| Zone | Site | Species | Dry wt. kg | Dry wt. $\mathrm{kg} / \mathrm{m}^{2}$ |  | AFDW $\mathrm{kg} / \mathrm{m}^{2}$ | Organic carbon $\mathrm{kg} / \mathrm{m}^{2}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Forest | Fb47 | I. opaca | 34.31 | 0.109 | 98 | 0.107 | 0.054 |
| Forest | Fb47 | l. opaca | 1.00 | 0.003 | 98 | 0.003 | 0.002 |
| Forest | Fb47 | l. opaca | 1.68 | 0.005 | 98 | 0.005 | 0.003 |
| Forest | Fb47 | l. opaca | 7.29 | 0.023 | 98 | 0.023 | 0.011 |
| Forest | Fb47 | l. opaca | 5.24 | 0.017 | 98 | 0.016 | 0.008 |
| Forest | Fb47 | l. opaca | 84.19 | 0.268 | 98 | 0.263 | 0.131 |
| Total |  |  | 7429.90 | 23.66 |  | 23.19 | 11.59 |
| Forest | Fb65 | P. taeda | 436.44 | 1.390 | 98 | 1.362 | 0.681 |
| Forest | Fb65 | P. taeda | 41.78 | 0.133 | 98 | 0.130 | 0.065 |
| Forest | Fb65 | P. taeda | 633.16 | 2.016 | 98 | 1.976 | 0.988 |
| Forest | Fb65 | P. taeda | 8.57 | 0.027 | 98 | 0.027 | 0.013 |
| Forest | Fb65 | P. taeda | 67.88 | 0.216 | 98 | 0.212 | 0.106 |
| Forest | Fb65 | P. taeda | 529.93 | 1.688 | 98 | 1.654 | 0.827 |
| Forest | Fb65 | P. taeda | 593.46 | 1.890 | 98 | 1.852 | 0.926 |
| Forest | Fb65 | P. taeda | 577.37 | 1.839 | 98 | 1.802 | 0.901 |
| Forest | Fb65 | P. taeda | 560.49 | 1.785 | 98 | 1.749 | 0.875 |
| Forest | Fb65 | P. taeda | 613.63 | 1.954 | 98 | 1.915 | 0.958 |
| Forest | Fb65 | P. taeda | 625.02 | 1.990 | 98 | 1.951 | 0.975 |
| Forest | Fb65 | P. taeda | 885.88 | 2.821 | 98 | 2.765 | 1.382 |
| Forest | Fb65 | P. taeda | 4.27 | 0.014 | 98 | 0.013 | 0.007 |
| Forest | Fb65 | P. taeda | 2.55 | 0.008 | 98 | 0.008 | 0.004 |
| Forest | Fb65 | J. virginiana | 74.39 | 0.237 | 98 | 0.232 | 0.116 |
| Forest | Fb65 | J. virginiana | 49.44 | 0.157 | 98 | 0.154 | 0.077 |
| Forest | Fb65 | J. virginiana | 24.25 | 0.077 | 98 | 0.076 | 0.038 |
| Forest | Fb65 | J. virginiana | 74.39 | 0.237 | 98 | 0.232 | 0.116 |
| Forest | Fb65 | J. virginiana | 15.16 | 0.048 | 98 | 0.047 | 0.024 |
| Forest | Fb65 | J. virginiana | 9.10 | 0.029 | 98 | 0.028 | 0.014 |
| Forest | Fb65 | J. virginiana | 49.44 | 0.157 | 98 | 0.154 | 0.077 |
| Forest | Fb65 | J. virginiana | 74.39 | 0.237 | 98 | 0.232 | 0.116 |
| Forest | Fb65 | J. virginiana | 24.25 | 0.077 | 98 | 0.076 | 0.038 |
| Forest | Fb65 | J. virginiana | 27.29 | 0.087 | 98 | 0.085 | 0.043 |
| Forest | Fb65 | J. virginiana | 27.29 | 0.087 | 98 | 0.085 | 0.043 |
| Forest | Fb65 | J. virginiana | 104.78 | 0.334 | 98 | 0.327 | 0.164 |
| Forest | Fb65 | l. opaca | 41.76 | 0.133 | 98 | 0.130 | 0.065 |
| Total |  |  | 6176.38 | 19.67 |  | 19.28 | 9.64 |
| Transition | Ta1 | J. virginiana | 18.19 | 0.058 | 98 | 0.057 | 0.028 |
| Transition | Ta1 | J. virginiana | 7.70 | 0.025 | 98 | 0.024 | 0.012 |
| Transition | Ta1 | J. virginiana | 15.16 | 0.048 | 98 | 0.047 | 0.024 |
| Transition | Ta1 | J. virginiana | 9.10 | 0.029 | 98 | 0.028 | 0.014 |
| Transition | Ta1 | J. virginiana | 9.10 | 0.029 | 98 | 0.028 | 0.014 |
| Total |  |  | 59.24 | 0.19 |  | 0.18 | 0.09 |


| Zone | Site | Species | Dry wt. kg | Dry wt. $\mathrm{kg} / \mathrm{m}^{2}$ | \% <br> Organic Matter | AFDW $\mathrm{kg} / \mathrm{m}^{2}$ | Organic carbon $\mathrm{kg} / \mathrm{m}^{2}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Transition | Ta4 | J. virginiana | 15.16 | 0.048 | 98 | 0.047 | 0.024 |
| Transition | Ta4 | J. virginiana | 9.10 | 0.029 | 98 | 0.028 | 0.014 |
| Transition | Ta4 | J. virginiana | 9.10 | 0.029 | 98 | 0.028 | 0.014 |
| Transition | Ta4 | J. virginiana | 15.16 | 0.048 | 98 | 0.047 | 0.024 |
| Transition | Ta4 | J. virginiana | 9.10 | 0.029 | 98 | 0.028 | 0.014 |
| Transition | Ta4 | J. virginiana | 7.70 | 0.025 | 98 | 0.024 | 0.012 |
| Transition | Ta4 | J. virginiana | 12.13 | 0.039 | 98 | 0.038 | 0.019 |
| Transition | Ta4 | J. virginiana | 18.19 | 0.058 | 98 | 0.057 | 0.028 |
| Transition | Ta4 | J. virginiana | 0.30 | 0.001 | 98 | 0.001 | 0.000 |
| Total |  |  | 95.93 | 0.31 |  | 0.30 | 0.150 |
| Transition | Ta13 | J. virginiana | 7.70 | 0.025 | 98 | 0.024 | 0.012 |
| Transition | Ta13 | J. virginiana | 9.10 | 0.029 | 98 | 0.028 | 0.014 |
| Transition | Ta13 | J. virginiana | 12.13 | 0.039 | 98 | 0.038 | 0.019 |
| Transition | Ta13 | J. virginiana | 7.70 | 0.025 | 98 | 0.024 | 0.012 |
| Transition | Ta13 | J. virginiana | 24.25 | 0.077 | 98 | 0.076 | 0.038 |
| Transition | Ta13 | J. virginiana | 7.70 | 0.025 | 98 | 0.024 | 0.012 |
| Total |  |  | 68.58 | 0.22 |  | 0.21 | 0.107 |
| Transition | Tb29 | N/A | 0.00 | 0.000 |  | 0.000 | 0.000 |
| Transition | Tc2 | J. virginiana | 2.66 | 0.008 | 98 | 0.008 | 0.004 |
| Transition | Tc2 | P. taeda | 4.19 | 0.013 | 98 | 0.013 | 0.007 |
| Transition | Tc2 | P. taeda | 0.66 | 0.002 | 98 | 0.002 | 0.001 |
| Transition | Tc2 | P. taeda | 1.53 | 0.005 | 98 | 0.005 | 0.002 |
| Transition | Tc2 | P. taeda | 1.53 | 0.005 | 98 | 0.005 | 0.002 |
| Transition | Tc2 | P. taeda | 1.73 | 0.006 | 98 | 0.005 | 0.003 |
| Transition | Tc2 | P. taeda | 1.53 | 0.005 | 98 | 0.005 | 0.002 |
| Transition | Tc2 | P. taeda | 2.35 | 0.007 | 98 | 0.007 | 0.004 |
| Transition | Tc2 | P. taeda | 2.35 | 0.007 | 98 | 0.007 | 0.004 |
| Transition | Tc2 | P. taeda | 2.35 | 0.007 | 98 | 0.007 | 0.004 |
| Transition | Tc2 | J. virginiana | 27.29 | 0.087 | 98 | 0.085 | 0.043 |
| Transition | Tc2 | J. virginiana | 18.19 | 0.058 | 98 | 0.057 | 0.028 |
| Transition | Tc2 | J. virginiana | 21.22 | 0.068 | 98 | 0.066 | 0.033 |
| Total |  |  | 87.59 | 0.28 |  | 0.27 | 0.14 |
| Transition | Tc8 | P. taeda | 0.20 | 0.001 | 98 | 0.001 | 0.000 |
| Transition | Tc8 | P. taeda | 3.55 | 0.011 | 98 | 0.011 | 0.006 |
| Transition | Tc8 | P. taeda | 120.98 | 0.385 | 98 | 0.378 | 0.189 |
| Transition | Tc8 | P. taeda | 49.86 | 0.159 | 98 | 0.156 | 0.078 |
| Transition | Tc8 | P. taeda | 9.11 | 0.029 | 98 | 0.028 | 0.014 |
| Transition | Tc8 | P. taeda | 13.41 | 0.043 | 98 | 0.042 | 0.021 |


| Zone | Site | Species | Dry wt. kg | Dry wt. $\mathrm{kg} / \mathrm{m}^{2}$ | \% <br> Organic Matter | AFDW $\mathrm{kg} / \mathrm{m}^{2}$ | Organic carbon $\mathrm{kg} / \mathrm{m}^{2}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Transition | Tc8 | P. taeda | 87.28 | 0.278 | 98 | 0.272 | 0.136 |
| Transition | Tc8 | P. taeda | 115.62 | 0.368 | 98 | 0.361 | 0.180 |
| Transition | Tc8 | P. taeda | 0.40 | 0.001 | 98 | 0.001 | 0.001 |
| Transition | Tc8 | P. taeda | 0.66 | 0.002 | 98 | 0.002 | 0.001 |
| Transition | Tc8 | P. taeda | 5.03 | 0.016 | 98 | 0.016 | 0.008 |
| Transition | Tc8 | P. taeda | 8.74 | 0.028 | 98 | 0.027 | 0.014 |
| Transition | Tc8 | P. taeda | 113.92 | 0.363 | 98 | 0.356 | 0.178 |
| Transition | Tc8 | P. taeda | 10.96 | 0.035 | 98 | 0.034 | 0.017 |
| Transition | Tc8 | P. taeda | 27.26 | 0.087 | 98 | 0.085 | 0.043 |
| Transition | Tc8 | P. taeda | 8.74 | 0.028 | 98 | 0.027 | 0.014 |
| Transition | Tc8 | P. taeda | 6.76 | 0.022 | 98 | 0.021 | 0.011 |
| Transition | Tc8 | P. taeda | 6.76 | 0.022 | 98 | 0.021 | 0.011 |
| Transition | Tc8 | P. taeda | 8.74 | 0.028 | 98 | 0.027 | 0.014 |
| Transition | Tc8 | P. taeda | 13.39 | 0.043 | 98 | 0.042 | 0.021 |
| Transition | Tc8 | P. taeda | 8.74 | 0.028 | 98 | 0.027 | 0.014 |
| Transition | Tc8 | P. taeda | 156.30 | 0.498 | 98 | 0.488 | 0.244 |
| Transition | Tc8 | P. taeda | 29.89 | 0.095 | 98 | 0.093 | 0.047 |
| Transition | Tc8 | P. taeda | 4.19 | 0.013 | 98 | 0.013 | 0.007 |
| Transition | Tc8 | P. taeda | 15.67 | 0.050 | 98 | 0.049 | 0.024 |
| Transition | Tc8 | P. taeda | 8.74 | 0.028 | 98 | 0.027 | 0.014 |
| Transition | Tc8 | P. taeda | 16.10 | 0.051 | 98 | 0.050 | 0.025 |
| Transition | Tc8 | P. taeda | 61.11 | 0.195 | 98 | 0.191 | 0.095 |
| Transition | Tc8 | P. taeda | 5.07 | 0.016 | 98 | 0.016 | 0.008 |
| Transition | Tc8 | P. taeda | 48.24 | 0.154 | 98 | 0.151 | 0.075 |
| Transition | Tc8 | P. taeda | 64.71 | 0.206 | 98 | 0.202 | 0.101 |
| Transition | Tc8 | P. taeda | 61.11 | 0.195 | 98 | 0.191 | 0.095 |
| Transition | Tc8 | P. taeda | 12.81 | 0.041 | 98 | 0.040 | 0.020 |
| Transition | Tc8 | P. taeda | 12.81 | 0.041 | 98 | 0.040 | 0.020 |
| Transition | Tc8 | P. taeda | 0.66 | 0.002 | 98 | 0.002 | 0.001 |
| Transition | Tc8 | P. taeda | 1.15 | 0.004 | 98 | 0.004 | 0.002 |
| Transition | Tc8 | P. taeda | 1.15 | 0.004 | 98 | 0.004 | 0.002 |
| Transition | Tc8 | P. taeda | 3.34 | 0.011 | 98 | 0.010 | 0.005 |
| Transition | Tc8 | P. taeda | 10.68 | 0.034 | 98 | 0.033 | 0.017 |
| Transition | Tc8 | J. virginiana | 9.10 | 0.029 | 98 | 0.028 | 0.014 |
| Transition | Tc8 | J. virginiana | 15.16 | 0.048 | 98 | 0.047 | 0.024 |
| Transition | Tc8 | J. virginiana | 15.16 | 0.048 | 98 | 0.047 | 0.024 |
| Transition | Tc8 | J. virginiana | 36.38 | 0.116 | 98 | 0.114 | 0.057 |
| Transition | Tc8 | J. virginiana | 30.32 | 0.097 | 98 | 0.095 | 0.047 |
| Transition | Tc8 | J. virginiana | 46.72 | 0.149 | 98 | 0.146 | 0.073 |
| Transition | Tc8 | J. virginiana | 9.10 | 0.029 | 98 | 0.028 | 0.014 |
| Transition | Tc8 | J. virginiana | 30.32 | 0.097 | 98 | 0.095 | 0.047 |
| Transition | Tc8 | J. virginiana | 21.22 | 0.068 | 98 | 0.066 | 0.033 |


| Zone | Site | Species | Dry wt. kg | Dry wt. $\mathrm{kg} / \mathrm{m}^{2}$ | \% <br> Organic Matter | AFDW $\mathrm{kg} / \mathrm{m}^{2}$ | Organic carbon $\mathrm{kg} / \mathrm{m}^{2}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Transition | Tc8 | J. virginiana | 27.29 | 0.087 | 98 | 0.085 | 0.043 |
| Transition | Tc8 | J. virginiana | 27.29 | 0.087 | 98 | 0.085 | 0.043 |
| Transition | Tc8 | J. virginiana | 18.19 | 0.058 | 98 | 0.057 | 0.028 |
| Transition | Tc8 | J. virginiana | 27.29 | 0.087 | 98 | 0.085 | 0.043 |
| Transition | Tc8 | J. virginiana | 51.88 | 0.165 | 98 | 0.162 | 0.081 |
| Transition | Tc8 | J. virginiana | 24.25 | 0.077 | 98 | 0.076 | 0.038 |
| Transition | Tc8 | J. virginiana | 18.19 | 0.058 | 98 | 0.057 | 0.028 |
| Transition | Tc8 | J. virginiana | 74.39 | 0.237 | 98 | 0.232 | 0.116 |
| Transition | Tc8 | J. virginiana | 9.10 | 0.029 | 98 | 0.028 | 0.014 |
| Transition | Tc8 | J. virginiana | 12.13 | 0.039 | 98 | 0.038 | 0.019 |
| Transition | Tc8 | J. virginiana | 15.16 | 0.048 | 98 | 0.047 | 0.024 |
| Total |  |  | 1652.44 | 5.26 |  | 5.16 | 2.58 |
| Transition | Tc18 | J. virginiana | 30.32 | 0.097 | 98 | 0.095 | 0.047 |
| Transition | Tc18 | J. virginiana | 15.16 | 0.048 | 98 | 0.047 | 0.024 |
| Transition | Tc18 | J. virginiana | 109.00 | 0.347 | 98 | 0.340 | 0.170 |
| Transition | Tc18 | J. virginiana | 18.19 | 0.058 | 98 | 0.057 | 0.028 |
| Transition | Tc18 | J. virginiana | 24.25 | 0.077 | 98 | 0.076 | 0.038 |
| Transition | Tc18 | J. virginiana | 27.29 | 0.087 | 98 | 0.085 | 0.043 |
| Transition | Tc18 | J. virginiana | 18.19 | 0.058 | 98 | 0.057 | 0.028 |
| Transition | Tc18 | J. virginiana | 15.16 | 0.048 | 98 | 0.047 | 0.024 |
| Transition | Tc18 | J. virginiana | 109.00 | 0.347 | 98 | 0.340 | 0.170 |
| Transition | Tc18 | J. virginiana | 9.10 | 0.029 | 98 | 0.028 | 0.014 |
| Transition | Tc18 | J. virginiana | 9.10 | 0.029 | 98 | 0.028 | 0.014 |
| Transition | Tc18 | J. virginiana | 12.13 | 0.039 | 98 | 0.038 | 0.019 |
| Transition | Tc18 | J. virginiana | 7.70 | 0.025 | 98 | 0.024 | 0.012 |
| Transition | Tc18 | J. virginiana | 12.13 | 0.039 | 98 | 0.038 | 0.019 |
| Transition | Tc18 | J. virginiana | 9.10 | 0.029 | 98 | 0.028 | 0.014 |
| Transition | Tc18 | J. virginiana | 7.70 | 0.025 | 98 | 0.024 | 0.012 |
| Transition | Tc18 | J. virginiana | 7.70 | 0.025 | 98 | 0.024 | 0.012 |
| Transition | Tc18 | J. virginiana | 18.19 | 0.058 | 98 | 0.057 | 0.028 |
| Transition | Tc18 | J. virginiana | 27.29 | 0.087 | 98 | 0.085 | 0.043 |
| Transition | Tc18 | J. virginiana | 15.16 | 0.048 | 98 | 0.047 | 0.024 |
| Transition | Tc18 | J. virginiana | 18.19 | 0.058 | 98 | 0.057 | 0.028 |
| Transition | Tc18 | J. virginiana | 15.16 | 0.048 | 98 | 0.047 | 0.024 |
| Transition | Tc18 | J. virginiana | 18.19 | 0.058 | 98 | 0.057 | 0.028 |
| Transition | Tc18 | J. virginiana | 33.35 | 0.106 | 98 | 0.104 | 0.052 |
| Transition | Tc18 | J. virginiana | 15.16 | 0.048 | 98 | 0.047 | 0.024 |
| Transition | Tc18 | J. virginiana | 30.32 | 0.097 | 98 | 0.095 | 0.047 |
| Transition | Tc18 | J. virginiana | 30.32 | 0.097 | 98 | 0.095 | 0.047 |


| Zone | Site | Species | Dry wt. $\mathrm{kg}$ | Dry wt. $\mathrm{kg} / \mathrm{m}^{2}$ | \% <br> Organic Matter | AFDW $\mathrm{kg} / \mathrm{m}^{2}$ | Organic carbon $\mathrm{kg} / \mathrm{m}^{2}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Transition | Tc18 | J. virginiana | 9.10 | 0.029 | 98 | 0.028 | 0.014 |
| Transition | Tc18 | P. taeda | 9.60 | 0.031 | 98 | 0.030 | 0.015 |
| Transition | Tc18 | P. taeda | 150.06 | 0.478 | 98 | 0.468 | 0.234 |
| Total |  |  | 831.28 | 2.65 |  | 2.59 | 1.30 |
| Transition | Tc31 | J. virginiana | 49.44 | 0.157 | 98 | 0.1543 | 0.077 |
| Transition | Tc31 | J. virginiana | 33.35 | 0.106 | 98 | 0.10409 | 0.052 |
| Transition | Tc31 | J. virginiana | 36.38 | 0.116 | 98 | 0.11355 | 0.057 |
| Transition | Tc31 | J. virginiana | 7.70 | 0.025 | 98 | 0.02403 | 0.012 |
| Transition | Tc31 | J. virginiana | 24.25 | 0.077 | 98 | 0.0757 | 0.038 |
| Transition | Tc31 | J. virginiana | 12.13 | 0.039 | 98 | 0.03785 | 0.019 |
| Transition | Tc31 | J. virginiana | 33.35 | 0.106 | 98 | 0.10409 | 0.052 |
| Total |  |  | 196.60 | 0.63 |  | 0.61 | 0.31 |
| Transition | Td2 | P. taeda | 1.76 | 0.006 | 98 | 0.005 | 0.003 |
| Transition | Td2 | P. taeda | 4.19 | 0.013 | 98 | 0.013 | 0.007 |
| Transition | Td2 | P. taeda | 1.76 | 0.006 | 98 | 0.005 | 0.003 |
| Transition | Td2 | P. taeda | 1.76 | 0.006 | 98 | 0.005 | 0.003 |
| Transition | Td2 | P. taeda | 2.06 | 0.007 | 98 | 0.006 | 0.003 |
| Transition | Td2 | P. taeda | 1.15 | 0.004 | 98 | 0.004 | 0.002 |
| Transition | Td2 | P. taeda | 62.42 | 0.199 | 98 | 0.195 | 0.097 |
| Transition | Td2 | P. taeda | 86.99 | 0.277 | 98 | 0.272 | 0.136 |
| Transition | Td2 | P. taeda | 169.74 | 0.541 | 98 | 0.530 | 0.265 |
| Transition | Td2 | P. taeda | 140.49 | 0.447 | 98 | 0.438 | 0.219 |
| Transition | Td2 | P. taeda | 102.52 | 0.327 | 98 | 0.320 | 0.160 |
| Transition | Td2 | P. taeda | 158.78 | 0.506 | 98 | 0.496 | 0.248 |
| Transition | Td2 | P. taeda | 33.04 | 0.105 | 98 | 0.103 | 0.052 |
| Transition | Td2 | P. taeda | 3.34 | 0.011 | 98 | 0.010 | 0.005 |
| Transition | Td2 | P. taeda | 22.18 | 0.071 | 98 | 0.069 | 0.035 |
| Transition | Td2 | P. taeda | 2.35 | 0.007 | 98 | 0.007 | 0.004 |
| Transition | Td2 | P. taeda | 3.34 | 0.011 | 98 | 0.010 | 0.005 |
| Transition | Td2 | P. taeda | 1.34 | 0.004 | 98 | 0.004 | 0.002 |
| Transition | Td2 | P. taeda | 1.34 | 0.004 | 98 | 0.004 | 0.002 |
| Transition | Td2 | P. taeda | 1.34 | 0.004 | 98 | 0.004 | 0.002 |
| Transition | Td2 | J. virginiana | 15.16 | 0.048 | 98 | 0.047 | 0.024 |
| Transition | Td2 | J. virginiana | 74.39 | 0.237 | 98 | 0.232 | 0.116 |
| Transition | Td2 | J. virginiana | 18.19 | 0.058 | 98 | 0.057 | 0.028 |
| Transition | Td2 | J. virginiana | 9.10 | 0.029 | 98 | 0.028 | 0.014 |
| Transition | Td2 | J. virginiana | 21.22 | 0.068 | 98 | 0.066 | 0.033 |
| Transition | Td2 | J. virginiana | 21.22 | 0.068 | 98 | 0.066 | 0.033 |
| Transition | Td2 | J. virginiana | 18.19 | 0.058 | 98 | 0.057 | 0.028 |
| Transition | Td2 | J. virginiana | 18.19 | 0.058 | 98 | 0.057 | 0.028 |


| Zone | Site | Species | Dry wt. kg | Dry wt. $\mathrm{kg} / \mathrm{m}^{2}$ |  | AFDW $\mathrm{kg} / \mathrm{m}^{2}$ | Organic carbon $\mathrm{kg} / \mathrm{m}^{2}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Transition | Td2 | J. virginiana | 49.44 | 0.157 | 98 | 0.154 | 0.077 |
| Transition | Td2 | J. virginiana | 9.10 | 0.029 | 98 | 0.028 | 0.014 |
| Transition | Td2 | J. virginiana | 36.38 | 0.116 | 98 | 0.114 | 0.057 |
| Transition | Td2 | J. virginiana | 30.32 | 0.097 | 98 | 0.095 | 0.047 |
| Transition | Td2 | J. virginiana | 33.35 | 0.106 | 98 | 0.104 | 0.052 |
| Total |  |  | 1156.11 | 3.68 |  | 3.61 | 1.80 |
| Transition | Td9 | J. virginiana | 9.10 | 0.029 | 98 | 0.03 | 0.014 |
| Transition | Td9 | J. virginiana | 9.10 | 0.029 | 98 | 0.03 | 0.014 |
| Transition | Td9 | J. virginiana | 12.13 | 0.039 | 98 | 0.04 | 0.019 |
| Transition | Td9 | J. virginiana | 12.13 | 0.039 | 98 | 0.04 | 0.019 |
| Transition | Td9 | J. virginiana | 27.29 | 0.087 | 98 | 0.09 | 0.043 |
| Transition | Td9 | J. virginiana | 33.35 | 0.106 | 98 | 0.10 | 0.052 |
| Transition | Td9 | J. virginiana | 74.39 | 0.237 | 98 | 0.23 | 0.116 |
| Transition | Td9 | J. virginiana | 18.19 | 0.058 | 98 | 0.06 | 0.028 |
| Transition | Td9 | J. virginiana | 30.32 | 0.097 | 98 | 0.09 | 0.047 |
| Transition | Td9 | J. virginiana | 30.32 | 0.097 | 98 | 0.09 | 0.047 |
| Transition | Td9 | J. virginiana | 36.38 | 0.116 | 98 | 0.11 | 0.057 |
| Transition | Td9 | J. virginiana | 18.19 | 0.058 | 98 | 0.06 | 0.028 |
| Transition | Td9 | J. virginiana | 21.22 | 0.068 | 98 | 0.07 | 0.033 |
| Transition | Td9 | J. virginiana | 74.39 | 0.237 | 98 | 0.23 | 0.116 |
| Transition | Td9 | J. virginiana | 33.35 | 0.106 | 98 | 0.10 | 0.052 |
| Transition | Td9 | J. virginiana | 74.39 | 0.237 | 98 | 0.23 | 0.116 |
| Transition | Td9 | J. virginiana | 36.38 | 0.116 | 98 | 0.11 | 0.057 |
| Transition | Td9 | J. virginiana | 18.19 | 0.058 | 98 | 0.06 | 0.028 |
| Transition | Td9 | J. virginiana | 30.32 | 0.097 | 98 | 0.09 | 0.047 |
| Transition | Td9 | J. virginiana | 104.78 | 0.334 | 98 | 0.33 | 0.164 |
| Transition | Td9 | J. virginiana | 21.22 | 0.068 | 98 | 0.07 | 0.033 |
| Transition | Td9 | J. virginiana | 18.19 | 0.058 | 98 | 0.06 | 0.028 |
| Transition | Td9 | J. virginiana | 74.39 | 0.237 | 98 | 0.23 | 0.116 |
| Transition | Td9 | J. virginiana | 21.22 | 0.068 | 98 | 0.07 | 0.033 |
| Transition | Td9 | J. virginiana | 49.44 | 0.157 | 98 | 0.15 | 0.077 |
| Transition | Td9 | J. virginiana | 24.25 | 0.077 | 98 | 0.08 | 0.038 |
| Transition | Td9 | P. taeda | 1.53 | 0.005 | 98 | 0.00 | 0.002 |
| Transition | Td9 | P. taeda | 2.35 | 0.007 | 98 | 0.01 | 0.004 |
| Transition | Td9 | P. taeda | 17.94 | 0.057 | 98 | 0.06 | 0.028 |
| Transition | Td9 | P. taeda | 0.48 | 0.002 | 98 | 0.00 | 0.001 |
| Transition | Td9 | P. taeda | 1.53 | 0.005 | 98 | 0.00 | 0.002 |
| Transition | Td9 | P. taeda | 5.63 | 0.018 | 98 | 0.02 | 0.009 |
| Transition | Td9 | P. taeda | 0.88 | 0.003 | 98 | 0.00 | 0.001 |
| Transition | Td9 | P. taeda | 8.74 | 0.028 | 98 | 0.03 | 0.014 |
| Transition | Td9 | P. taeda | 5.63 | 0.018 | 98 | 0.02 | 0.009 |
| Total |  |  | 957.33 | 3.05 |  | 2.99 | 1.49 |


| Zone | Site | Species | Dry wt. kg | Dry wt. $\mathrm{kg} / \mathrm{m}^{2}$ | \% <br> Organic Matter | AFDW $\mathrm{kg} / \mathrm{m}^{2}$ | Organic carbon $\mathrm{kg} / \mathrm{m}^{2}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Transition | Td15 | J. virginiana | 21.22 | 0.068 | 98 | 0.066 | 0.033 |
| Transition | Td15 | J. virginiana | 15.16 | 0.048 | 98 | 0.047 | 0.024 |
| Transition | Td15 | J. virginiana | 18.19 | 0.058 | 98 | 0.057 | 0.028 |
| Transition | Td15 | J. virginiana | 15.16 | 0.048 | 98 | 0.047 | 0.024 |
| Transition | Td15 | J. virginiana | 15.16 | 0.048 | 98 | 0.047 | 0.024 |
| Transition | Td15 | J. virginiana | 9.10 | 0.029 | 98 | 0.028 | 0.014 |
| Transition | Td15 | J. virginiana | 24.25 | 0.077 | 98 | 0.076 | 0.038 |
| Transition | Td15 | J. virginiana | 15.16 | 0.048 | 98 | 0.047 | 0.024 |
| Transition | Td15 | J. virginiana | 9.10 | 0.029 | 98 | 0.028 | 0.014 |
| Transition | Td15 | J. virginiana | 27.29 | 0.087 | 98 | 0.085 | 0.043 |
| Transition | Td15 | J. virginiana | 12.13 | 0.039 | 98 | 0.038 | 0.019 |
| Transition | Td15 | J. virginiana | 24.25 | 0.077 | 98 | 0.076 | 0.038 |
| Transition | Td15 | J. virginiana | 27.29 | 0.087 | 98 | 0.085 | 0.043 |
| Transition | Td15 | J. virginiana | 18.19 | 0.058 | 98 | 0.057 | 0.028 |
| Transition | Td15 | J. virginiana | 33.35 | 0.106 | 98 | 0.104 | 0.052 |
| Transition | Td15 | J. virginiana | 15.16 | 0.048 | 98 | 0.047 | 0.024 |
| Transition | Td15 | J. virginiana | 21.22 | 0.068 | 98 | 0.066 | 0.033 |
| Transition | Td15 | J. virginiana | 12.13 | 0.039 | 98 | 0.038 | 0.019 |
| Transition | Td15 | J. virginiana | 12.13 | 0.039 | 98 | 0.038 | 0.019 |
| Transition | Td15 | J. virginiana | 33.35 | 0.106 | 98 | 0.104 | 0.052 |
| Transition | Td15 | J. virginiana | 30.32 | 0.097 | 98 | 0.095 | 0.047 |
| Transition | Td15 | J. virginiana | 9.10 | 0.029 | 98 | 0.028 | 0.014 |
| Transition | Td15 | J. virginiana | 27.29 | 0.087 | 98 | 0.085 | 0.043 |
| Transition | Td15 | J. virginiana | 12.13 | 0.039 | 98 | 0.038 | 0.019 |
| Transition | Td15 | J. virginiana | 302.55 | 0.964 | 98 | 0.944 | 0.472 |
| Transition | Td15 | J. virginiana | 162.39 | 0.517 | 98 | 0.507 | 0.253 |
| Transition | Td15 | J. virginiana | 192.78 | 0.614 | 98 | 0.602 | 0.301 |
| Transition | Td15 | J. virginiana | 302.55 | 0.964 | 98 | 0.944 | 0.472 |
| Transition | Td15 | P. taeda | 70.89 | 0.226 | 98 | 0.221 | 0.111 |
| Transition | Td15 | P. taeda | 72.40 | 0.231 | 98 | 0.226 | 0.113 |
| Total |  |  | 1561.35 | 4.97 |  | 4.87 | 2.44 |
| Transition | Td33 | P. taeda | 12.28 | 0.039 | 98 | 0.038 | 0.019 |
| Transition | Td33 | P. taeda | 10.03 | 0.032 | 98 | 0.031 | 0.016 |
| Transition | Td33 | P. taeda | 2.50 | 0.008 | 98 | 0.008 | 0.004 |
| Transition | Td33 | P. taeda | 11.15 | 0.036 | 98 | 0.035 | 0.017 |
| Transition | Td33 | J. virginiana | 24.25 | 0.077 | 98 | 0.076 | 0.038 |
| Transition | Td33 | J. virginiana | 15.16 | 0.048 | 98 | 0.047 | 0.024 |
| Transition | Td33 | J. virginiana | 15.16 | 0.048 | 98 | 0.047 | 0.024 |
| Transition | Td33 | J. virginiana | 18.19 | 0.058 | 98 | 0.057 | 0.028 |
| Transition | Td33 | J. virginiana | 24.25 | 0.077 | 98 | 0.076 | 0.038 |
| Transition | Td33 | J. virginiana | 21.22 | 0.068 | 98 | 0.066 | 0.033 |


| Zone | Site | Species | $\begin{gathered} \text { Dry wt. } \\ \text { kg } \\ \hline \hline \end{gathered}$ | Dry wt. $\mathrm{kg} / \mathrm{m}^{2}$ |  | AFDW $\mathrm{kg} / \mathrm{m}^{2}$ | Organic carbon $\mathrm{kg} / \mathrm{m}^{2}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Transition | Td33 | J. virginiana | 36.38 | 0.116 | 98 | 0.114 | 0.057 |
| Transition | Td33 | J. virginiana | 24.25 | 0.077 | 98 | 0.076 | 0.038 |
| Transition | Td33 | J. virginiana | 24.25 | 0.077 | 98 | 0.076 | 0.038 |
| Total |  |  | 239.09 | 0.76 |  | 0.75 | 0.37 |
| Transition | Td39 | P. taeda | 44.71 | 0.142378 | 98 | 0.140 | 0.070 |
| Transition | Td39 | P. taeda | 147.35 | 0.46928 | 98 | 0.460 | 0.230 |
| Transition | Td39 | J. virginiana | 140.62 | 0.447819 | 98 | 0.439 | 0.219 |
| Transition | Td39 | J. virginiana | 15.16 | 0.048277 | 98 | 0.047 | 0.024 |
| Transition | Td39 | J. virginiana | 12.13 | 0.038622 | 98 | 0.038 | 0.019 |
| Transition | Td39 | J. virginiana | 12.13 | 0.038622 | 98 | 0.038 | 0.019 |
| Transition | Td39 | J. virginiana | 18.19 | 0.057932 | 98 | 0.057 | 0.028 |
| Transition | Td39 | J. virginiana | 51.58 | 0.164268 | 98 | 0.161 | 0.080 |
| Transition | Td39 | J. virginiana | 66.16 | 0.210701 | 98 | 0.206 | 0.103 |
| Transition | Td39 | J. virginiana | 24.25 | 0.077243 | 98 | 0.076 | 0.038 |
| Transition | Td39 | J. virginiana | 27.29 | 0.086899 | 98 | 0.085 | 0.043 |
| Transition | Td39 | J. virginiana | 30.32 | 0.096554 | 98 | 0.095 | 0.047 |
| Transition | Td39 | J. virginiana | 9.10 | 0.028966 | 98 | 0.028 | 0.014 |
| Transition | Td39 | J. virginiana | 12.13 | 0.038622 | 98 | 0.038 | 0.019 |
| Transition | Td39 | J. virginiana | 12.13 | 0.038622 | 98 | 0.038 | 0.019 |
| Transition | Td39 | J. virginiana | 9.10 | 0.028966 | 98 | 0.028 | 0.014 |
| Transition | Td39 | J. virginiana | 33.35 | 0.10621 | 98 | 0.104 | 0.052 |
| Transition | Td39 | J. virginiana | 33.35 | 0.10621 | 98 | 0.104 | 0.052 |
| Transition | Td39 | J. virginiana | 27.29 | 0.086899 | 98 | 0.085 | 0.043 |
| Transition | Td39 | J. virginiana | 15.16 | 0.048277 | 98 | 0.047 | 0.024 |
| Transition | Td39 | J. virginiana | 49.44 | 0.157452 | 98 | 0.154 | 0.077 |
| Transition | Td39 | J. virginiana | 104.78 | 0.333694 | 98 | 0.327 | 0.164 |
| Transition | Td39 | J. virginiana | 27.29 | 0.086899 | 98 | 0.085 | 0.043 |
| Transition | Td39 | J. virginiana | 36.38 | 0.115865 | 98 | 0.114 | 0.057 |
| Transition | Td39 | J. virginiana | 104.78 | 0.333694 | 98 | 0.327 | 0.164 |
| Transition | Td39 | J. virginiana | 18.19 | 0.057932 | 98 | 0.057 | 0.028 |
| Transition | Td39 | J. virginiana | 15.16 | 0.048277 | 98 | 0.047 | 0.024 |
| Transition | Td39 | J. virginiana | 12.13 | 0.038622 | 98 | 0.038 | 0.019 |
| Transition | Td39 | J. virginiana | 9.10 | 0.028966 | 98 | 0.028 | 0.014 |
| Transition | Td39 | J. virginiana | 41.87 | 0.133344 | 98 | 0.131 | 0.065 |
| Transition | Td39 | J. virginiana | 27.29 | 0.086899 | 98 | 0.085 | 0.043 |
| Transition | Td39 | J. virginiana | 15.16 | 0.048277 | 98 | 0.047 | 0.024 |
| Transition | Td39 | J. virginiana | 12.13 | 0.038622 | 98 | 0.038 | 0.019 |
| Transition | Td39 | J. virginiana | 30.32 | 0.096554 | 98 | 0.095 | 0.047 |
| Transition | Td39 | J. virginiana | 27.29 | 0.086899 | 98 | 0.085 | 0.043 |
| Transition | Td39 | J. virginiana | 30.32 | 0.096554 | 98 | 0.095 | 0.047 |
| Transition | Td39 | J. virginiana | 140.62 | 0.447834 | 98 | 0.439 | 0.219 |
| Total |  |  | 1443.69 | 4.60 |  | 4.51 | 2.25 |


|  |  |  |  |  | Dry wt. <br> kg | Dry wt. <br> $\mathrm{kg} / \mathrm{m}^{2}$ | Organic <br> Matter |
| :--- | :--- | :--- | :---: | :---: | :---: | :---: | :---: |
| Zone | Site | Species | AFDW <br> $\mathrm{kg} / \mathrm{m}^{2}$ | Organic <br> carbon <br> $\mathrm{kg} / \mathrm{m}^{2}$ |  |  |  |
| Transition | Td50 | J. virginiana | 24.25 | 0.077 | 98 | 0.076 | 0.038 |
| Transition | Td50 | J. virginiana | 33.35 | 0.106 | 98 | 0.104 | 0.052 |
| Transition | Td50 | J. virginiana | 33.35 | 0.106 | 98 | 0.104 | 0.052 |
| Transition | Td50 | J. virginiana | 27.29 | 0.087 | 98 | 0.085 | 0.043 |
| Transition | Td50 | P. taeda | 22.18 | 0.071 | 98 | 0.069 | 0.035 |
| Total |  |  | 140.42 | 0.45 |  | 0.44 | 0.22 |

APPENDIX C. SUMMATION OF BIOMASS, ASF-FREE DRY WEIGHT, AND CARRON MA.S.S OF SHRIIR.S PFR FORFST AND TRAN.SITON SITF

| Zone | Site | Species | Biomass kg | $\begin{array}{r} \text { Biomass } \\ \mathrm{kg} / \mathrm{m}^{2} \\ \hline \hline \end{array}$ |  | AFDW $\mathrm{kg} / \mathrm{m}^{2}$ | Organic carbon $\mathrm{kg} / \mathrm{m}^{2}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Forest | Fa18 | N/A | 0.0 | 0.0 |  | 0.0 | 0.0 |
| Forest | Fa30 | N/A | 0.0 | 0.0 |  | 0.0 | 0.0 |
| Forest | Fb29 | Myrica cerifera | 3.68 | 0.05 | 98 | 0.05 | 0.023 |
| Forest | Fb29 | M. cerifera | 2.78 | 0.04 | 98 | 0.03 | 0.017 |
| Forest | Fb29 | M. cerifera | 1.88 | 0.02 | 98 | 0.02 | 0.012 |
| Forest | Fb29 | M. cerifera | 3.68 | 0.05 | 98 | 0.05 | 0.023 |
| Forest | Fb29 | M. cerifera | 3.68 | 0.05 | 98 | 0.05 | 0.023 |
| Forest | Fb29 | M. cerifera | 0.56 | 0.01 | 98 | 0.01 | 0.003 |
| Total |  |  | 16.27 | 0.21 |  | 0.20 | 0.10 |
| Forest | Fb47 |  | 0.0 | 0.0 |  | 0.0 | 0.0 |
| Forest | Fb65 | M. cerifera | 3.68 | 0.05 | 98 | 0.05 | 0.023 |
| Forest | Fb65 | M. cerifera | 2.78 | 0.04 | 98 | 0.03 | 0.017 |
| Forest | Fb65 | M. cerifera | 3.68 | 0.05 | 98 | 0.05 | 0.023 |
| Forest | Fb65 | M. cerifera | 3.68 | 0.05 | 98 | 0.05 | 0.023 |
| Forest | Fb65 | M. cerifera | 0.97 | 0.01 | 98 | 0.01 | 0.006 |
| Total |  |  | 14.80 | 0.19 |  | 0.18 | 0.09 |
| Transition | Ta1 | Baccharis halimifolia | 0.46 | 0.006 | 98 | 0.01 | 0.003 |
| Transition | Ta1 | B. halimifolia | 1.02 | 0.013 | 98 | 0.01 | 0.006 |
| Transition | Ta1 | B. halimifolia | 1.02 | 0.013 | 98 | 0.01 | 0.006 |
| Transition | Ta1 | B. halimifolia | 0.37 | 0.005 | 98 | 0.00 | 0.002 |
| Total |  |  | 2.86 | 0.04 |  | 0.04 | 0.02 |
| Transition | Ta4 | B. halimifolia | 0.46 | 0.006 | 98 | 0.006 | 0.003 |
| Transition | Ta4 | B. halimifolia | 1.20 | 0.015 | 98 | 0.015 | 0.007 |
| Transition | Ta4 | B. halimifolia | 0.37 | 0.005 | 98 | 0.005 | 0.002 |
| Total |  |  | 2.03 | 0.03 | 98 | 0.03 | 0.01 |
| Transition | Ta13 | Iva frutescens | 0.24 | 0.003 | 98 | 0.003 | 0.002 |
| Transition | Ta13 | l. frutescens | 0.17 | 0.002 | 98 | 0.002 | 0.001 |
| Transition | Ta13 | I. frutescens | 0.69 | 0.009 | 98 | 0.009 | 0.004 |
| Transition | Ta13 | l. frutescens | 0.69 | 0.009 | 98 | 0.009 | 0.004 |
| Transition | Ta13 | l. frutescens | 0.52 | 0.007 | 98 | 0.007 | 0.003 |
| Transition | Ta13 | l. frutescens | 0.24 | 0.003 | 98 | 0.003 | 0.002 |
| Transition | Ta13 | I. frutescens | 0.35 | 0.004 | 98 | 0.004 | 0.002 |
| Transition | Ta13 | I. frutescens | 0.24 | 0.003 | 98 | 0.003 | 0.002 |
| Transition | Ta13 | l. frutescens | 0.90 | 0.012 | 98 | 0.011 | 0.006 |


| Zone | Site | Species | Biomass kg | Biomass $\mathrm{kg} / \mathrm{m}^{2}$ |  | AFDW $\mathrm{kg} / \mathrm{m}^{2}$ | Organic carbon $\mathrm{kg} / \mathrm{m}^{2}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Transition | Ta13 | I. frutescens | 0.69 | 0.009 | 98 | 0.009 | 0.004 |
| Transition | Ta13 | I. frutescens | 0.35 | 0.004 | 98 | 0.004 | 0.002 |
| Transition | Ta13 | I. frutescens | 0.17 | 0.002 | 98 | 0.002 | 0.001 |
| Transition | Ta13 | I. frutescens | 0.52 | 0.007 | 98 | 0.007 | 0.003 |
| Transition | Ta13 | I. frutescens | 1.04 | 0.013 | 98 | 0.013 | 0.007 |
| Transition | Ta13 | I. frutescens | 0.90 | 0.012 | 98 | 0.011 | 0.006 |
| Transition | Ta13 | I. frutescens | 0.83 | 0.011 | 98 | 0.010 | 0.005 |
| Transition | Ta13 | I. frutescens | 0.83 | 0.011 | 98 | 0.010 | 0.005 |
| Transition | Ta13 | I. frutescens | 1.22 | 0.015 | 98 | 0.015 | 0.008 |
| Transition | Ta13 | I. frutescens | 0.69 | 0.009 | 98 | 0.009 | 0.004 |
| Transition | Ta13 | I. frutescens | 0.17 | 0.002 | 98 | 0.002 | 0.001 |
| Transition | Ta13 | I. frutescens | 0.52 | 0.007 | 98 | 0.007 | 0.003 |
| Transition | Ta13 | B. halimifolia | 1.02 | 0.013 | 98 | 0.013 | 0.006 |
| Transition | Ta13 | B. halimifolia | 0.74 | 0.009 | 98 | 0.009 | 0.005 |
| Transition | Ta13 | B. halimifolia | 1.58 | 0.020 | 98 | 0.020 | 0.010 |
| Total |  |  | 15.36 | 0.20 |  | 0.19 | 0.10 |
| Transition | Tb29 | I. frutescens | 0.23 | 0.003 | 98 | 0.003 | 0.001 |
| Transition | Tb29 | I. frutescens | 0.23 | 0.003 | 98 | 0.003 | 0.001 |
| Transition | Tb29 | I. frutescens | 0.23 | 0.003 | 98 | 0.003 | 0.001 |
| Transition | Tb29 | I. frutescens | 0.23 | 0.003 | 98 | 0.003 | 0.001 |
| Transition | Tb29 | I. frutescens | 0.23 | 0.003 | 98 | 0.003 | 0.001 |
| Transition | Tb29 | I. frutescens | 0.46 | 0.006 | 98 | 0.006 | 0.003 |
| Transition | Tb29 | I. frutescens | 0.69 | 0.009 | 98 | 0.009 | 0.004 |
| Transition | Tb29 | I. frutescens | 0.69 | 0.009 | 98 | 0.009 | 0.004 |
| Transition | Tb29 | I. frutescens | 0.46 | 0.006 | 98 | 0.006 | 0.003 |
| Transition | Tb29 | I. frutescens | 0.46 | 0.006 | 98 | 0.006 | 0.003 |
| Transition | Tb29 | I. frutescens | 0.23 | 0.003 | 98 | 0.003 | 0.001 |
| Transition | Tb29 | I. frutescens | 0.23 | 0.003 | 98 | 0.003 | 0.001 |
| Transition | Tb29 | I. frutescens | 0.23 | 0.003 | 98 | 0.003 | 0.001 |
| Transition | Tb29 | I. frutescens | 0.46 | 0.006 | 98 | 0.006 | 0.003 |
| Transition | Tb29 | I. frutescens | 0.46 | 0.006 | 98 | 0.006 | 0.003 |
| Transition | Tb29 | I. frutescens | 0.46 | 0.006 | 98 | 0.006 | 0.003 |
| Transition | Tb29 | I. frutescens | 0.46 | 0.006 | 98 | 0.006 | 0.003 |
| Transition | Tb29 | I. frutescens | 0.23 | 0.003 | 98 | 0.003 | 0.001 |
| Transition | Tb29 | I. frutescens | 0.46 | 0.006 | 98 | 0.006 | 0.003 |
| Transition | Tb29 | I. frutescens | 0.46 | 0.006 | 98 | 0.006 | 0.003 |
| Transition | Tb29 | I. frutescens | 0.23 | 0.003 | 98 | 0.003 | 0.001 |
| Transition | Tb29 | I. frutescens | 0.46 | 0.006 | 98 | 0.006 | 0.003 |
| Transition | Tb29 | I. frutescens | 0.23 | 0.003 | 98 | 0.003 | 0.001 |
| Transition | Tb29 | I. frutescens | 0.23 | 0.003 | 98 | 0.003 | 0.001 |
| Transition | Tb29 | I. frutescens | 0.23 | 0.003 | 98 | 0.003 | 0.001 |
| Transition | Tb29 | I. frutescens | 0.46 | 0.006 | 98 | 0.006 | 0.003 |
| Transition | Tb29 | I. frutescens | 0.23 | 0.003 | 98 | 0.003 | 0.001 |


| Zone | Site | Species | Biomass kg | $\begin{array}{r} \text { Biomass } \\ \mathrm{kg} / \mathrm{m}^{2} \\ \hline \hline \end{array}$ |  | AFDW $\mathrm{kg} / \mathrm{m}^{2}$ | Organic carbon $\mathrm{kg} / \mathrm{m}^{2}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Transition | Tb29 | I. frutescens | 0.69 | 0.009 | 98 | 0.009 | 0.004 |
| Transition | Tb29 | I. frutescens | 0.46 | 0.006 | 98 | 0.006 | 0.003 |
| Transition | Tb29 | I. frutescens | 0.46 | 0.006 | 98 | 0.006 | 0.003 |
| Transition | Tb29 | I. frutescens | 0.23 | 0.003 | 98 | 0.003 | 0.001 |
| Transition | Tb29 | I. frutescens | 0.23 | 0.003 | 98 | 0.003 | 0.001 |
| Transition | Tb29 | I. frutescens | 0.46 | 0.006 | 98 | 0.006 | 0.003 |
| Transition | Tb29 | I. frutescens | 0.23 | 0.003 | 98 | 0.003 | 0.001 |
| Transition | Tb29 | I. frutescens | 1.19 | 0.015 | 98 | 0.015 | 0.007 |
| Transition | Tb29 | I. frutescens | 1.10 | 0.014 | 98 | 0.014 | 0.007 |
| Transition | Tb29 | l. frutescens | 0.92 | 0.012 | 98 | 0.011 | 0.006 |
| Transition | Tb29 | I. frutescens | 0.46 | 0.006 | 98 | 0.006 | 0.003 |
| Transition | Tb29 | I. frutescens | 0.23 | 0.003 | 98 | 0.003 | 0.001 |
| Transition | Tb29 | I. frutescens | 0.23 | 0.003 | 98 | 0.003 | 0.001 |
| Transition | Tb29 | I. frutescens | 0.23 | 0.003 | 98 | 0.003 | 0.001 |
| Transition | Tb29 | I. frutescens | 0.23 | 0.003 | 98 | 0.003 | 0.001 |
| Total |  |  | 16.99 | 0.22 |  | 0.21 | 0.11 |
| Transition | Tc2 | B. halimifolia | 0.37 | 0.005 | 98 | 0.005 | 0.002 |
| Transition | Tc2 | B. halimifolia | 0.46 | 0.006 | 98 | 0.006 | 0.003 |
| Transition | Tc2 | B. halimifolia | 0.12 | 0.002 | 98 | 0.001 | 0.001 |
| Transition | Tc2 | B. halimifolia | 0.78 | 0.010 | 98 | 0.010 | 0.005 |
| Transition | Tc2 | M. cerifera | 1.88 | 0.024 | 98 | 0.023 | 0.012 |
| Total |  |  | 3.61 | 0.05 |  | 0.05 | 0.02 |
| Transition | Tc8 | M. cerifera | 0.97 | 0.012 | 98 | 0.012 | 0.006 |
| Transition | Tc8 | M. cerifera | 0.56 | 0.007 | 98 | 0.007 | 0.003 |
| Transition | Tc8 | M. cerifera | 1.88 | 0.024 | 98 | 0.023 | 0.012 |
| Transition | Tc8 | M. cerifera | 6.39 | 0.081 | 98 | 0.080 | 0.040 |
| Transition | Tc8 | M. cerifera | 7.30 | 0.093 | 98 | 0.091 | 0.046 |
| Transition | Tc8 | M. cerifera | 2.78 | 0.035 | 98 | 0.035 | 0.017 |
| Transition | Tc8 | M. cerifera | 2.78 | 0.035 | 98 | 0.035 | 0.017 |
| Transition | Tc8 | M. cerifera | 0.56 | 0.007 | 98 | 0.007 | 0.003 |
| Transition | Tc8 | M. cerifera | 0.97 | 0.012 | 98 | 0.012 | 0.006 |
| Transition | Tc8 | M. cerifera | 1.88 | 0.024 | 98 | 0.023 | 0.012 |
| Transition | Tc8 | M. cerifera | 0.97 | 0.012 | 98 | 0.012 | 0.006 |
| Transition | Tc8 | M. cerifera | 6.39 | 0.081 | 98 | 0.080 | 0.040 |
| Transition | Tc8 | M. cerifera | 7.30 | 0.093 | 98 | 0.091 | 0.046 |
| Transition | Tc8 | M. cerifera | 0.97 | 0.012 | 98 | 0.012 | 0.006 |
| Transition | Tc8 | M. cerifera | 2.78 | 0.035 | 98 | 0.035 | 0.017 |
| Transition | Tc8 | M. cerifera | 2.78 | 0.035 | 98 | 0.035 | 0.017 |
| Transition | Tc8 | M. cerifera | 3.68 | 0.047 | 98 | 0.046 | 0.023 |
| Transition | Tc8 | M. cerifera | 2.78 | 0.035 | 98 | 0.035 | 0.017 |


| Zone | Site | Species | Biomass kg | Biomass $\mathrm{kg} / \mathrm{m}^{2}$ |  | AFDW $\mathrm{kg} / \mathrm{m}^{2}$ | Organic carbon $\mathrm{kg} / \mathrm{m}^{2}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Transition | Tc8 | M. cerifera | 0.97 | 0.012 | 98 | 0.012 | 0.006 |
| Transition | Tc8 | M. cerifera | 1.88 | 0.024 | 98 | 0.023 | 0.012 |
| Transition | Tc8 | M. cerifera | 2.78 | 0.035 | 98 | 0.035 | 0.017 |
| Transition | Tc8 | M. cerifera | 3.68 | 0.047 | 98 | 0.046 | 0.023 |
| Transition | Tc8 | M. cerifera | 0.56 | 0.007 | 98 | 0.007 | 0.003 |
| Transition | Tc8 | M. cerifera | 5.49 | 0.070 | 98 | 0.069 | 0.034 |
| Transition | Tc8 | M. cerifera | 7.30 | 0.093 | 98 | 0.091 | 0.046 |
| Total |  |  | 76.39 | 0.97 |  | 0.95 | 0.48 |
| Transition | Tc18 | M. cerifera | 3.68 | 0.047 | 98 | 0.046 | 0.023 |
| Transition | Tc18 | M. cerifera | 0.97 | 0.012 | 98 | 0.012 | 0.006 |
| Transition | Tc18 | M. cerifera | 2.78 | 0.035 | 98 | 0.035 | 0.017 |
| Transition | Tc18 | M. cerifera | 6.39 | 0.081 | 98 | 0.080 | 0.040 |
| Transition | Tc18 | M. cerifera | 2.78 | 0.035 | 98 | 0.035 | 0.017 |
| Transition | Tc18 | M. cerifera | 0.97 | 0.012 | 98 | 0.012 | 0.006 |
| Transition | Tc18 | M. cerifera | 0.97 | 0.012 | 98 | 0.012 | 0.006 |
| Transition | Tc18 | M. cerifera | 0.97 | 0.012 | 98 | 0.012 | 0.006 |
| Transition | Tc18 | M. cerifera | 2.78 | 0.035 | 98 | 0.035 | 0.017 |
| Transition | Tc18 | M. cerifera | 4.59 | 0.058 | 98 | 0.057 | 0.029 |
| Transition | Tc18 | M. cerifera | 6.39 | 0.081 | 98 | 0.080 | 0.040 |
| Transition | Tc18 | M. cerifera | 1.88 | 0.024 | 98 | 0.023 | 0.012 |
| Transition | Tc18 | M. cerifera | 4.59 | 0.058 | 98 | 0.057 | 0.029 |
| Transition | Tc18 | M. cerifera | 1.88 | 0.024 | 98 | 0.023 | 0.012 |
| Transition | Tc18 | M. cerifera | 1.88 | 0.024 | 98 | 0.023 | 0.012 |
| Transition | Tc18 | M. cerifera | 3.68 | 0.047 | 98 | 0.046 | 0.023 |
| Transition | Tc18 | M. cerifera | 0.56 | 0.007 | 98 | 0.007 | 0.003 |
| Transition | Tc18 | M. cerifera | 1.88 | 0.024 | 98 | 0.023 | 0.012 |
| Transition | Tc18 | M. cerifera | 2.78 | 0.035 | 98 | 0.035 | 0.017 |
| Transition | Tc18 | M. cerifera | 1.88 | 0.024 | 98 | 0.023 | 0.012 |
| Transition | Tc18 | M. cerifera | 2.78 | 0.035 | 98 | 0.035 | 0.017 |
| Transition | Tc18 | M. cerifera | 0.97 | 0.012 | 98 | 0.012 | 0.006 |
| Transition | Tc18 | M. cerifera | 0.97 | 0.012 | 98 | 0.012 | 0.006 |
| Transition | Tc18 | M. cerifera | 1.88 | 0.024 | 98 | 0.023 | 0.012 |
| Transition | Tc18 | M. cerifera | 1.88 | 0.024 | 98 | 0.023 | 0.012 |
| Total |  |  | 62.77 | 0.80 |  | 0.78 | 0.39 |
| Transition | Tc31 | I. frutescens | 1.61 | 0.020 | 98 | 0.020 | 0.010 |
| Transition | Tc31 | B. halimifolia | 0.46 | 0.006 | 98 | 0.006 | 0.003 |
| Transition | Tc31 | B. halimifolia | 0.78 | 0.010 | 98 | 0.010 | 0.005 |
| Total |  |  | 2.85 | 0.04 |  | 0.04 | 0.02 |
| Transition | Td2 | M. cerifera | 0.97 | 0.012 | 98 | 0.012 | 0.006 |
| Transition | Td2 | M. cerifera | 0.97 | 0.012 | 98 | 0.012 | 0.006 |


| Zone | Site | Species | Biomass kg | Biomass $\mathrm{kg} / \mathrm{m}^{2}$ | \% <br> Organic Matter | AFDW $\mathrm{kg} / \mathrm{m}^{2}$ | Organic carbon $\mathrm{kg} / \mathrm{m}^{2}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Transition | Td2 | M. cerifera | 2.78 | 0.035 | 98 | 0.035 | 0.017 |
| Transition | Td2 | M. cerifera | 0.97 | 0.012 | 98 | 0.012 | 0.006 |
| Transition | Td2 | M. cerifera | 3.68 | 0.047 | 98 | 0.046 | 0.023 |
| Transition | Td2 | M. cerifera | 0.97 | 0.012 | 98 | 0.012 | 0.006 |
| Transition | Td2 | M. cerifera | 0.97 | 0.012 | 98 | 0.012 | 0.006 |
| Transition | Td2 | M. cerifera | 0.97 | 0.012 | 98 | 0.012 | 0.006 |
| Transition | Td2 | M. cerifera | 2.78 | 0.035 | 98 | 0.035 | 0.017 |
| Transition | Td2 | M. cerifera | 3.68 | 0.047 | 98 | 0.046 | 0.023 |
| Transition | Td2 | M. cerifera | 0.97 | 0.012 | 98 | 0.012 | 0.006 |
| Transition | Td2 | M. cerifera | 0.97 | 0.012 | 98 | 0.012 | 0.006 |
| Transition | Td2 | M. cerifera | 0.97 | 0.012 | 98 | 0.012 | 0.006 |
| Transition | Td2 | M. cerifera | 0.97 | 0.012 | 98 | 0.012 | 0.006 |
| Transition | Td2 | M. cerifera | 1.88 | 0.024 | 98 | 0.023 | 0.012 |
| Transition | Td2 | M. cerifera | 1.88 | 0.024 | 98 | 0.023 | 0.012 |
| Transition | Td2 | M. cerifera | 1.88 | 0.024 | 98 | 0.023 | 0.012 |
| Transition | Td2 | M. cerifera | 1.88 | 0.024 | 98 | 0.023 | 0.012 |
| Transition | Td2 | M. cerifera | 2.78 | 0.035 | 98 | 0.035 | 0.017 |
| Transition | Td2 | M. cerifera | 7.30 | 0.093 | 98 | 0.091 | 0.046 |
| Transition | Td2 | B. halimifolia | 1.93 | 0.025 | 98 | 0.024 | 0.012 |
| Total |  |  | 42.17 | 0.54 |  | 0.53 | 0.26 |
| Transition | Td9 | M. cerifera | 1.88 | 0.024 | 98 | 0.023 | 0.012 |
| Transition | Td9 | M. cerifera | 0.97 | 0.012 | 98 | 0.012 | 0.006 |
| Transition | Td9 | M. cerifera | 2.78 | 0.035 | 98 | 0.035 | 0.017 |
| Transition | Td9 | M. cerifera | 0.97 | 0.012 | 98 | 0.012 | 0.006 |
| Transition | Td9 | M. cerifera | 0.97 | 0.012 | 98 | 0.012 | 0.006 |
| Transition | Td9 | M. cerifera | 3.68 | 0.047 | 98 | 0.046 | 0.023 |
| Transition | Td9 | M. cerifera | 4.14 | 0.053 | 98 | 0.052 | 0.026 |
| Transition | Td9 | M. cerifera | 5.49 | 0.070 | 98 | 0.069 | 0.034 |
| Transition | Td9 | M. cerifera | 0.97 | 0.012 | 98 | 0.012 | 0.006 |
| Transition | Td9 | M. cerifera | 3.68 | 0.047 | 98 | 0.046 | 0.023 |
| Transition | Td9 | M. cerifera | 0.97 | 0.012 | 98 | 0.012 | 0.006 |
| Transition | Td9 | M. cerifera | 0.97 | 0.012 | 98 | 0.012 | 0.006 |
| Transition | Td9 | M. cerifera | 2.78 | 0.035 | 98 | 0.035 | 0.017 |
| Transition | Td9 | M. cerifera | 1.88 | 0.024 | 98 | 0.023 | 0.012 |
| Transition | Td9 | M. cerifera | 2.78 | 0.035 | 98 | 0.035 | 0.017 |
| Transition | Td9 | M. cerifera | 0.97 | 0.012 | 98 | 0.012 | 0.006 |
| Transition | Td9 | M. cerifera | 0.97 | 0.012 | 98 | 0.012 | 0.006 |
| Transition | Td9 | M. cerifera | 2.78 | 0.035 | 98 | 0.035 | 0.017 |
| Transition | Td9 | M. cerifera | 2.88 | 0.037 | 98 | 0.036 | 0.018 |
| Total |  |  | 42.54 | 0.54 |  | 0.53 | 0.27 |


| Zone | Site | Species | Biomass <br> kg | Biomass <br> $\mathrm{kg} / \mathrm{m}^{2}$ | Organic <br> Matter | AFDW <br> $\mathrm{kg} / \mathrm{m}^{2}$ | Organic <br> carbon <br> $\mathrm{kg} / \mathrm{m}^{2}$ |
| :--- | :--- | :--- | ---: | :--- | ---: | ---: | ---: |
| Transition | Td15 | M. cerifera | 4.14 | 0.05 | 98 | 0.05 | 0.03 |
| Transition |  |  |  |  |  |  |  |
|  | Td33 | N/A | 0.00 | 0.000 |  | 0.00 | 0.000 |
| Transition | Td39 | M. cerifera | 2.88 | 0.037 | 98 | 0.036 | 0.018 |
| Transition | Td39 | M. cerifera | 0.97 | 0.012 | 98 | 0.012 | 0.006 |
| Transition | Td39 | M. cerifera | 2.78 | 0.035 | 98 | 0.035 | 0.017 |
| Transition | Td39 | M. cerifera | 2.78 | 0.035 | 98 | 0.035 | 0.017 |
| Transition | Td39 | M. cerifera | 0.97 | 0.012 | 98 | 0.012 | 0.006 |
| Total |  |  | 10.39 | 0.13 |  | 0.12 | 0.06 |
|  |  |  |  |  |  |  |  |
| Transition | Td50 | M. cerifera | 0.97 | 0.012 | 98 | 0.012 | 0.006 |
| Transition | Td50 | M. cerifera | 1.88 | 0.024 | 98 | 0.023 | 0.012 |
| Transition | Td50 | M. cerifera | 0.97 | 0.012 | 98 | 0.012 | 0.006 |
| Transition | Td50 | M. cerifera | 2.62 | 0.033 | 98 | 0.033 | 0.016 |
| Transition | Td50 | M. cerifera | 1.88 | 0.024 | 98 | 0.023 | 0.012 |
| Transition | Td50 | M. cerifera | 1.88 | 0.024 | 98 | 0.023 | 0.012 |
| Total |  |  | 10.19 | 0.13 |  | 0.13 | 0.06 |

APPENDIX D. HERBACEOUS VEGETATION: SUMMARY OF BIOMASS, ASH-FREE DRY WEIGHT, AND ORGANIC CARBON (KG/M²) PER SITE.

| Zone | Site | Biomass | AFDW | Organic C |
| :---: | :---: | :---: | :---: | :---: |
| Forest | Fa18 | 0.09 | 0.09 | 0.05 |
| Forest | Fa30 | 0.01 | 0.01 | 0.00 |
| Forest | Fb29 | 0.03 | 0.03 | 0.02 |
| Forest | Fb47 | 0.01 | 0.01 | 0.00 |
| Forest | Fb65 | 0.05 | 0.05 | 0.02 |
| Transition | Ta1 | 1.09 | 1.00 | 0.50 |
| Transition | Ta4 | 1.21 | 1.11 | 0.56 |
| Transition | Ta13 | 0.90 | 0.83 | 0.41 |
| Transition | Tb29 | 1.56 | 1.43 | 0.72 |
| Transition | Tc2 | 0.97 | 0.90 | 0.45 |
| Transition | Tc8 | 0.61 | 0.56 | 0.28 |
| Transition | Tc18 | 1.14 | 1.04 | 0.52 |
| Transition | Tc31 | 0.54 | 0.50 | 0.25 |
| Transition | Td2 | 0.80 | 0.74 | 0.37 |
| Transition | Td9 | 1.16 | 1.06 | 0.53 |
| Transition | Td15 | 0.97 | 0.90 | 0.45 |
| Transition | Td33 | 1.02 | 0.94 | 0.47 |
| Transition | Td39 | 0.80 | 0.73 | 0.37 |
| Transition | Td50 | 1.61 | 1.48 | 0.74 |
| High Marsh | Ha5 | 1.26 | 1.16 | 0.58 |
| High Marsh | Ha14 | 1.24 | 1.14 | 0.57 |
| High Marsh | Ha19 | 0.71 | 0.66 | 0.33 |
| High Marsh | Ha21 | 0.49 | 0.45 | 0.22 |
| High Marsh | Ha36 | 1.03 | 0.95 | 0.47 |
| High Marsh | Ha41 | 0.92 | 0.85 | 0.42 |
| High Marsh | Ha52 | 0.45 | 0.41 | 0.21 |
| High Marsh | Hb1 | 0.85 | 0.78 | 0.39 |
| High Marsh | Hb4 | 1.06 | 0.97 | 0.49 |
| High Marsh | Hc1 | 1.43 | 1.32 | 0.66 |
| High Marsh | Hc30 | 2.56 | 2.35 | 1.18 |
| High Marsh | Hc37 | 0.05 | 0.04 | 0.02 |
| High Marsh | Hc41 | 0.46 | 0.42 | 0.21 |
| High Marsh | Hc 42 | 2.74 | 2.52 | 1.26 |
| High Marsh | Hc 48 | 1.54 | 1.38 | 0.69 |
| High Marsh | Hc72 | 1.05 | 0.97 | 0.48 |
| High Marsh | Hc75 | 1.83 | 1.64 | 0.82 |
| High Marsh | Hc84 | 0.69 | 0.64 | 0.32 |
| High Marsh | Hc 92 | 0.66 | 0.60 | 0.30 |
| High Marsh | Ma1 | 2.71 | 2.49 | 1.25 |
| High Marsh | Ma15 | 1.53 | 1.41 | 0.70 |


| Zone | Site | Biomass | AFDW | Organic C |
| :--- | :--- | :---: | :---: | :---: |
| High Marsh | Mc2 | 0.47 | 0.43 | 0.22 |
| High Marsh | Tb4 | 0.95 | 0.88 | 0.44 |
| High Marsh | Tb14 | 0.81 | 0.75 | 0.37 |
| High Marsh | Tb23 | 1.30 | 1.20 | 0.60 |
| LMSS | LSa8 | 0.36 | 0.30 | 0.15 |
| LMSS | LSa12 | 0.73 | 0.58 | 0.29 |
| LMSS | LSb2 | 0.57 | 0.45 | 0.22 |
| LMSS | LSb16 | 0.78 | 0.68 | 0.34 |
| LMSS | LSb17 | 0.87 | 0.69 | 0.34 |
| LMSS | LSb46 | 0.60 | 0.50 | 0.25 |
| LMSS | LSb53 | 0.58 | 0.45 | 0.23 |
| LMSS | LSb55 | 0.70 | 0.54 | 0.27 |
| LMSS | Mb4 | 0.42 | 0.34 | 0.17 |
| LMTS | LTS6 | 0.82 | 0.67 | 0.33 |
| LMTS | LTS9 | 1.06 | 0.85 | 0.43 |
| LMTS | LT6 | 1.66 | 1.33 | 0.67 |
| LMTS | LT12 | 1.48 | 1.22 | 0.61 |
| LMTS | LT25 | 1.12 | 0.92 | 0.46 |
| LMTS | LT29 | 0.58 | 0.44 | 0.22 |
| LMTS | LT33 | 1.23 | 0.96 | 0.48 |

APPENDIX E. QUALITATIVE CHARACTERISTICS - DAUBENMIRE'S PERCENT GROUNDCOVER, DOMINANT HERBACEOUS SPECIES, DEPTH OF SOIL ORGANIC MATTER, AND WOODY VINES.

| Zone | Site | Daub. \% Herb.Quadrat ground coverspecies |  |  | Soil organic depth cm | Woody vines Species \#/ha |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Forest | Fa18 | 1 | 0 | none | 5 | R. radicans | 2737 |
| Forest | Fa18 | 2 | 15 | Smilax spp. | 4 | Campsis radicans | 64 |
| Forest | Fa18 | 3 | 15 | Smilax spp. | 3 |  |  |
| Forest | Fa18 | 4 | 15 | Smilax spp. | 3 |  |  |
| Forest | Fa18 | 5 | 37.5 | Smilax spp. | 6 |  |  |
| Forest | Fa30 | 1 | 2.5 | Rhus radicans | 6 | R. radicans | 2642 |
| Forest | Fa30 | 2 | 2.5 | R. radicans | 5 | C. radicans | 382 |
| Forest | Fa30 | 3 | 2.5 | R. radicans | 4 |  |  |
| Forest | Fa30 | 4 | 2.5 | R. radicans | 3 |  |  |
| Forest | Fa30 | 5 | 2.5 | R. radicans | 5 |  |  |
| Forest | Fb29 | 1 | 15 | R. radicans Smilax spp. | 8 | R. radicans | 891 |
| Forest | Fb29 | 2 | 2.5 | R. radicans | 5 | C. radicans | 32 |
| Forest | Fb29 | 3 | 2.5 | R. radicans | 5 |  |  |
| Forest | Fb29 | 4 | 2.5 | R. radicans | 4 |  |  |
| Forest | Fb29 | 5 | 0 | none | 3 |  |  |
| Forest | Fb47 | 1 | 2.5 | Mitchella repens | 7 | R. radicans | 318 |
| Forest | Fb47 | 2 | 15 | Smilax spp. Mitchella | 6 | C. radicans | 95 |
| Forest | Fb47 | 3 | 2.5 | repens | 8 |  |  |
| Forest | Fb47 | 4 | 15 | M. repens/ Smilax spp. | 6 |  |  |
| Forest | Fb47 | 5 | 0 | none | 7 |  |  |


| Zone | Site | Quadr | Daub. \% round cov | Dominant Herb. rspecies | Soil organic depth cm | Woody Species | vines \# / ha |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Forest | Fb65 | 1 | 15 | Panicum virgatum | 5 | R. radicans | 318 |
| Forest | Fb65 | 2 | 15 | Smilax spp. | 7 | C. radicans | 64 |
| Forest | Fb65 | 3 | 15 | Smilax spp. | 7 |  |  |
| Forest | Fb65 | 4 | 0 | none | 4 |  |  |
| Forest | Fb65 | 5 | 15 | Smilax spp. | 6 |  |  |
| Transition | Ta1 | 1 | 67.5 | S. patens/ D. spicata | 30 | R. radicans | 64 |
| Transition | Ta1 | 2 | 67.5 | D. spicata | 30 |  |  |
| Transition | Ta1 | 3 | 85 | S. patens | 20 |  |  |
| Transition | Ta1 | 4 | 100 | S. patens | 20 |  |  |
| Transition | Ta1 | 5 | 100 | S. patens | unknown |  |  |
| Transition | Ta13 | 1 | 97.5 | S. patens | 20 | none |  |
| Transition | Ta13 | 2 | 97.5 | D. spicata | 20 |  |  |
| Transition | Ta13 | 3 | 97.5 | S. patens | 20 |  |  |
| Transition | Ta13 | 4 | 97.5 | S. patens/ D. spicata | 20 |  |  |
| Transition | Ta13 | 5 | 100 | S. patens/ D. spicata | 15 |  |  |
| Transition | Ta4 | 1 | 67.5 | D. spicata | 20 | R. radicans | 32 |
| Transition | Ta4 | 2 | 67.5 | D. spicata | 20 |  |  |
| Transition | Ta4 | 3 | 85 | D. spicata | 20 |  |  |
| Transition | Ta4 | 4 | 85 | D. spicata | 10 |  |  |
| Transition | Ta4 | 5 | 97.5 | S. patens | 10 |  |  |
| Transition | Tb29 | 1 | 97.5 | S. patens | 20 | none |  |


| Zone | Site | Quadr | Daub. \% ound cove | Dominant Herb. species | Soil organic depth cm | Woody vines Species \#/h |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Transition | Tb29 | 2 | 97.5 | S. patens/ D. spicata | 15 |  |  |
| Transition | Tb29 | 3 | 97.5 | S. patens/ D. spicata | 7 |  |  |
| Transition | Tc2 | 1 | 67.5 | S. patens | 20 | none |  |
| Transition | Tc2 | 2 | 67.5 | Juncus roemerianus | 20 |  |  |
| Transition | Tc2 | 3 | 67.5 | D. spicata | 15 |  |  |
| Transition | Tc2 | 4 | 67.5 | D. spicata | 20 |  |  |
| Transition | Tc2 | 5 | 100 | S. patens | 20 |  |  |
| Transition | Tc8 | 1 | 97.5 | S. patens/D. spicata/ $P$. virgatum | 20 | R. radicans | 286 |
| Transition | Tc8 | 2 | 100 | S. patens | 20 |  |  |
| Transition | Tc8 | 3 | 100 | Juncus/ S. patens | 10 |  |  |
| Transition | Tc8 | 4 | 100 | Panicum/S. patens | 20 |  |  |
| Transition | Tc8 | 5 | 85 | S. patens | 15 |  |  |
| Transition | Tc18 | 1 | 67.5 | Juncus | 15 | none |  |
| Transition | Tc18 | 2 | 37.5 | Juncus | 10 |  |  |
| Transition | Tc18 | 3 | 85 | S. patens | 10 |  |  |
| Transition | Tc18 | 4 | 67.5 | D. spicata | 8 |  |  |
| Transition | Tc18 | 5 | 85 | D. spicata | 15 |  |  |
| Transition | Tc31 | 1 | 97.5 | S. patens/ D. spicata | 20 | R. radicans | 32 |
| Transition | Tc31 | 2 | 67.5 | D. spicata | 20 |  |  |
| Transition | Tc31 | 3 | 85 | D. spicata/ Scirpus spp. | 15 |  |  |


| Zone | Site | Quadr | Daub. \% round cov | Dominant Herb. rspecies | Soil organic depth cm | Woody vines Species \#/ ha |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Transition | Tc31 | 4 | 97.5 | S. patens/ D. spicata | 20 |  |  |
| Transition | Tc31 | 5 | 85 | S. patens | 15 |  |  |
| Transition | Td15 | 1 | 97.5 | D. spicata | 18 | R. radicans | 64 |
| Transition | Td15 | 2 | 97.5 | S. patens/ D. spicata | 10 |  |  |
| Transition | Td15 | 3 | 97.5 | S. patens/D. spicata | 10 |  |  |
| Transition | Td15 | 4 | 85 | D. spicata | 10 |  |  |
| Transition | Td15 | 5 | 97.5 | D. spicata | 20 |  |  |
| Transition | Td2 | 1 | 100 | D. spicata | 15 | R. radicans | 668 |
| Transition | Td2 | 2 | 97.5 | D. spicata | 19 |  |  |
| Transition | Td2 | 3 | 97.5 | S. patens | 17 |  |  |
| Transition | Td2 | 4 | 100 | Juncus | 10 |  |  |
| Transition | Td2 | 5 | 100 | S. patens | 15 |  |  |
| Transition | Td33 | 1 | 67.5 | D. spicata | 30 | R. radicans | 32 |
| Transition | Td33 | 2 | 67.5 | D. spicata | 25 |  |  |
| Transition | Td33 | 3 | 85 | D. spicata | 25 |  |  |
| Transition | Td33 | 4 | 85 | D. spicata | 20 |  |  |
| Transition | Td33 | 5 | 100 | S. patens | 15 |  |  |
| Transition | Td39 | 1 | 85 | D. spicata | 25 | R. radicans | 95 |
| Transition | Td39 | 2 | 100 | S. patens | 20 |  |  |
| Transition | Td39 | 3 | 100 | S. patens/D. spicata | 10 |  |  |

$\left.\begin{array}{llllllll}\hline & & & & \text { Daub. \% } \begin{array}{l}\text { Derbinant }\end{array} & \begin{array}{c}\text { Soil } \\ \text { organic } \\ \text { depth } \mathrm{cm}\end{array} & \begin{array}{c}\text { Woody vines } \\ \text { Species }\end{array} \\ \text { \#/ ha }\end{array}\right]$

| Zone | Site | Daub. \% ground cover | Dominant Herb. species | Soil organic depth cm |
| :---: | :---: | :---: | :---: | :---: |
| High marsh | Ha41 | 97.5 | S. patens/D. spicata | 5 |
| High marsh | Ha52 | 85.0 | D. spicata | 7 |
| High marsh | Hb1 | 100.0 | S. patens/D. spicata | 8 |
| High marsh | Hb4 | 100.0 | S. patens/D. spicata | 16 |
| High marsh | Hc 1 | 100.0 | Juncus/D. spicata | 33 |
| High marsh | Hc30 | 100.0 | Juncus | 33 |
| High marsh | Hc37 | 37.5 | D. spicata | 35 |
| High marsh | Hc 41 | 37.5 | D. spicata | 26 |
| High marsh | Hc42 | 100.0 | Juncus | 28 |
| High marsh | Hc68 | 100.0 | S. patens / D. spicata | 30 |
| High marsh | Hc72 | 67.5 | S. alterniflora | 38 |
| High marsh | Hc75 | 97.5 | D. spicata / <br> S. patens | 20 |
| High marsh | Hc84 | 100.0 | S. patens | 14 |
| High marsh | Hc92 | 100.0 | S. patens / <br> D. spicata | 57 |
| High marsh | Ma1 | 100.0 | Juncus | 22 |
| High marsh | Ma15 | 67.5 | S. patens / <br> D. spicata | 8 |
| High marsh | Mc2 | 85.0 | S. alterniflora | 70 |
| LMSS | Mb4 | 67.5 | S. alterniflora | 0 |
| LMSS | LSa8 | 37.5 | S. alterniflora | 0 |


| Zone | Site | Daub. \% ground cover | Dominant Herb. species | Soil organic depth cm |
| :---: | :---: | :---: | :---: | :---: |
| LMSS | LSa12 | 37.5 | S. alterniflora | 0 |
|  |  |  |  |  |
| LMSS | LSb2 | 67.5 | S. alterniflora | 0 |
|  |  |  |  |  |
| LMSS | LSb16 | 37.5 | D. spicata | 0 |
|  |  |  |  |  |
| LMSS | LSb17 | 67.5 |  |  |
|  |  |  | S. alterniflora | 0 |
| LMSS | LSb46 | 67.5 | S. alterniflora |  |
|  |  |  |  | 0 |
| LMSS | LSb53 | 67.5 | S. alterniflora | 0 |
| LMSS | LSb54 | 67.5 | S. alterniflora | 0 |
| LMTS | LT6 | 37.5 | S. alterniflora | 0 |
| LMTS | LT12 | 15.0 | S. alterniflora | 0 |
| LMTS | LT25 | 15.0 | S. alterniflora | 0 |
| LMTS | LT29 | 37.5 | S. alterniflora | 0 |
| LMTS | LT33 | 37.5 | S. alterniflora | 0 |
| LMTS | LTS6 | 15.0 | S. alterniflora | 0 |
| LMTS | LTS9 | 15.0 | S. alterniflora | 0 |

APPENDIX F. MACROORGANIC MATTER OVEN-DRY MASS AND ORGANIC CARBON MASS (KG/M ${ }^{2}$ ) PER SITE.

|  |  | 0 to 10 cm |  | 10 to 20 cm |  | 20 to 30 cm |  | Total 0 to 30 cm |  | 30 to 50 cm |  | Total 0 to 50 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Zone | Site | Dry | OC | Dry | OC | Dry | OC | Dry | OC | Dry | OC | Dry | OC |
| Forest | Fa18 | 1.79 | 0.77 | 0.52 | 0.24 | 0.29 | 0.13 | 2.60 | 1.13 | 0.29 | 0.13 | 2.89 | 1.26 |
| Forest | Fa30 | - | - | - | - | - | - | 2.01 | 0.91 | 0.32 | 0.14 | 2.33 | 1.04 |
| Forest | Fb29 | 3.17 | 1.43 | 0.68 | 0.25 | 0.27 | 0.12 | 4.12 | 1.80 | 1.10 | 0.47 | 5.22 | 2.27 |
| Forest | Fb47 | 2.07 | 0.92 | 0.73 | 0.32 | 0.32 | 0.14 | 3.12 | 1.38 | 0.25 | 0.10 | 3.36 | 1.48 |
| Forest | Fb65 | 3.70 | 1.72 | 0.64 | 0.29 | 0.33 | 0.14 | 4.67 | 2.16 | 0.33 | 0.14 | 5.00 | 2.30 |
| Transition | Ta1 | 3.01 | 1.38 | 0.64 | 0.23 | 0.22 | 0.08 | 3.87 | 1.69 |  |  |  |  |
| Transition | Ta4 | 3.60 | 1.59 | 0.47 | 0.19 | 0.28 | 0.10 | 4.35 | 1.87 |  |  |  |  |
| Transition | Ta13 | 3.01 | 1.37 | 0.65 | 0.27 | 0.37 | 0.15 | 4.02 | 1.79 |  |  |  |  |
| Transition | Tb29 | 2.31 | 0.99 | 1.26 | 0.48 | 0.31 | 0.11 | 3.88 | 1.58 |  |  |  |  |
| Transition | Tc2 | 3.70 | 1.66 | 0.51 | 0.22 | 0.21 | 0.09 | 4.42 | 1.97 |  |  |  |  |
| Transition | Tc8 | 2.26 | 0.97 | 0.60 | 0.22 | 0.31 | 0.13 | 3.17 | 1.32 |  |  |  |  |
| Transition | Tc18 | 2.01 | 0.91 | 0.35 | 0.15 | 0.12 | 0.05 | 2.48 | 1.12 |  |  |  |  |
| Transition | Tc31 | 2.20 | 1.00 | 0.74 | 0.30 | 0.32 | 0.10 | 3.26 | 1.40 |  |  |  |  |
| Transition | Td2 | 3.30 | 1.51 | 0.39 | 0.18 | 0.14 | 0.06 | 3.84 | 1.74 |  |  |  |  |
| Transition | Td9 | 5.19 | 2.26 | 1.19 | 0.53 | 0.22 | 0.10 | 6.60 | 2.89 |  |  |  |  |
| Transition | Td15 | 3.32 | 1.55 | 1.01 | 0.46 | 0.28 | 0.11 | 4.61 | 2.12 |  |  |  |  |
| Transition | Td33 | 3.59 | 1.60 | 1.76 | 0.78 | 1.20 | 0.49 | 6.55 | 2.87 |  |  |  |  |
| Transition | Td39 | 2.96 | 1.34 | 0.96 | 0.44 | 0.33 | 0.15 | 4.25 | 1.93 |  |  |  |  |
| Transition | Td50 | 4.10 | 1.90 | 1.42 | 0.62 | 0.58 | 0.25 | 6.10 | 2.77 |  |  |  |  |
| High Marsh | $\mathrm{Ha5}$ | 7.01 | 3.02 | 1.09 | 0.36 | 0.27 | 0.23 | 8.38 | 3.61 |  |  |  |  |
| High Marsh | Ha14 | 7.41 | 3.28 | 1.74 | 0.76 | 0.27 | 0.23 | 9.43 | 4.27 |  |  |  |  |
| High Marsh | Ha19 | 2.64 | 1.26 | 1.57 | 0.68 | 0.44 | 0.18 | 4.64 | 2.12 |  |  |  |  |
| High Marsh | Ha21 | 3.85 | 1.79 | 3.55 | 1.61 | 0.31 | 0.13 | 7.71 | 3.54 |  |  |  |  |
| High Marsh | Ha36 | 6.53 | 3.00 | 2.48 | 1.12 | 0.27 | 0.11 | 9.28 | 4.23 |  |  |  |  |
| High Marsh | Ha41 | 4.55 | 2.06 | 0.71 | 0.31 | 0.23 | 0.09 | 5.49 | 2.46 |  |  |  |  |
| High Marsh | Ha52 | 5.87 | 2.76 | 0.85 | 0.39 | 0.12 | 0.05 | 6.84 | 3.20 |  |  |  |  |
| High Marsh | Hc 1 | 5.38 | 2.44 | 2.90 | 1.21 | 1.27 | 0.58 | 9.55 | 4.23 |  |  |  |  |
| High Marsh | Hc 30 | 3.36 | 1.51 | 2.72 | 1.23 | 1.18 | 0.55 | 7.27 | 3.29 |  |  |  |  |


|  |  | 0 to 10 cm |  | 10 to 20 cm |  | 20 to 30 cm | Total 0 to 30 cm |  |  |
| :--- | :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Zone | Site | Dry | OC | Dry | OC | Dry | OC | Dry | OC |
| High Marsh | Hc37 | 3.78 | 1.58 | 2.61 | 1.15 | 1.28 | 0.54 | 7.66 | 3.27 |
| High Marsh | Hc41 | 3.24 | 1.54 | 3.64 | 1.58 | 1.85 | 0.85 | 8.72 | 3.96 |
| High Marsh | Hc42 | 5.83 | 2.70 | 4.09 | 1.88 | 1.31 | 0.57 | 11.23 | 5.15 |
| High Marsh | Hc68 | 5.86 | 2.70 | 5.00 | 2.33 | 1.19 | 0.56 | 12.06 | 5.58 |
| High Marsh | Hc72 | 4.64 | 2.10 | 3.01 | 1.37 | 0.74 | 0.28 | 8.40 | 3.75 |
| High Marsh | Hc75 | 5.71 | 2.58 | 3.40 | 1.54 | 4.39 | 1.20 | 13.49 | 5.32 |
| High Marsh | Hc84 | 5.42 | 2.52 | 2.36 | 1.05 | 2.31 | 0.81 | 10.08 | 4.38 |
| High Marsh | Hc92 | 7.50 | 3.14 | 5.53 | 2.51 | 3.53 | 1.66 | 16.56 | 7.32 |
| High Marsh | Ma1 | 4.71 | 2.16 | 2.84 | 1.30 | 1.60 | 0.67 | 9.15 | 4.13 |
| High Marsh | Ma15 | 4.06 | 1.91 | 3.67 | 1.60 | 1.18 | 0.55 | 8.91 | 4.06 |
| High Marsh | Mc2 | 5.14 | 2.32 | 4.01 | 1.76 | 4.12 | 1.77 | 13.27 | 5.85 |
| High Marsh | Tb4 | 5.52 | 2.33 | 0.25 | 0.10 | 0.14 | 0.05 | 5.92 | 2.48 |
| High Marsh | Tb14 | 2.01 | 0.88 | 0.18 | 0.07 | 0.12 | 0.03 | 2.31 | 0.98 |
| High Marsh | Tb23 | 2.95 | 1.40 | 1.17 | 0.49 | 0.35 | 0.14 | 4.47 | 2.04 |
| LMSS | LSa8 | 4.49 | 1.86 | 1.79 | 0.84 | 1.09 | 0.44 | 7.37 | 3.13 |
| LMSS | LSa12 | 2.48 | 1.07 | 1.55 | 0.55 | 0.33 | 0.08 | 4.37 | 1.69 |
| LMSS | LSb2 | 5.17 | 2.25 | 4.37 | 1.91 | 2.05 | 0.84 | 11.60 | 4.99 |
| LMSS | LSb16 | 3.77 | 1.58 | 0.89 | 0.30 | 0.23 | 0.08 | 4.89 | 1.96 |
| LMSS | LSb17 | 4.66 | 2.11 | 2.89 | 1.04 | 0.52 | 0.16 | 8.07 | 3.31 |
| LMSS | LSb46 | 7.19 | 3.20 | 3.54 | 1.64 | 1.74 | 0.80 | 12.47 | 5.64 |
| LMSS | LSb53 | 3.36 | 1.43 | 2.72 | 1.14 | 1.18 | 0.43 | 7.27 | 3.00 |
| LMSS | LSb54 | 4.53 | 1.96 | 3.05 | 1.35 | 1.54 | 0.68 | 9.12 | 3.98 |
| LMSS | Mb4 | 5.30 | 2.24 | 3.73 | 1.54 | 1.05 | 0.41 | 10.08 | 4.19 |
| LMTS | LT6 | 1.75 | 0.68 | 1.66 | 0.65 | 1.28 | 0.45 | 4.69 | 1.78 |
| LMTS | LT12 | 2.48 | 1.05 | 1.55 | 0.67 | 1.30 | 0.52 | 5.34 | 2.23 |
| LMTS | LT25 | 2.92 | 1.19 | 2.69 | 1.16 | 2.37 | 1.05 | 7.98 | 3.40 |
| LMTS | LT29 | 2.03 | 0.79 | 1.38 | 0.54 | 0.72 | 0.25 | 4.12 | 1.58 |
| LMTS | LT33 | 1.23 | 0.56 | 1.20 | 0.52 | 1.09 | 0.44 | 3.52 | 1.52 |
| LMTS | LTS6 | 3.49 | 1.43 | 3.76 | 1.52 | 0.82 | 0.31 | 8.06 | 3.25 |
| LMTS | LTS9 | 4.12 | 1.81 | 2.67 | 1.13 | 1.30 | 0.52 | 8.10 | 3.46 |
|  |  |  |  |  |  |  |  |  |  |

APPENDIX G. BELOWGROUND TOTAL ORGANIC MATTER AND TOTAL ORGANIC CARBON IN 10 cm INCREMENTS FOR ALL SITES.


| Zone | Site | Quadrat | Depth cm | Bulk Density $\mathrm{g} / \mathrm{cm}^{3}$ | \% Organic Matter | $\begin{array}{r} \mathrm{TOM} \\ \mathrm{~kg} / \mathrm{m}^{2} \\ \hline \hline \end{array}$ | $\begin{array}{r} \mathrm{TOC} \\ \mathrm{~kg} / \mathrm{m}^{2} \\ \hline \hline \end{array}$ | Mean TOM $\mathrm{kg} / \mathrm{m}^{2}$ | Mean TOC $\mathrm{kg} / \mathrm{m}^{2}$ | P. taeda root OC $\mathrm{kg} / \mathrm{m}^{2}$ | Fine root OC kg/m ${ }^{2}$ <br> @ 30 to 50 cm |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Forest | Fa30 | 4 | 10 to 20 | 1.381 | 5.29 | 7.30 | 3.65 |  |  |  |  |
| Forest | Fa30 | 5 | 10 to 20 | 1.121 | 2.69 | 3.02 | 1.51 | 5.28 | 2.64 |  |  |
| Forest | Fa30 | 1 | 20 to 30 | 1.510 | 4.70 | 7.11 | 3.55 |  |  |  |  |
| Forest | Fa30 | 2 | 20 to 30 | 1.321 | 5.05 | 6.67 | 3.33 |  |  |  |  |
| Forest | Fa30 | 3 | 20 to 30 | 1.448 | 3.10 | 4.49 | 2.25 |  |  |  |  |
| Forest | Fa30 | 4 | 20 to 30 | 1.398 | 4.26 | 5.95 | 2.97 |  |  |  |  |
| Forest | Fa30 | 5 | 20 to 30 | 1.348 | 2.08 | 2.81 | 1.40 | 5.40 | 2.70 | 4.00 | 0.14 |
| Total 0 to 30 |  |  |  |  |  |  |  | 23.19 | 11.59 |  |  |
| Site Total TOC |  |  |  |  |  |  |  |  |  |  | 15.74 |
| Forest | Fb29 | 1 | 0 to 10 | 0.384 | 39.58 | 15.20 | 7.60 |  |  |  |  |
| Forest | Fb29 | 2 | 0 to 10 | 0.348 | 23.67 | 8.24 | 4.12 |  |  |  |  |
| Forest | Fb29 | 3 | 0 to 10 | 0.085 | 66.18 | 5.62 | 2.81 |  |  |  |  |
| Forest | Fb29 | 4 | 0 to 10 | 0.524 | 15.40 | 8.07 | 4.04 | 9.29 | 4.64 |  |  |
| Forest | Fb29 | 1 | 10 to 20 | 0.736 | 5.35 | 3.93 | 1.97 |  |  |  |  |
| Forest | Fb29 | 2 | 10 to 20 | 0.712 | 3.99 | 2.84 | 1.42 |  |  |  |  |
| Forest | Fb29 | 3 | 10 to 20 | 0.266 | 14.44 | 3.83 | 1.92 |  |  |  |  |
| Forest | Fb29 | 4 | 10 to 20 | 0.692 | 5.31 | 3.67 | 1.84 |  |  |  |  |
| Forest | Fb29 | 5 | 10 to 20 | 1.148 | 2.49 | 2.86 | 1.43 | 3.43 | 1.71 |  |  |
| Forest | Fb29 | 1 | 20 to 30 | 1.456 | 2.79 | 4.06 | 2.03 |  |  |  |  |
| Forest | Fb29 | 2 | 20 to 30 | 1.036 | 4.73 | 4.90 | 2.45 |  |  |  |  |
| Forest | Fb29 | 4 | 20 to 30 | 1.094 | 5.28 | 5.78 | 2.89 |  |  |  |  |
| Forest | Fb29 | 5 | 20 to 30 | 1.029 | 2.39 | 2.46 | 1.23 | 4.30 | 2.15 | 1.13 | 0.47 |
| Total 0 to 30 cm |  |  |  |  |  |  |  | 17.01 | 8.51 |  |  |
| Site Total TOC |  |  |  |  |  |  |  |  |  |  | 10.11 |


| Zone | Site | Quadrat | Depth cm | Bulk Density $\mathrm{g} / \mathrm{cm}^{3}$ | \% Organic Matter | $\begin{array}{r} \mathrm{TOM} \\ \mathrm{~kg} / \mathrm{m}^{2} \\ \hline \end{array}$ | $\begin{array}{r} \mathrm{TOC} \\ \mathrm{~kg} / \mathrm{m}^{2} \\ \hline \end{array}$ | Mean TOM $\mathrm{kg} / \mathrm{m}^{2}$ | Mean TOC $\mathrm{kg} / \mathrm{m}^{2}$ | P. taeda root OC $\mathrm{kg} / \mathrm{m}^{2}$ | Fine root $\mathrm{OC} \mathrm{kg} / \mathrm{m}^{2}$ <br> @ 30 to <br> 50 cm |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Forest | Fb47 | 1 | 0 to 10 | 0.567 | 15.57 | 8.82 | 4.41 |  |  |  |  |
| Forest | Fb47 | 2 | 0 to 10 | 0.395 | 20.92 | 8.27 | 4.13 |  |  |  |  |
| Forest | Fb47 | 3 | 0 to 10 | 0.301 | 43.38 | 13.06 | 6.53 |  |  |  |  |
| Forest | Fb47 | 4 | 0 to 10 | 0.421 | 24.43 | 10.29 | 5.14 |  |  |  |  |
| Forest | Fb47 | 5 | 0 to 10 | 0.652 | 12.99 | 8.48 | 4.24 | 9.78 | 4.89 |  |  |
| Forest | Fb47 | 1 | 10 to 20 | 0.692 | 9.07 | 6.28 | 3.14 |  |  |  |  |
| Forest | Fb47 | 2 | 10 to 20 | 0.715 | 5.96 | 4.26 | 2.13 |  |  |  |  |
| Forest | Fb47 | 3 | 10 to 20 | 0.812 | 6.68 | 5.42 | 2.71 |  |  |  |  |
| Forest | Fb47 | 4 | 10 to 20 | 0.688 | 7.69 | 5.29 | 2.64 |  |  |  |  |
| Forest | Fb47 | 5 | 10 to 20 | 0.760 | 6.91 | 5.25 | 2.63 | 5.30 | 2.65 |  |  |
| Forest | Fb47 | 1 | 20 to 30 | 1.103 | 6.15 | 6.78 | 3.39 |  |  |  |  |
| Forest | Fb47 | 2 | 20 to 30 | 0.842 | 3.95 | 3.33 | 1.66 |  |  |  |  |
| Forest | Fb47 | 3 | 20 to 30 | 0.870 | 4.56 | 3.96 | 1.98 |  |  |  |  |
| Forest | Fb47 | 4 | 20 to 30 | 0.772 | 6.19 | 4.78 | 2.39 |  |  |  |  |
| Forest | Fb47 | 5 | 20 to 30 | 0.801 | 4.27 | 3.42 | 1.71 | 4.46 | 2.23 | 2.33 | 0.10 |
| Total 0 to 30 cm |  |  |  |  |  |  |  | 19.54 | 9.77 |  |  |
| Site Total TOC |  |  |  |  |  |  |  |  |  |  | 12.20 |
| Forest | Fb65 | 1 | 0 to 10 | 0.325 | 45.29 | 14.70 | 7.35 |  |  |  |  |
| Forest | Fb65 | 2 | 0 to 10 | 0.241 | 69.13 | 16.68 | 8.34 |  |  |  |  |
| Forest | Fb65 | 3 | 0 to 10 | 0.186 | 49.94 | 9.27 | 4.63 |  |  |  |  |
| Forest | Fb65 | 4 | 0 to 10 | 0.384 | 26.45 | 10.15 | 5.08 |  |  |  |  |
| Forest | Fb65 | 5 | 0 to 10 | 0.487 | 15.86 | 7.73 | 3.86 | 11.70 | 5.85 |  |  |
| Forest | Fb65 | 1 | 10 to 20 | 0.673 | 4.98 | 3.35 | 1.68 |  |  |  |  |
| Forest | Fb65 | 2 | 10 to 20 | 1.119 | 4.08 | 4.56 | 2.28 |  |  |  |  |
| Forest | Fb65 | 4 | 10 to 20 | 0.705 | 5.27 | 3.72 | 1.86 |  |  |  |  |
| Forest | Fb65 | 5 | 10 to 20 | 0.798 | 4.08 | 3.25 | 1.63 | 3.72 | 1.86 |  |  |


| Zone | Site | Quadrat | Depth cm | Bulk Density $\mathrm{g} / \mathrm{cm}$ | \% Organic Matter | TOM $\mathrm{kg} / \mathrm{m}^{2}$ | $\begin{array}{r} \mathrm{TOC} \\ \mathrm{~kg} / \mathrm{m}^{2} \\ \hline \hline \end{array}$ | Mean TOM $\mathrm{kg} / \mathrm{m}^{2}$ | $\begin{array}{r} \text { Mean } \\ \text { TOC } \\ \mathrm{kg} / \mathrm{m}^{2} \\ \hline \end{array}$ | P. taeda root OC $\mathrm{kg} / \mathrm{m}^{2}$ | Fine root OC kg/m² <br> @ 30 to <br> 50 cm |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Forest | Fb65 | 1 | 20 to 30 | 0.801 | 3.84 | 3.08 | 1.54 |  |  |  |  |
| Forest | Fb65 | 2 | 20 to 30 | 1.084 | 2.45 | 2.65 | 1.33 |  |  |  |  |
| Forest | Fb65 | 4 | 20 to 30 | 0.851 | 4.42 | 3.77 | 1.88 |  |  |  |  |
| Forest | Fb65 | 5 | 20 to 30 | 1.157 | 2.81 | 3.25 | 1.63 | 3.19 | 1.59 | 2.61 | 0.14 |
| Total 0 to 30 cm |  |  |  |  |  |  |  | 18.61 | 9.31 |  |  |
|  |  |  |  |  |  |  |  |  |  |  | 12.06 |
| Transition | Ta1 | 1 | 0 to 10 | 0.160 | 66.54 | 10.66 | 5.33 |  |  |  |  |
| Transition | Ta1 | 2 | 0 to 10 | 0.130 | 73.54 | 9.52 | 4.76 |  |  |  |  |
| Transition | Ta1 | 4 | 0 to 10 | 0.176 | 60.79 | 10.69 | 5.35 |  |  |  |  |
| Transition | Ta1 | 5 | 0 to 10 | 0.327 | 23.08 | 7.55 | 3.78 | 9.61 | 4.80 |  |  |
| Transition | Ta1 | 1 | 10 to 20 | 0.425 | 25.90 | 11.00 | 5.50 |  |  |  |  |
| Transition | Ta1 | 2 | 10 to 20 | 0.536 | 19.60 | 10.51 | 5.26 |  |  |  |  |
| Transition | Ta1 | 4 | 10 to 20 | 0.508 | 17.68 | 8.97 | 4.49 |  |  |  |  |
| Transition | Ta1 | 5 | 10 to 20 | 1.157 | 4.13 | 4.78 | 2.39 | 8.82 | 4.41 |  |  |
| Transition | Ta1 | 1 | 20 to 30 | 0.907 | 6.92 | 6.27 | 3.14 |  |  |  |  |
| Transition | Ta1 | 2 | 20 to 30 | 0.891 | 8.48 | 7.55 | 3.78 |  |  |  |  |
| Transition | Ta1 | 3 | 20 to 30 | 0.938 | 4.60 | 4.32 | 2.16 |  |  |  |  |
| Transition | Ta1 | 4 | 20 to 30 | 1.013 | 4.88 | 4.94 | 2.47 |  |  |  |  |
| Transition | Ta1 | 5 | 20 to 30 | 1.356 | 1.53 | 2.08 | 1.04 | 5.03 | 2.52 | 0.00 | N/A |
| Total 0 to 30 cm |  |  |  |  |  |  |  | 23.46 | 11.73 |  |  |
| Site Total TOC |  |  |  |  |  |  |  |  |  | 11.73 |  |
| Transition | Ta4 | 1 | 0 to 10 | 0.192 | 48.60 | 9.33 | 4.67 |  |  |  |  |
| Transition | Ta4 | 2 | 0 to 10 | 0.351 | 28.16 | 9.89 | 4.94 |  |  |  |  |
| Transition | Ta4 | 3 | 0 to 10 | 0.172 | 41.91 | 7.21 | 3.60 |  |  |  |  |
| Transition | Ta4 | 4 | 0 to 10 | 0.129 | 68.72 | 8.89 | 4.45 |  |  |  |  |


| Zone | Site | Quadrat | Depth cm | Bulk Density $\mathrm{g} / \mathrm{cm}$ | \% Organic Matter | $\begin{array}{r} \mathrm{TOM} \\ \mathrm{~kg} / \mathrm{m}^{2} \\ \hline \end{array}$ | $\begin{array}{r} \mathrm{TOC} \\ \mathrm{~kg} / \mathrm{m}^{2} \\ \hline \hline \end{array}$ | Mean TOM $\mathrm{kg} / \mathrm{m}^{2}$ |  | P. taeda root OC $\mathrm{kg} / \mathrm{m}^{2}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Transition | Ta4 | 5 | 0 to 10 | 0.089 | 65.94 | 5.85 | 2.93 | 8.23 | 4.12 |  |
| Transition | Ta4 | 1 | 10 to 20 | 0.447 | 16.40 | 7.33 | 3.67 |  |  |  |
| Transition | Ta4 | 2 | 10 to 20 | 0.860 | 9.35 | 8.04 | 4.02 |  |  |  |
| Transition | Ta4 | 3 | 10 to 20 | 1.130 | 6.33 | 7.16 | 3.58 |  |  |  |
| Transition | Ta4 | 4 | 10 to 20 | 1.021 | 7.56 | 7.72 | 3.86 |  |  |  |
| Transition | Ta4 | 5 | 10 to 20 | 0.277 | 22.17 | 6.14 | 3.07 | 7.28 | 3.64 |  |
| Transition | Ta4 | 1 | 20 to 30 | 0.811 | 6.30 | 5.11 | 2.56 |  |  |  |
| Transition | Ta4 | 3 | 20 to 30 | 1.594 | 2.38 | 3.80 | 1.90 |  |  |  |
| Transition | Ta4 | 4 | 20 to 30 | 1.243 | 3.07 | 3.81 | 1.91 |  |  |  |
| Transition | Ta4 | 5 | 20 to 30 | 0.833 | 11.53 | 9.61 | 4.80 | 5.58 | 2.79 | 0.002 |
| Total 0 to 30 | cm |  |  |  |  |  |  | 21.09 | 10.55 |  |
| Site total T |  |  |  |  |  |  |  |  |  | 10.55 |
| Transition | Ta13 | 1 | 0 to 10 | 0.320 | 35.23 | 11.26 | 5.63 |  |  |  |
| Transition | Ta13 | 2 | 0 to 10 | 0.340 | 32.08 | 10.90 | 5.45 |  |  |  |
| Transition | Ta13 | 3 | 0 to 10 | 0.485 | 23.31 | 11.30 | 5.65 |  |  |  |
| Transition | Ta13 | 4 | 0 to 10 | 0.473 | 24.42 | 11.56 | 5.78 |  |  |  |
| Transition | Ta13 | 5 | 0 to 10 | 0.400 | 27.26 | 10.92 | 5.46 | 11.19 | 5.59 |  |
| Transition | Ta13 | 1 | 10 to 20 | 0.964 | 8.53 | 8.23 | 4.11 |  |  |  |
| Transition | Ta13 | 2 | 10 to 20 | 1.092 | 5.80 | 6.33 | 3.16 |  |  |  |
| Transition | Ta13 | 4 | 10 to 20 | 1.160 | 5.91 | 6.86 | 3.43 |  |  |  |
| Transition | Ta13 | 5 | 10 to 20 | 1.241 | 6.27 | 7.77 | 3.89 | 7.30 | 3.65 |  |
| Transition | Ta13 | 1 | 20 to 30 | 1.211 | 3.06 | 3.71 | 1.85 |  |  |  |
| Transition | Ta13 | 2 | 20 to 30 | 1.473 | 1.99 | 2.93 | 1.47 |  |  |  |


| Zone | Site | Quadrat | Depth cm | Bulk Density $\mathrm{g} / \mathrm{cm}$ | \% Organic Matter | $\begin{array}{r} \mathrm{TOM} \\ \mathrm{~kg} / \mathrm{m}^{2} \\ \hline \hline \end{array}$ | $\begin{array}{r} \mathrm{TOC} \\ \mathrm{~kg} / \mathrm{m}^{2} \\ \hline \end{array}$ | Mean TOM $\mathrm{kg} / \mathrm{m}^{2}$ |  | P. taeda root OC $\mathrm{kg} / \mathrm{m}^{2}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Transition | Ta13 | 3 | 20 to 30 | 1.571 | 2.04 | 3.21 | 1.61 |  |  |  |
| Transition | Ta13 | 4 | 20 to 30 | 1.543 | 2.03 | 3.13 | 1.57 |  |  |  |
| Transition | Ta13 | 5 | 20 to 30 | 1.561 | 2.97 | 4.63 | 2.32 | 3.52 | 1.76 | 0.00 |
| Total 0 to 30 cm |  |  |  |  |  |  |  | 22.01 | 11.00 |  |
| Site Total TOC |  |  |  |  |  |  |  |  |  | 11.00 |
| Transition | Tb29 | 1 | 0 to 10 | 0.314 | 32.36 | 10.17 | 5.08 |  |  |  |
| Transition | Tb29 | 2 | 0 to 10 | 0.164 | 51.94 | 8.52 | 4.26 |  |  |  |
| Transition | Tb29 | 3 | 0 to 10 | 0.746 | 31.95 | 23.85 | 11.92 | 14.18 | 7.09 |  |
| Transition | Tb29 | 1 | 10 to 20 | 0.898 | 4.70 | 4.22 | 2.11 |  |  |  |
| Transition | Tb29 | 2 | 10 to 20 | 0.690 | 13.48 | 9.29 | 4.65 | 6.76 | 3.38 |  |
| Transition | Tb29 | 1 | 20 to 30 | 1.610 | 3.34 | 5.37 | 2.69 |  |  |  |
| Transition | Tb29 | 2 | 20 to 30 | 1.246 | 3.46 | 4.31 | 2.16 |  |  |  |
| Transition | Tb29 | 3 | 20 to 30 | 1.508 | 3.69 | 5.56 | 2.78 | 5.08 | 2.54 | 0.00 |
| Total 0 to 30 cm |  |  |  |  |  |  |  | 26.02 | 13.01 |  |
| Site Total TOC |  |  |  |  |  |  |  |  |  | 13.01 |
| Transition | Tc2 | 1 | 0 to 10 | 0.203 | 51.05 | 10.35 | 5.18 |  |  |  |
| Transition | Tc2 | 2 | 0 to 10 | 0.234 | 46.31 | 10.86 | 5.43 |  |  |  |
| Transition | Tc2 | 3 | 0 to 10 | 0.248 | 38.10 | 9.45 | 4.73 |  |  |  |
| Transition | Tc2 | 4 | 0 to 10 | 0.181 | 38.27 | 6.93 | 3.46 |  |  |  |
| Transition | Tc2 | 5 | 0 to 10 | 0.176 | 63.17 | 11.14 | 5.57 | 9.75 | 4.87 |  |
| Transition | Tc2 | 1 | 10 to 20 | 0.780 | 7.96 | 6.20 | 3.10 |  |  |  |
| Transition | Tc2 | 2 | 10 to 20 | 0.725 | 8.19 | 5.93 | 2.97 |  |  |  |
| Transition | Tc2 | 3 | 10 to 20 | 1.060 | 5.21 | 5.52 | 2.76 |  |  |  |
| Transition | Tc2 | 4 | 10 to 20 | 0.970 | 5.57 | 5.40 | 2.70 |  |  |  |
| Transition | Tc2 | 5 | 10 to 20 | 0.590 | 9.47 | 5.59 | 2.80 | 5.73 | 2.87 |  |
| Transition | Tc2 | 1 | 20 to 30 | 0.906 | 3.83 | 3.47 | 1.74 |  |  |  |


| Zone | Site | Quadrat | Depth cm | Bulk Density $\mathrm{g} / \mathrm{cm}^{3}$ | $\begin{gathered} \text { \% Organic } \\ \text { Matter } \\ \hline \hline \end{gathered}$ | $\begin{array}{r} \text { TOM } \\ \mathrm{kg} / \mathrm{m}^{2} \\ \hline \hline \end{array}$ | $\begin{array}{r} \mathrm{TOC} \\ \mathrm{~kg} / \mathrm{m}^{2} \\ \hline \hline \end{array}$ | $\begin{gathered} \text { Mean } \\ \text { TOM } \\ \mathrm{kg} / \mathrm{m}^{2} \end{gathered}$ | $\begin{gathered} \text { Mean } \\ \mathrm{TOC} \\ \mathrm{~kg} / \mathrm{m}^{2} \\ \hline \end{gathered}$ | P. taeda root OC $\mathrm{kg} / \mathrm{m}^{2}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Transition | Tc2 | 2 | 20 to 30 | 1.285 | 2.51 | 3.23 | 1.62 |  |  |  |
| Transition | Tc2 | 3 | 20 to 30 | 1.460 | 3.44 | 5.03 | 2.51 |  |  |  |
| Transition | Tc2 | 4 | 20 to 30 | 1.083 | 2.52 | 2.73 | 1.36 |  |  |  |
| Transition | Tc2 | 5 | 20 to 30 | 0.752 | 5.04 | 3.79 | 1.90 | 3.65 | 1.83 | 0.011 |
| Total 0 to 30 cm |  |  |  |  |  |  |  | 19.13 | 9.56 |  |
| Site Total TOC |  |  |  |  |  |  |  |  |  | 9.57 |
| Transition | Tc8 | 1 | 0 to 10 | 0.210 | 40.30 | 8.46 | 4.23 |  |  |  |
| Transition | Tc8 | 2 | 0 to 10 | 0.177 | 42.32 | 7.47 | 3.74 |  |  |  |
| Transition | Tc8 | 3 | 0 to 10 | 0.325 | 43.66 | 14.19 | 7.09 |  |  |  |
| Transition | Tc8 | 4 | 0 to 10 | 0.260 | 40.66 | 10.58 | 5.29 |  |  |  |
| Transition | Tc8 | 5 | 0 to 10 | 0.135 | 72.33 | 9.73 | 4.87 | 10.09 | 5.04 |  |
| Transition | Tc8 | 1 | 10 to 20 | 0.700 | 10.28 | 7.20 | 3.60 |  |  |  |
| Transition | Tc8 | 2 | 10 to 20 | 0.374 | 30.99 | 11.58 | 5.79 |  |  |  |
| Transition | Tc8 | 3 | 10 to 20 | 0.694 | 26.30 | 18.25 | 9.13 |  |  |  |
| Transition | Tc8 | 4 | 10 to 20 | 0.700 | 17.44 | 12.20 | 6.10 |  |  |  |
| Transition | Tc8 | 5 | 10 to 20 | 0.656 | 17.61 | 11.56 | 5.78 | 12.16 | 6.08 |  |
| Transition | Tc8 | 1 | 20 to 30 | 1.013 | 4.16 | 4.21 | 2.11 |  |  |  |
| Transition | Tc8 | 2 | 20 to 30 | 1.399 | 6.49 | 9.08 | 4.54 |  |  |  |
| Transition | Tc8 | 3 | 20 to 30 | 1.422 | 5.26 | 7.48 | 3.74 |  |  |  |
| Transition | Tc8 | 4 | 20 to 30 | 1.176 | 6.04 | 7.10 | 3.55 |  |  |  |
| Transition | Tc8 | 5 | 20 to 30 | 1.063 | 6.03 | 6.41 | 3.21 | 6.86 | 3.43 | 0.68 |
| Total 0 to 30 cm |  |  |  |  |  |  |  | 29.10 | 14.55 |  |
| Site Total TOC |  |  |  |  |  |  |  |  |  | 15.23 |
| Transition | Tc18 | 1 | 0 to 10 | 0.246 | 42.31 | 10.39 | 5.20 |  |  |  |
| Transition | Tc18 | 2 | 0 to 10 | 0.205 | 40.01 | 8.19 | 4.09 |  |  |  |
| Transition | Tc18 | 3 | 0 to 10 | 0.157 | 51.11 | 8.00 | 4.00 |  |  |  |


| Zone | Site | Quadrat | Depth cm | Bulk Density $\mathrm{g} / \mathrm{cm}^{3}$ | \% Organic Matter | $\begin{array}{r} \mathrm{TOM} \\ \mathrm{~kg} / \mathrm{m}^{2} \\ \hline \end{array}$ | $\begin{array}{r} \mathrm{TOC} \\ \mathrm{~kg} / \mathrm{m}^{2} \\ \hline \hline \end{array}$ | Mean TOM $\mathrm{kg} / \mathrm{m}^{2}$ | Mean TOC $\mathrm{kg} / \mathrm{m}^{2}$ | P. taeda root OC $\mathrm{kg} / \mathrm{m}^{2}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Transition | Tc18 | 4 | 0 to 10 | 0.581 | 11.67 | 6.78 | 3.39 |  |  |  |
| Transition | Tc18 | 5 | 0 to 10 | 0.506 | 20.07 | 10.16 | 5.08 | 8.70 | 4.35 |  |
| Transition | Tc18 | 1 | 10 to 20 | 1.167 | 6.53 | 7.62 | 3.81 |  |  |  |
| Transition | Tc18 | 2 | 10 to 20 | 0.794 | 10.61 | 8.43 | 4.21 |  |  |  |
| Transition | Tc18 | 3 | 10 to 20 | 0.926 | 9.41 | 8.71 | 4.36 |  |  |  |
| Transition | Tc18 | 4 | 10 to 20 | 1.668 | 3.39 | 5.65 | 2.83 |  |  |  |
| Transition | Tc18 | 5 | 10 to 20 | 1.165 | 4.80 | 5.59 | 2.79 | 7.20 | 3.60 |  |
| Transition | Tc18 | 1 | 20 to 30 | 1.293 | 2.50 | 3.24 | 1.62 |  |  |  |
| Transition | Tc18 | 2 | 20 to 30 | 1.100 | 5.00 | 5.50 | 2.75 |  |  |  |
| Transition | Tc18 | 3 | 20 to 30 | 1.376 | 4.68 | 6.44 | 3.22 |  |  |  |
| Transition | Tc18 | 4 | 20 to 30 | 1.599 | 3.55 | 5.67 | 2.84 |  |  |  |
| Transition | Tc18 | 5 | 20 to 30 | 1.875 | 3.18 | 5.95 | 2.98 | 5.36 | 2.68 | 0.010 |
| Total 0 to 3 | cm |  |  |  |  |  |  | 21.27 | 10.63 |  |
| Site Total |  |  |  |  |  |  |  |  |  | 10.64 |
| Transition | Tc31 | 1 | 0 to 10 | 0.216 | 48.03 | 10.37 | 5.19 |  |  |  |
| Transition | Tc31 | 2 | 0 to 10 | 0.284 | 36.90 | 10.47 | 5.24 |  |  |  |
| Transition | Tc31 | 3 | 0 to 10 | 0.286 | 27.15 | 7.78 | 3.89 |  |  |  |
| Transition | Tc31 | 4 | 0 to 10 | 0.157 | 63.55 | 9.95 | 4.97 |  |  |  |
| Transition | Tc31 | 5 | 0 to 10 | 0.258 | 39.24 | 10.12 | 5.06 | 9.74 | 4.87 |  |
| Transition | Tc31 | 1 | 10 to 20 | 0.740 | 7.20 | 5.32 | 2.66 |  |  |  |
| Transition | Tc31 | 2 | 10 to 20 | 0.803 | 6.23 | 5.00 | 2.50 |  |  |  |
| Transition | Tc31 | 3 | 10 to 20 | 1.222 | 4.91 | 6.00 | 3.00 |  |  |  |
| Transition | Tc31 | 4 | 10 to 20 | 0.633 | 16.17 | 10.24 | 5.12 |  |  |  |
| Transition | Tc31 | 5 | 10 to 20 | 0.881 | 6.96 | 6.13 | 3.06 | 6.54 | 3.27 |  |
| Transition | Tc31 | 1 | 20 to 30 | 0.884 | 2.92 | 2.58 | 1.29 |  |  |  |
| Transition | Tc31 | 2 | 20 to 30 | 1.010 | 4.14 | 4.18 | 2.09 |  |  |  |
| Transition | Tc31 | 3 | 20 to 30 | 1.462 | 3.83 | 5.60 | 2.80 |  |  |  |


| Zone | Site | Quadrat | Depth cm | Bulk Density $\mathrm{g} / \mathrm{cm}$ | \% Organic Matter | $\begin{array}{r} \mathrm{TOM} \\ \mathrm{~kg} / \mathrm{m}^{2} \\ \hline \end{array}$ | $\begin{array}{r} \mathrm{TOC} \\ \mathrm{~kg} / \mathrm{m}^{2} \\ \hline \end{array}$ | Mean TOM $\mathrm{kg} / \mathrm{m}^{2}$ | Mean TOC $\mathrm{kg} / \mathrm{m}^{2}$ | P. taeda root OC $\mathrm{kg} / \mathrm{m}^{2}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Transition | Tc31 | 4 | 20 to 30 | 1.240 | 3.96 | 4.91 | 2.45 |  |  |  |
| Transition | Tc31 | 5 | 20 to 30 | 0.961 | 4.17 | 4.01 | 2.00 | 4.26 | 2.13 | 0.00 |
| Total 0 to | cm |  |  |  |  |  |  | 20.53 | 10.27 |  |
| Site Total TOC |  |  |  |  |  |  |  |  |  | 10.27 |
| Transition | Td2 | 1 | 0 to 10 | 0.220 | 50.75 | 11.16 | 5.58 |  |  |  |
| Transition | Td2 | 2 | 0 to 10 | 0.234 | 50.26 | 11.79 | 5.89 |  |  |  |
| Transition | Td2 | 3 | 0 to 10 | 0.248 | 74.85 | 18.57 | 9.29 |  |  |  |
| Transition | Td2 | 4 | 0 to 10 | 0.181 | 38.89 | 7.04 | 3.52 |  |  |  |
| Transition | Td2 | 5 | 0 to 10 | 0.176 | 40.52 | 7.14 | 3.57 | 11.14 | 5.57 |  |
| Transition | Td2 | 1 | 10 to 20 | 0.825 | 7.84 | 6.47 | 3.24 |  |  |  |
| Transition | Td2 | 2 | 10 to 20 | 0.777 | 9.40 | 7.30 | 3.65 |  |  |  |
| Transition | Td2 | 3 | 10 to 20 | 0.383 | 19.35 | 7.41 | 3.70 |  |  |  |
| Transition | Td2 | 4 | 10 to 20 | 0.932 | 6.04 | 5.63 | 2.81 |  |  |  |
| Transition | Td2 | 5 | 10 to 20 | 0.999 | 5.99 | 5.99 | 2.99 | 6.56 | 3.28 |  |
| Transition | Td2 | 1 | 20 to 30 | 1.598 | 3.55 | 5.67 | 2.83 |  |  |  |
| Transition | Td2 | 2 | 20 to 30 | 1.215 | 2.48 | 3.01 | 1.51 |  |  |  |
| Transition | Td2 | 3 | 20 to 30 | 1.133 | 3.84 | 4.35 | 2.18 |  |  |  |
| Transition | Td2 | 4 | 20 to 30 | 1.420 | 3.34 | 4.75 | 2.37 |  |  |  |
| Transition | Td2 | 5 | 20 to 30 | 1.047 | 3.29 | 3.45 | 1.72 | 4.24 | 2.12 | 0.456 |
| Total 0 to 30 cm |  |  |  |  |  |  |  | 21.94 | 10.97 |  |
| Site Total TOC |  |  |  |  |  |  |  |  |  | 11.43 |
| Transition | Td9 | 1 | 0 to 10 | 0.106 | 72.52 | 7.68 | 3.84 |  |  |  |
| Transition | Td9 | 2 | 0 to 10 | 0.144 | 54.84 | 7.89 | 3.95 |  |  |  |
| Transition | Td9 | 3 | 0 to 10 | 0.164 | 60.88 | 10.01 | 5.00 |  |  |  |
| Transition | Td9 | 4 | 0 to 10 | 0.222 | 54.31 | 12.05 | 6.03 |  |  |  |
| Transition | Td9 | 5 | 0 to 10 | 0.093 | 71.54 | 6.68 | 3.34 | 8.86 | 4.43 |  |


| Zone | Site | Quadrat | Depth cm | Bulk Density $\mathrm{g} / \mathrm{cm}$ | \% Organic Matter | $\begin{array}{r} \mathrm{TOM} \\ \mathrm{~kg} / \mathrm{m}^{2} \\ \hline \hline \end{array}$ | $\begin{array}{r} \mathrm{TOC} \\ \mathrm{~kg} / \mathrm{m}^{2} \\ \hline \hline \end{array}$ | Mean TOM $\mathrm{kg} / \mathrm{m}^{2}$ | Mean TOC kg/m | P. taeda root OC $\mathrm{kg} / \mathrm{m}^{2}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Transition | Td9 | 1 | 10 to 20 | 0.300 | 27.06 | 8.13 | 4.06 |  |  |  |
| Transition | Td9 | 2 | 10 to 20 | 0.675 | 12.01 | 8.11 | 4.06 |  |  |  |
| Transition | Td9 | 3 | 10 to 20 | 0.708 | 7.24 | 5.13 | 2.56 |  |  |  |
| Transition | Td9 | 4 | 10 to 20 | 1.061 | 4.44 | 4.71 | 2.36 |  |  |  |
| Transition | Td9 | 5 | 10 to 20 | 0.441 | 15.84 | 6.99 | 3.49 | 6.61 | 3.31 |  |
| Transition | Td9 | 1 | 20 to 30 | 0.904 | 4.77 | 4.31 | 2.16 |  |  |  |
| Transition | Td9 | 2 | 20 to 30 | 1.196 | 2.17 | 2.59 | 1.30 |  |  |  |
| Transition | Td9 | 3 | 20 to 30 | 1.265 | 2.87 | 3.63 | 1.82 |  |  |  |
| Transition | Td9 | 4 | 20 to 30 | 1.171 | 2.85 | 3.34 | 1.67 |  |  |  |
| Transition | Td9 | 5 | 20 to 30 | 1.350 | 2.50 | 3.38 | 1.69 | 3.45 | 1.73 | 0.025 |
| Total 0 to | cm |  |  |  |  |  |  | 18.93 | 9.46 |  |
| Site Total |  |  |  |  |  |  |  |  |  | 9.49 |
| Transition | Td15 | 1 | 0 to 10 | 0.087 | 71.44 | 6.21 | 3.11 |  |  |  |
| Transition | Td15 | 2 | 0 to 10 | 0.098 | 68.96 | 6.75 | 3.37 |  |  |  |
| Transition | Td15 | 3 | 0 to 10 | 0.137 | 51.69 | 7.08 | 3.54 |  |  |  |
| Transition | Td15 | 4 | 0 to 10 | 0.304 | 32.53 | 9.90 | 4.95 |  |  |  |
| Transition | Td15 | 5 | 0 to 10 | 0.102 | 71.57 | 7.34 | 3.67 | 7.45 | 3.73 |  |
| Transition | Td15 | 1 | 10 to 20 | 0.268 | 25.11 | 6.74 | 3.37 |  |  |  |
| Transition | Td15 | 2 | 10 to 20 | 0.534 | 12.78 | 6.82 | 3.41 |  |  |  |
| Transition | Td15 | 3 | 10 to 20 | 0.696 | 5.46 | 3.80 | 1.90 |  |  |  |
| Transition | Td15 | 4 | 10 to 20 | 0.755 | 6.37 | 4.81 | 2.40 |  |  |  |
| Transition | Td15 | 5 | 10 to 20 | 0.377 | 21.71 | 8.18 | 4.09 | 6.07 | 3.04 |  |
| Transition | Td15 | 1 | 20 to 30 | 0.725 | 4.81 | 3.48 | 1.74 |  |  |  |
| Transition | Td15 | 2 | 20 to 30 | 1.081 | 3.55 | 3.84 | 1.92 |  |  |  |


| Zone | Site | Quadrat | Depth cm | Bulk Density $\mathrm{g} / \mathrm{cm}$ | \% Organic Matter | $\begin{array}{r} \mathrm{TOM} \\ \mathrm{~kg} / \mathrm{m}^{2} \\ \hline \end{array}$ | $\begin{array}{r} \mathrm{TOC} \\ \mathrm{~kg} / \mathrm{m}^{2} \\ \hline \end{array}$ | Mean TOM $\mathrm{kg} / \mathrm{m}^{2}$ | Mean TOC $\mathrm{kg} / \mathrm{m}^{2}$ | P. taeda root OC $\mathrm{kg} / \mathrm{m}^{2}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Transition | Td15 | 3 | 20 to 30 | 0.946 | 3.47 | 3.28 | 1.64 |  |  |  |
| Transition | Td15 | 4 | 20 to 30 | 1.303 | 2.73 | 3.56 | 1.78 |  |  |  |
| Transition | Td15 | 5 | 20 to 30 | 1.181 | 2.97 | 3.51 | 1.75 | 3.53 | 1.77 | 0.110 |
| Site Total TOC |  |  |  |  |  |  |  | 17.06 | 8.53 |  |
|  |  |  |  |  |  |  |  |  |  | 8.64 |
| Transition | Td33 | 1 | 0 to 10 | 0.189 | 36.96 | 6.99 | 3.49 |  |  |  |
| Transition | Td33 | 2 | 0 to 10 | 0.067 | 81.69 | 5.47 | 2.73 |  |  |  |
| Transition | Td33 | 3 | 0 to 10 | 0.106 | 67.71 | 7.17 | 3.58 |  |  |  |
| Transition | Td33 | 4 | 0 to 10 | 0.056 | 75.77 | 4.25 | 2.12 |  |  |  |
| Transition | Td33 | 5 | 0 to 10 | 0.115 | 57.69 | 6.65 | 3.32 | 6.10 | 3.05 |  |
| Transition | Td33 | 1 | 10 to 20 | 0.257 | 19.90 | 5.11 | 2.56 |  |  |  |
| Transition | Td33 | 2 | 10 to 20 | 0.092 | 65.55 | 6.01 | 3.01 |  |  |  |
| Transition | Td33 | 3 | 10 to 20 | 0.351 | 15.03 | 5.28 | 2.64 |  |  |  |
| Transition | Td33 | 4 | 10 to 20 | 0.141 | 57.46 | 8.13 | 4.06 |  |  |  |
| Transition | Td33 | 5 | 10 to 20 | 1.029 | 6.61 | 6.80 | 3.40 | 6.27 | 3.13 |  |
| Transition | Td33 | 1 | 20 to 30 | 0.786 | 4.59 | 3.61 | 1.81 |  |  |  |
| Transition | Td33 | 2 | 20 to 30 | 0.795 | 8.12 | 6.45 | 3.22 |  |  |  |
| Transition | Td33 | 3 | 20 to 30 | 1.046 | 5.45 | 5.70 | 2.85 |  |  |  |
| Transition | Td33 | 4 | 20 to 30 | 1.273 | 5.95 | 7.58 | 3.79 |  |  |  |
| Transition | Td33 | 5 | 20 to 30 | 1.590 | 2.35 | 3.74 | 1.87 | 5.42 | 2.71 | 0.027 |
| Total 0 to 30 cm |  |  |  |  |  |  |  | 17.78 | 8.89 |  |
| Site Total TOC |  |  |  |  |  |  |  |  |  | 8.92 |
| Transition | Td39 | 1 | 0 to 10 | 0.066 | 78.70 | 5.23 | 2.61 |  |  |  |
| Transition | Td39 | 2 | 0 to 10 | 0.177 | 48.21 | 8.54 | 4.27 |  |  |  |
| Transition | Td39 | 3 | 0 to 10 | 0.228 | 32.87 | 7.51 | 3.75 |  |  |  |
| Transition | Td39 | 4 | 0 to 10 | 0.339 | 31.46 | 10.67 | 5.33 |  |  |  |


| Zone | Site | Quadrat | Depth cm | Bulk Density $\mathrm{g} / \mathrm{cm}^{3}$ | \% Organic Matter | $\begin{array}{r} \mathrm{TOM} \\ \mathrm{~kg} / \mathrm{m}^{2} \\ \hline \hline \end{array}$ | $\begin{array}{r} \mathrm{TOC} \\ \mathrm{~kg} / \mathrm{m}^{2} \\ \hline \hline \end{array}$ | Mean TOM $\mathrm{kg} / \mathrm{m}^{2}$ | Mean TOC $\mathrm{kg} / \mathrm{m}^{2}$ | P. taeda root OC $\mathrm{kg} / \mathrm{m}^{2}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Transition | Td39 | 5 | 0 to 10 | 0.478 | 19.08 | 9.13 | 4.56 | 8.22 | 4.11 |  |
| Transition | Td39 | 1 | 10 to 20 | 0.245 | 43.88 | 10.73 | 5.37 |  |  |  |
| Transition | Td39 | 2 | 10 to 20 | 0.530 | 15.07 | 7.98 | 3.99 |  |  |  |
| Transition | Td39 | 3 | 10 to 20 | 0.761 | 4.38 | 3.33 | 1.67 |  |  |  |
| Transition | Td39 | 4 | 10 to 20 | 0.797 | 3.99 | 3.18 | 1.59 |  |  |  |
| Transition | Td39 | 5 | 10 to 20 | 0.881 | 4.62 | 4.07 | 2.04 | 5.86 | 2.93 |  |
| Transition | Td39 | 1 | 20 to 30 | 0.476 | 16.62 | 7.91 | 3.96 |  |  |  |
| Transition | Td39 | 2 | 20 to 30 | 0.879 | 5.13 | 4.51 | 2.26 |  |  |  |
| Transition | Td39 | 3 | 20 to 30 | 1.111 | 2.31 | 2.56 | 1.28 |  |  |  |
| Transition | Td39 | 4 | 20 to 30 | 0.970 | 4.93 | 4.78 | 2.39 |  |  |  |
| Transition | Td39 | 5 | 20 to 30 | 1.271 | 3.38 | 4.29 | 2.14 | 4.81 | 2.41 | 0.107 |
| Total 0 to 30 | cm |  |  |  |  |  |  | 18.89 | 9.44 |  |
| Site Total |  |  |  |  |  |  |  |  |  | 9.55 |
| Transition | Td50 | 1 | 0 to 10 | 0.101 | 78.34 | 7.88 | 3.94 |  |  |  |
| Transition | Td50 | 2 | 0 to 10 | 0.107 | 89.83 | 9.62 | 4.81 |  |  |  |
| Transition | Td50 | 3 | 0 to 10 | 0.090 | 67.21 | 6.06 | 3.03 |  |  |  |
| Transition | Td50 | 4 | 0 to 10 | 0.079 | 72.84 | 5.74 | 2.87 |  |  |  |
| Transition | Td50 | 5 | 0 to 10 | 0.138 | 70.81 | 9.78 | 4.89 | 7.82 | 3.91 |  |
| Transition | Td50 | 1 | 10 to 20 | 0.254 | 33.69 | 8.57 | 4.28 |  |  |  |
| Transition | Td50 | 2 | 10 to 20 | 0.404 | 23.57 | 9.53 | 4.76 |  |  |  |
| Transition | Td50 | 3 | 10 to 20 | 0.438 | 18.83 | 8.24 | 4.12 |  |  |  |
| Transition | Td50 | 4 | 10 to 20 | 0.315 | 39.67 | 12.51 | 6.26 |  |  |  |
| Transition | Td50 | 5 | 10 to 20 | 0.646 | 15.42 | 9.96 | 4.98 | 9.76 | 4.88 |  |


| Zone | Site | Quadrat | Depth cm | Bulk Density $\mathrm{g} / \mathrm{cm}$ | \% Organic Matter | TOM $\mathrm{kg} / \mathrm{m}^{2}$ | $\begin{array}{r} \mathrm{TOC} \\ \mathrm{~kg} / \mathrm{m}^{2} \\ \hline \end{array}$ | Mean TOM $\mathrm{kg} / \mathrm{m}^{2}$ | Mean TOC $\mathrm{kg} / \mathrm{m}^{2}$ | P. taeda root OC $\mathrm{kg} / \mathrm{m}^{2}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Transition | Td50 | 1 | 20 to30 | 1.093 | 7.22 | 7.89 | 3.95 |  |  |  |
| Transition | Td50 | 2 | 20 to 30 | 1.082 | 6.54 | 7.08 | 3.54 |  |  |  |
| Transition | Td50 | 3 | 20 to 30 | 1.336 | 4.24 | 5.67 | 2.84 |  |  |  |
| Transition | Td50 | 4 | 20 to 30 | 1.357 | 6.14 | 8.33 | 4.16 |  |  |  |
| Transition | Td50 | 5 | 20 to 30 | 1.513 | 3.03 | 4.59 | 2.29 | 6.71 | 3.36 | 0.017 |
| Total 0 to 30 | cm |  |  |  |  |  |  | 24.29 | 12.15 |  |
| Site Total T |  |  |  |  |  |  |  |  |  | 12.16 |
| High Marsh | $\mathrm{Ha5}$ |  | 0 to 10 | 0.350 | 44.16 | 15.47 | 7.73 |  |  |  |
| High Marsh | $\mathrm{Ha5}$ |  | 10 to 20 | 1.483 | 7.33 | 10.87 | 5.44 |  |  |  |
| High Marsh | $\mathrm{Ha5}$ |  | 20 to 30 | 1.519 | 2.22 | 3.38 | 1.69 |  |  |  |
| Total 0 to 30 | cm |  |  |  |  | 29.72 | 14.86 |  |  |  |
| High Marsh | Ha14 |  | 0 to 10 | 0.181 | 67.56 | 12.24 | 6.12 |  |  |  |
| High Marsh | Ha14 |  | 10 to 20 | 0.513 | 36.17 | 18.57 | 9.29 |  |  |  |
| High Marsh | Ha14 |  | 20 to 30 | * | * | 9.25 | 4.63 |  |  |  |
| Total 0 to 30 | cm |  |  |  |  | 40.06 | 20.03 |  |  |  |
| High Marsh | Ha19 |  | 0 to 10 | 0.048 | 71.71 | 3.41 | 1.71 |  |  |  |
| High Marsh | Ha19 |  | 10 to 20 | 0.200 | 34.04 | 6.80 | 3.40 |  |  |  |
| High Marsh | Ha19 |  | 20 to 30 | 1.485 | 4.56 | 6.77 | 3.38 |  |  |  |
| Total 0 to 30 | cm |  |  |  |  | 16.98 | 8.49 |  |  |  |
| High Marsh | Ha 21 |  | 0 to 10 | 0.098 | 72.22 | 7.09 | 3.55 |  |  |  |
| High Marsh | Ha21 |  | 10 to 20 | * | * | 10.95 | 5.48 |  |  |  |
| High Marsh | Ha21 |  | 20 to 30 | * | * | 9.25 | 4.63 |  |  |  |
| Total 0 to 30 | cm |  |  |  |  | 27.29 | 13.65 |  |  |  |


| Zone Site | Quadrat Depth cm | Bulk Density $\mathrm{g} / \mathrm{cm}^{3}$ | \% Organic Matter | TOM $\mathrm{kg} / \mathrm{m}^{2}$ | $\begin{array}{r} \mathrm{TOC} \\ \mathrm{~kg} / \mathrm{m}^{2} \\ \hline \hline \end{array}$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| High Marsh Ha36 | 0 to 10 | 0.153 | 65.34 | 9.98 | 4.99 |
| High Marsh Ha36 | 10 to 20 | 0.556 | 17.29 | 9.61 | 4.80 |
| High Marsh Ha36 | 20 to 30 | * | * | 9.25 | 4.63 |
| Total 0 to 30 cm |  |  |  | 28.84 | 14.42 |
| High Marsh Ha41 | 0 to 10 | 0.197 | 55.32 | 10.91 | 5.45 |
| High Marsh Ha41 | 10 to 20 | 1.579 | 5.63 | 8.89 | 4.45 |
| High Marsh Ha41 | 20 to 30 | 1.642 | 4.29 | 7.05 | 3.52 |
| Total 0 to 30 cm |  |  |  | 26.85 | 13.42 |
| High Marsh Ha52 | 0 to 10 | 0.139 | 70.52 | 9.77 | 4.88 |
| High Marsh Ha52 | 10 to 20 | 0.479 | 34.93 | 16.73 | 8.36 |
| High Marsh Ha52 | 20 to 30 | 0.960 | 15.51 | 14.88 | 7.44 |
| Total 0 to 30 cm |  |  |  | 41.38 | 20.69 |
| High Marsh Hb1 | 0 to 10 | 0.131 | 66.40 | 8.71 | 4.36 |
| High Marsh Hb1 | 10 to 20 | 0.517 | 37.58 | 19.41 | 9.71 |
| High Marsh Hb1 | 20 to 30 | * | * | 9.25 | 4.63 |
| Total 0 to 30 cm |  |  |  | 37.38 | 18.69 |
| High Marsh Hb4 | 0 to 10 | 0.140 | 66.31 | 9.25 | 4.63 |
| High Marsh Hb4 | 10 to 20 | 0.357 | 36.10 | 12.88 | 6.44 |
| High Marsh Hb4 | 20 to 30 | 1.405 | 10.08 | 14.16 | 7.08 |
| Total 0 to 30 cm |  |  |  | 36.30 | 18.15 |


| Zone Site | Quadrat Depth cm | Bulk Density g/cm | \% Organic Matter | TOM $\mathrm{kg} / \mathrm{m}^{2}$ | $\begin{array}{r} \mathrm{TOC} \\ \mathrm{~kg} / \mathrm{m}^{2} \\ \hline \end{array}$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| High Marsh Hc1 | 0 to 10 | 0.177 | 58.35 | 10.30 | 5.15 |
| High Marsh Hc1 | 10 to 20 | 0.165 | 58.55 | 9.63 | 4.82 |
| High Marsh Hc1 | 20 to 30 | 0.494 | 18.05 | 8.92 | 4.46 |
| Total 0 to 30 cm |  |  |  | 28.86 | 14.43 |
| High Marsh Hc30 | 0 to 10 | 0.133 | 50.54 | 6.70 | 3.35 |
| High Marsh Hc30 | 10 to 20 | 0.163 | 47.75 | 7.79 | 3.90 |
| High Marsh Hc30 | 20 to 30 | 0.168969 | 48.98 | 8.28 | 4.14 |
| Total 0 to 30 cm |  |  |  | 22.77 | 11.39 |
| High Marsh Hc37 | 0 to 10 | 0.115 | 58.03 | 6.70 | 3.35 |
| High Marsh Hc37 | 10 to 20 | 0.152 | 52.90 | 8.05 | 4.02 |
| High Marsh Hc37 | 20 to 30 | 0.208 | 45.82 | 9.54 | 4.77 |
| Total 0 to 30 cm |  |  |  | 24.29 | 12.14 |
| High Marsh Hc41 | 0 to 10 | 0.106 | 71.08 | 7.55 | 3.78 |
| High Marsh Hc41 | 10 to 20 | 0.154 | 58.91 | 9.06 | 4.53 |
| High Marsh Hc41 | 20 to 30 | * | * | 7.86 | 3.93 |
| Total 0 to 30 cm |  |  |  | 24.48 | 12.24 |
| High Marsh Hc42 | 0 to 10 | 0.101 | 80.37 | 8.13 | 4.06 |
| High Marsh Hc42 | 10 to 20 | 0.135 | 56.30 | 7.61 | 3.80 |
| High Marsh Hc42 | 20 to 30 | 0.183 | 59.62 | 10.88 | 5.44 |
| Total 0 to 30 cm |  |  |  | 26.62 | 13.31 |
| High Marsh Hc68 | 0 to 10 | 0.135 | 59.25 | 8.01 | 4.01 |
| High Marsh Hc68 | 10 to 20 | 0.149 | 60.97 | 9.08 | 4.54 |
| High Marsh Hc68 | 20 to 30 | 0.141 | 58.29 | 8.21 | 4.10 |
| Total 0 to 30 cm |  |  |  | 25.30 | 12.65 |


| Zone Site | Quadrat Depth cm | Bulk Density $\mathrm{g} / \mathrm{cm}^{3}$ | \% Organic Matter | $\begin{array}{r} \mathrm{TOM} \\ \mathrm{~kg} / \mathrm{m}^{2} \\ \hline \end{array}$ | $\begin{array}{r} \text { TOC } \\ \mathrm{kg} / \mathrm{m}^{2} \\ \hline \end{array}$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| High Marsh Hc72 | 0 to 10 | 0.112 | 64.11 | 7.18 | 3.59 |
| High Marsh Hc72 | 10 to 20 | 0.162 | 45.84 | 7.44 | 3.72 |
| High Marsh Hc72 | 20 to 30 | 0.239 | 33.18 | 7.92 | 3.96 |
| Total 0 to 30 cm |  |  |  | 22.53 | 11.27 |
| High Marsh Hc75 | 0 to 10 | 0.199 | 55.97 | 11.12 | 5.56 |
| High Marsh Hc75 | 10 to 20 | 0.414 | 25.34 | 10.48 | 5.24 |
| High Marsh Hc75 | 20 to 30 | 0.930 | 7.64 | 7.11 | 3.55 |
| Total 0 to 30 cm |  |  |  | 28.70 | 14.35 |
| High Marsh Hc84 | 0 to 10 | 0.141 | 66.51 | 9.36 | 4.68 |
| High Marsh Hc84 | 10 to 20 | 0.690 | 16.91 | 11.67 | 5.84 |
| High Marsh Hc84 | 20 to 30 | 1.565 | 2.75 | 4.30 | 2.15 |
| Total 0 to 30 cm |  |  |  | 25.33 | 12.67 |
| High Marsh Hc92 | 0 to 10 | 0.229 | 50.13 | 11.48 | 5.74 |
| High Marsh Hc92 | 10 to 20 | 0.190 | 52.27 | 9.94 | 4.97 |
| High Marsh Hc92 | 20 to 30 | 0.159 | 47.92 | 7.64 | 3.82 |
| Total 0 to 30 cm |  |  |  | 29.06 | 14.53 |
| High Marsh Ma1 | 0 to 10 | 0.094 | 70.88 | 6.66 | 3.33 |
| High Marsh Ma1 | 10 to 20 | 0.176 | 59.58 | 10.49 | 5.24 |
| High Marsh Ma1 | 20 to 30 | * | * | 9.25 | 4.63 |
| Total 0 to 30 cm |  |  |  | 26.40 | 13.20 |
| High Marsh Ma15 | 0 to 10 | 0.131 | 51.26 | 6.73 | 3.37 |
| High Marsh Ma15 | 10 to 20 | 0.232 | 30.57 | 7.09 | 3.55 |
| High Marsh Ma15 | 20 to 30 | 0.697 | 8.98 | 6.26 | 3.13 |
| Total 0 to 30 cm |  |  |  | 20.08 | 10.04 |


| Zone Site | Quadrat Depth cm | Bulk Density $\mathrm{g} / \mathrm{cm}^{3}$ | \% Organic Matter | $\begin{array}{r} \mathrm{TOM} \\ \mathrm{~kg} / \mathrm{m}^{2} \\ \hline \end{array}$ | $\begin{array}{r} \mathrm{TOC} \\ \mathrm{~kg} / \mathrm{m}^{2} \\ \hline \end{array}$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| High Marsh Mc2 | 0 to 10 | 0.222 | 48.885 | 10.84 | 5.42 |
| High Marsh Mc2 | 10 to 20 | 0.223 | 36.116 | 8.07 | 4.03 |
| High Marsh Mc2 | 20 to 30 | 0.192 | 38.850 | 7.46 | 3.73 |
| Total 0 to 30 cm |  |  |  | 26.37 | 13.19 |
| High Marsh Tb4 | 0 to 10 | 0.598 | 13.24 | 7.92 | 3.96 |
| High Marsh Tb4 | 10 to 20 | 1.525 | 3.16 | 4.82 | 2.41 |
| High Marsh Tb4 | 20 to 30 | 1.663 | 3.55 | 5.90 | 2.95 |
| Total 0 to 30 cm |  |  |  | 18.65 | 9.32 |
| High Marsh Tb14 | 0 to 10 | 0.869 | 8.99 | 7.82 | 3.91 |
| High Marsh Tb14 | 10 to 20 | 1.172 | 2.93 | 3.43 | 1.71 |
| High Marsh Tb14 | 20 to 30 | 1.473 | 2.63 | 3.87 | 1.94 |
| Total 0 to 30 cm |  |  |  | 15.12 | 7.56 |
| High Marsh Tb23 | 0 to 10 | 0.135 | 59.64 | 8.03 | 4.02 |
| High Marsh Tb23 | 10 to 20 | 0.722 | 13.86 | 10.00 | 5.00 |
| High Marsh Tb23 | 20 to 30 | 1.704 | 3.60 | 6.14 | 3.07 |
| Total 0 to 30 cm |  |  |  | 24.18 | 12.09 |
| LMSS LSa8 | 0 to 10 | 0.380 | 28.74 | 10.91 | 5.46 |
| LMSS LSa8 | 10 to 20 | 0.327 | 26.09 | 8.53 | 4.26 |
| LMSS LSa8 | 20 to 30 | * | * | 7.11 | 3.56 |
| Total 0 to 30 cm |  |  |  | 26.55 | 13.28 |
| LMSS LSa12 | 0 to 10 | 1.102 | 7.91 | 8.72 | 4.36 |
| LMSS LSa12 | 10 to 20 | 1.447 | 5.28 | 7.64 | 3.82 |
| LMSS LSa12 | 20 to 30 | * | * | 7.11 | 3.56 |
| Total 0 to 30 cm |  |  |  | 23.46 | 11.73 |


| Zone | Site | Quadrat | Depth cm | Bulk Density $\mathrm{g} / \mathrm{cm}^{3}$ | \% Organic Matter | $\begin{gathered} \mathrm{TOM} \\ \mathrm{~kg} / \mathrm{m}^{2} \end{gathered}$ | $\begin{array}{r} \mathrm{TOC} \\ \mathrm{~kg} / \mathrm{m}^{2} \\ \hline \hline \end{array}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| LMSS | LSb2 |  | 0 to 10 | 0.281 | 28.66 | 8.05 | 4.03 |
| LMSS | LSb2 |  | 10 to 20 | 0.668 | 12.64 | 8.44 | 4.22 |
| LMSS | LSb2 |  | 20 to 30 | 0.967 | 8.04 | 7.78 | 3.89 |
| Total 0 to 30 cm |  |  |  |  |  | 24.26 | 12.13 |
| LMSS | LSb16 |  | 0 to 10 | 1.347 | 3.76 | 5.06 | 2.53 |
| LMSS | LSb16 |  | 10 to 20 | 1.721 | 2.43 | 4.18 | 2.09 |
| LMSS | LSb16 |  | 20 to 30 | 1.749 | 2.78 | 4.87 | 2.44 |
| Total 0 to 30 cm |  |  |  |  |  | 14.11 | 7.05 |
| LMSS | LSb17 |  | 0 to 10 | 0.408 | 28.87 | 11.77 | 5.89 |
| LMSS | LSb17 |  | 10 to 20 | 1.364 | 4.53 | 6.17 | 3.09 |
| LMSS | LSb17 |  | 20 to 30 | 1.571 | 2.50 | 3.92 | 1.96 |
| Total 0 to 30 cm |  |  |  |  |  | 21.87 | 10.93 |
| LMSS | LSb46 |  | 0 to 10 | 0.292 | 41.59 | 12.13 | 6.07 |
| LMSS | LSb46 |  | 10 to 20 | 0.232 | 35.14 | 8.14 | 4.07 |
| LMSS | LSb46 |  | 20 to 30 | 0.376 | 21.59 | 8.11 | 4.06 |
| Total 0 to 30 cm |  |  |  |  |  | 28.38 | 14.19 |
| LMSS | LSb53 |  | 0 to 10 | 0.312 | 31.77 | 9.90 | 4.95 |
| LMSS | LSb53 |  | 10 to 20 | 0.358 | 22.04 | 7.88 | 3.94 |
| LMSS | LSb53 |  | 20 to 30 | 0.838 | 6.62 | 5.55 | 2.77 |
| Total 0 to 30 cm |  |  |  |  |  | 23.34 | 11.67 |
| LMSS | LSb54 |  | 0 to 10 | 0.277 | 36.95 | 10.23 | 5.11 |
| LMSS | LSb54 |  | 10 to 20 | 0.308 | 28.52 | 8.77 | 4.39 |
| LMSS | LSb54 |  | 20 to 30 | 0.432 | 28.88 | 12.46 | 6.23 |
| Total 0 to 30 cm |  |  |  |  |  | 31.46 | 15.73 |


| Zone | Site | Quadrat | Depth cm | Bulk Density $\mathrm{g} / \mathrm{cm}$ | \% Organic Matter | $\begin{array}{r} \mathrm{TOM} \\ \mathrm{~kg} / \mathrm{m}^{2} \\ \hline \hline \end{array}$ | $\begin{array}{r} \mathrm{TOC} \\ \mathrm{~kg} / \mathrm{m}^{2} \\ \hline \end{array}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| LMSS | Mb4 |  | 0 to 10 | 0.288 | 38.58 | 11.11 | 5.55 |
| LMSS | Mb4 |  | 10 to 20 | 0.254 | 37.64 | 9.55 | 4.77 |
| LMSS | Mb4 |  | 20 to 30 | * | * | 7.11 | 3.56 |
| Total 0 to 30 cm |  |  |  |  |  | 27.76 | 13.88 |
| LMTS | LT6 |  | 0 to 10 | 0.368 | 22.85 | 8.40 | 4.20 |
| LMTS | LT6 |  | 10 to 20 | 0.415 | 19.76 | 8.20 | 4.10 |
| LMTS | LT6 |  | 20 to 30 | 0.364 | 21.87 | 7.97 | 3.98 |
| Total 0 to 30 cm |  |  |  |  |  | 24.56 | 12.28 |
| LMTS | LT12 |  | 0 to 10 | 0.330 | 19.97 | 6.58 | 3.29 |
| LMTS | LT12 |  | 10 to 20 | * | * | 7.41 | 3.71 |
| LMTS | LT12 |  | 20 to 30 | * | * | 6.56 | 3.28 |
| Total 0 to 30 cm |  |  |  |  |  | 20.55 | 10.27 |
| LMTS | LT25 |  | 0 to 10 | 0.365 | 20.24 | 7.39 | 3.69 |
| LMTS | LT25 |  | 10 to 20 | 0.362 | 15.32 | 5.55 | 2.77 |
| LMTS | LT25 |  | 20 to 30 | 0.376 | 18.43 | 6.93 | 3.46 |
| Total 0 to 30 cm |  |  |  |  |  | 19.86 | 9.93 |
| LMTS | LT29 |  | 0 to 10 | 0.623 | 14.15 | 8.82 | 4.41 |
| LMTS | LT29 |  | 10 to 20 | 0.963 | 9.12 | 8.79 | 4.39 |
| LMTS | LT29 |  | 20 to 30 | * | * | 6.56 | 3.28 |
| Total 0 to 30 cm |  |  |  |  |  | 24.17 | 12.08 |
| LMTS | LT33 |  | 0 to 10 | 0.301 | 15.41 | 4.64 | 2.32 |
| LMTS | LT33 |  | 10 to 20 | 0.851 | 6.92 | 5.89 | 2.94 |
| LMTS | LT33 |  | 20 to 30 | 0.888 | 3.81 | 3.39 | 1.69 |
| Total 0 to 30 cm |  |  |  |  |  | 13.92 | 6.96 |


| Zone | Site | Quadrat Depth cm | Bulk Density $\mathrm{g} / \mathrm{cm}$ | \% Organic Matter | $\begin{gathered} \text { TOM } \\ \mathrm{kg} / \mathrm{m}^{2} \\ \hline \end{gathered}$ | $\begin{array}{r} \mathrm{TOC} \\ \mathrm{~kg} / \mathrm{m}^{2} \\ \hline \end{array}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| LMTS | LTS6 | 0 to 10 | 0.331 | 28.35 | 9.37 | 4.69 |
| LMTS | LTS6 | 10 to 20 | 0.356 | 22.60 | 8.05 | 4.02 |
| LMTS | LTS6 | 20 to 30 | * | * | 6.56 | 3.28 |
| Total 0 to 30 cm |  |  |  |  | 23.98 | 11.99 |
| LMTS | LTS9 | 0 to 10 | 0.389 | 20.72 | 8.05 | 4.03 |
| LMTS | LTS9 | 10 to 20 | 0.438 | 18.21 | 7.98 | 3.99 |
| LMTS | LTS9 | 20 to 30 | 0.532 | 14.92 | 7.94 | 3.97 |
| Total 0 to 30 cm |  |  |  |  | 23.98 | 11.99 |
| Tidal Creek | C1 | 0 to 10 | 0.555 | 12.57 | 6.98 | 3.49 |
| Tidal Creek | C1 | 10 to 20 | 0.439 | 12.35 | 5.42 | 2.71 |
| Tidal Creek | C1 | 20 to 30 | 0.685 | 7.67 | 5.25 | 2.63 |
| Total 0 to 30 cm |  |  |  |  | 17.65 | 8.83 |
| Tidal Creek | C2 | 0 to 10 | 0.598 | 8.87 | 5.30 | 2.65 |
| Tidal Creek | C2 | 10 to 20 | 0.605 | 11.17 | 6.75 | 3.38 |
| Tidal Creek | C2 | 20 to 30 | 0.684 | 9.24 | 6.32 | 3.16 |
| Total 0 to 30 cm |  |  |  |  | 18.38 | 9.19 |
| Tidal Creek | C3 | 0 to 10 | 0.608 | 9.92 | 6.04 | 3.02 |
| Tidal Creek | C3 | 10 to 20 | 0.839 | 5.42 | 4.55 | 2.27 |
| Tidal Creek | C3 | 20 to 30 | 0.843 | 6.89 | 5.80 | 2.90 |
| Total 0 to 30 cm |  |  |  |  | 16.39 | 8.20 |

APPENDIX H. CHN ANALYSES OF SOIL CORES.

|  |  |  |  |
| :--- | :---: | ---: | :---: |
| Sample ID | Depth (cm) | \%C | $\%$ N |
| Fa18Q1 | $0-10$ | 6.11 | 0.35 |
| Fa18Q2 | $10-20$ | 1.64 | 0.08 |
| Fa18Q4 | $20-30$ | 0.69 | 0.04 |
| Fa30Q1 | $10-19$ | 2.16 | 0.15 |
| Fb47Q4 | $0-10$ | 13.75 | 0.38 |
| Fb47Q4 | $10-20$ | 2.56 | 0.07 |
| Fb65Q4 | $0-10$ | 13.88 | 0.42 |
| Ha5 | $10-20$ | 2.94 | 0.12 |
| Ha19 | $0-10$ | 34.40 | 1.63 |
| Ha36 | $10-14$ | 7.36 | 0.35 |
| Ha41 | $10-20$ | 2.62 | 0.08 |
| Ha52 | $10-20$ | 14.42 | 0.94 |
| Hc1 | $20-30$ | 7.93 | 0.50 |
| Hc37 | $0-10$ | 18.82 | 1.17 |
| Hc72 | $10-20$ | 12.42 | 0.83 |
| Hc84 | $10-20$ | 6.72 | 0.34 |
| LSa8 | $10-20$ | 10.04 | 0.56 |
| LSa12 | $0-10$ | 3.08 | 0.15 |
| LSb2 | $0-10$ | 12.75 | 0.62 |
| LSb16 | $0-10$ | 1.09 | 0.06 |
| LSb16 | $20-30$ | 0.35 | 0.02 |
| LSb53 | $20-30$ | 2.43 | 0.13 |
| LTS9 | $20-30$ | 5.71 | 0.32 |
| LT33 | $10-20$ | 2.60 | 0.14 |
| Tb4 | $0-10$ | 5.98 | 0.28 |
| Tb4 | $10-20$ | 1.02 | 0.03 |
| Tb4 | $20-30$ | 0.72 | 0.03 |
| Tb23Q2 | $10-20$ | 8.60 | 0.47 |
| Tb23Q3 | $10-20$ | 6.96 | 0.28 |
| Tb23Q4 | $0-10$ | 7.61 | 0.42 |
| Tb29Q1 | $10-20$ | 2.14 | 0.07 |
| Tb29Q1 | $20-30$ | 0.75 | 0.03 |
| Tb29Q3* | $10-20$ | 1.52 | 0.05 |
| Tc2Q2 | $20-30$ | 1.14 | 0.03 |
| Tc2Q3 | $20-30$ | 0.85 | 0.03 |
| Tc2Q5 | $0-10$ | 30.08 | 1.08 |
| Tc2Q5 | $20-30$ | 1.71 | 0.09 |
| Tc8Q3 | $20-30$ | 1.63 | 0.07 |
| Tc8Q4 | $10-20$ | 8.15 | 0.34 |
| Tc8Q4 | $20-30$ | 2.36 | 0.09 |
|  |  |  |  |


| Sample ID | Depth (cm) | \%C | $\%$ N |
| :--- | :---: | ---: | ---: |
| Tc31Q2 | $10-20$ | 2.63 | 0.10 |
| Ta1Q2 | $0-10$ | 33.96 | 1.62 |
| Ta4Q4 | $10-20$ | 3.42 | 0.12 |
| Ta13Q1 | $10-20$ | 3.71 | 0.17 |
| Ta13Q5 | $10-20$ | 3.11 | 0.13 |
| Td2Q5 | $10-20$ | 2.70 | 0.12 |
| Td15Q2 | $20-30$ | 1.49 | 0.03 |
| Td15Q4 | $20-30$ | 0.95 | 0.03 |
| Td15Q5 | $0-10$ | 34.34 | 1.67 |
| Td39Q4 | $10-20$ | 1.46 | 0.04 |
| *Soil sample was not used in |  |  |  |

*Soil sample was not used in cubic regression due to possible error in bulk density $\left(\mathrm{g} / \mathrm{cm}^{3}\right)$.

APPENDIX I. GPS SURVEY OF SITE
COORDINATES AND ELEVATION FOR ALL
SITES EXCEPT TIDAL CREEK.

|  |  |  |  |
| :--- | :--- | :--- | :---: |
| Site | Latitude | Longitude | Elevation $(\mathrm{m})$ |
| Fa18 | 37.45827275 | -75.83101698 | 2.540 |
| Fa30 | 37.45800192 | -75.8309846 | 1.844 |
| Fb29 | 37.46429748 | -75.83741756 | 2.960 |
| Fb47 | 37.4647559 | -75.83097758 | 2.210 |
| Fb65 | 37.4629122 | -75.83337509 | 1.201 |
| Ta1 | 37.462267 | -75.82931942 | 1.367 |
| Ta4 | 37.46158267 | -75.8358191 | 0.730 |
| Ta13 | 37.4596393 | -75.83129823 | 1.087 |
| Tb29 | 37.46084767 | -75.8340213 | 0.794 |
| Tc2 | 37.4624025 | -75.82980253 | 1.132 |
| Tc8 | 37.46216677 | -75.83096015 | 1.316 |
| Tc18 | 37.4616915 | -75.83180489 | 1.122 |
| Tc31 | 37.46174132 | -75.8324005 | 1.157 |
| Td2 | 37.46269529 | -75.83326499 | 1.195 |
| Td9 | 37.46279632 | -75.83386952 | 1.187 |
| Td15 | 37.46287196 | -75.83508715 | 1.225 |
| Td33 | 37.46318967 | -75.83431557 | 1.099 |
| Td39 | 37.46317467 | -75.83601544 | 1.245 |
| Td50 | 37.4634608 | -75.83402833 | 0.995 |
| Ha5 | 37.4595168 | -75.83155473 | 1.229 |
| Ha14 | 37.4601146 | -75.83092076 | 1.063 |
| Ha19 | 37.46020715 | -75.83205501 | 1.047 |
| Ha21 | 37.46049289 | -75.83221814 | 0.957 |
| Ha36 | 37.46056984 | -75.83084546 | $*$ |
| Ha41 | 37.46073699 | -75.8311535 | 1.051 |
| Ha52 | 37.46181218 | -75.83039906 | 1.030 |
| Hb1 | 37.46146852 | -75.83229173 | 1.104 |
| Hb4 | 37.4612343 | -75.83287377 | 1.067 |
| Hc1 | 37.46316958 | -75.83769199 | 0.954 |
| Hc30 | 37.46228915 | -75.83692389 | 0.940 |
| Hc37 | 37.46185523 | -75.8350857 | 1.086 |
| Hc41 | 37.46199484 | -75.83426657 | 0.922 |
| Hc42 | 37.4620176 | -75.83396079 | 1.085 |
| Hc68 | 37.46110735 | -75.83547743 | 1.059 |
| Hc72 | 37.46174742 | -75.83674655 | 0.926 |
| Hc75 | 37.46143113 | -75.83676859 | 0.886 |
| Hc84 | 37.46089095 | -75.83450192 | 1.090 |
| Hc92 | 37.4606908 | -75.83561184 | 0.914 |
| Ma1 | 37.46005265 | -75.83216438 | 1.012 |
| Ma15 | 37.45932557 | -75.83262928 | 0.978 |
| Mc2 | 37.46030785 | -75.83657352 | 0.698 |
| Tb4 | 37.46089833 | -75.83243869 | 1.102 |
|  |  |  |  |
|  |  |  |  |


| Site | Latitude | Longitude | Elevation $(\mathrm{m})$ |
| :--- | :--- | :--- | :---: |
| Tb14 | 37.45977767 | -75.83315822 | 1.005 |
| Tb23 | 37.46125428 | -75.8334436 | 1.023 |
| LSa8 | 37.45863325 | -75.83277866 | 0.784 |
| LSa12 | 37.45880534 | -75.83241855 | 0.852 |
| LSb2 | 37.45902331 | -75.8332824 | 0.767 |
| LSb16 | 37.45944985 | -75.83354545 | 0.850 |
| LSb17 | 37.45921942 | -75.83364574 | 0.801 |
| LSb46 | 37.45910627 | -75.83523896 | 0.722 |
| LSb53 | 37.46024436 | -75.83571531 | 0.674 |
| LSb54 | 37.46047421 | -75.8358191 | 0.730 |
| Mb4 | 37.45981122 | -75.83595583 | 0.738 |
| LT6 | 37.45932196 | -75.83627024 | 0.447 |
| LT12 | 37.45905437 | -75.83555928 | 0.416 |
| LT25 | 37.45848405 | -75.83441722 | 0.488 |
| LT29 | 37.45827532 | -75.83328139 | 0.268 |
| LT33 | 37.45884185 | -75.83298936 | 0.348 |
| LTS6 | 37.46063246 | -75.83695996 | 0.484 |
| LTS9 | 37.46016193 | -75.83702832 | 0.516 |

[^1]
[^0]:    ${ }^{\text {a }}$ Canary Creek marsh, ${ }^{\text {b }}$ Blackbird marsh, ${ }^{\text {c }}$ Snow's marsh, ${ }^{\text {a lower marsh, }}{ }^{e}$ upper marsh

[^1]:    * Error in elevation reading

